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Shifts in dimethylated sulfur concentrations and microbiome composition in the red-tide causing dinoflagellate *Alexandrium minutum* during a simulated marine heatwave

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Abstract. The biogenic sulfur compounds dimethyl sulfide (DMS), dimethyl sulfoniopropionate (DMSP) and dimethyl sulfoxide (DMSO) are produced and transformed by diverse populations of marine microorganisms and have substantial physiological, ecological and biogeochemical importance spanning organism to global scales. Understanding the production and transformation dynamics of these compounds under shifting environmental conditions is important for predicting their roles in a changing ocean. Here, we report the physiological and biochemical response of a robust strain of *Alexandrium minutum*, a dinoflagellate with the highest reported intracellular DMSP content, exposed to a 6 d increase in temperature mimicking mild and extreme coastal marine heatwave conditions ($+4$ and $+12\,°C$). Under mild temperature increases ($+4\,°C$), *A. minutum* growth was enhanced, with no measurable physiological stress response. However, under a very acute increase in temperature ($+12\,°C$) triggering thermal stress, *A. minutum* growth declined, photosynthetic efficiency (F_V/F_M) was impaired, and enhanced oxidative stress was observed. These physiological responses indicative of thermal stress were accompanied by increased DMS and DMSO concentrations followed by decreased DMSP concentration. At this temperature extreme, we observed a cascading stress response in *A. minutum*, which was initiated 6 h after the start of the experiment by a spike in DMS and DMSO concentrations and a rapid decrease in F_V/F_M. This was followed by an increase in reactive oxygen species (ROS) and an abrupt decline in DMS and DMSO on day 2 of the experiment. A subsequent decrease in DMSP coupled with a decline in the growth rate of both *A. minutum* and its associated total bacterial assemblage coincided with a shift in the composition of the *A. minutum* microbiome. Specifically, an increase in the relative abundance of the operational taxonomic units (OTUs) matching *Oceanicaulis* (17.0 %), Phycisphaeraceae *SM1A02* (8.8 %) and *Balneola* (4.9 %) as well as a decreased relative abundance of *Maribacter* (24.4 %), *Marinoscillum* (4.7 %) and *Seohaeicola* (2.7 %) were primarily responsible for differences in microbiome structure observed between temperature treatments. These shifts in microbiome structure are likely to have been driven by either the temperature itself, the changing physiological state of *A. minutum* cells, shifts in biogenic sulfur concentrations, the presence of other solutes, or a combination of all. Nevertheless, we suggest that these results point to the significant effect of extreme heatwaves on the physiology, growth and microbiome composition of the red-tide causing dinoflagellate *A. minutum*, as well as potential implications for biogenic sulfur cycling processes and marine DMS emissions.

1 Introduction

Many marine phytoplankton produce the organic sulfur compound dimethyl sulfoniopropionate (DMSP) (Zhou et al., 2009; Berdalet et al., 2011; Caruana and Malin, 2014), as it can function as an antioxidant, osmolyte, chemoattractant and currency in reciprocal chemical exchanges with heterotrophic bacteria (Stefels, 2000; Sunda et al., 2002; Kiene et al., 2000; Seymour et al., 2010). Phytoplankton-derived DMSP is in fact a major source of sulfur and carbon for marine heterotrophic bacteria (Kiene et al., 2000), which in turn

play a major role in the cycling and turnover of organosulfur compounds in the ocean (Todd et al., 2007; Curson et al., 2011). The subsequent cycling of DMSP into other biogenic sulfur molecules including dimethyl sulfide (DMS) and dimethyl sulfoxide (DMSO) by a suite of microbial transformation pathways (Kiene et al., 2000; Sunda et al., 2002) and physical drivers (Brimblecombe and Shooter, 1986) have important ecological and biogeochemical implications spanning from cellular to global scales (Sunda et al., 2002; Charlson et al., 1987; DeBose et al., 2008; Van Alstyne et al., 2001; Knight, 2012; Nevitt et al., 1995).

Among DMSP-producing phytoplankton, the dinoflagellate *Alexandrium minutum* has the highest recorded DMSP cell content, with an average concentration of $14.2 \, \text{pmol cell}^{-1}$, compared with less than $1 \, \text{pmol cell}^{-1}$ in most other dinoflagellates (Caruana and Malin, 2014). Blooms of *A. minutum* occur from the Mediterranean Sea to the South Pacific coast in sea surface waters within temperature ranges of 12 to $25\,^\circ\text{C}$ (Laabir et al., 2011). Notably, some strains of *Alexandrium*, including *A. minutum*, produce saxitoxins, which lead to paralytic shellfish poisoning (PSP) and are responsible for the most harmful algal blooms in terms of magnitude, distribution and consequences on human health (Anderson et al., 2012).

A. minutum commonly inhabits shallow coastal and estuarine waters (Anderson, 1998), which are globally experiencing substantial shifts in environmental conditions, including increases in sea surface temperature (SST) associated with climate change (Harley et al., 2006). Although generally less studied than chronic temperature rises associated with global climate change (Frölicher and Laufkötter, 2018), acute ephemeral temperature increases known as marine heatwaves (MHWs) (Hobday et al., 2016) have recently been demonstrated to be becoming more frequent and persistent as a consequence of climate change (Oliver et al., 2018). Increases in MHW occurrence are anticipated to become particularly frequent within the shallow coastal and estuarine waters, where *A. minutum* blooms occur (Ummenhofer and Meehl, 2017; Anderson, 1998).

Coastal MHW events have recently had dramatic impacts on coastal environments. MHWs of up to $6\,^\circ\text{C}$ increase in temperature in Western Australian coastal waters (2011) and the northeast Pacific (2013–2015) resulted in significant ecosystem shifts with increases in novel species at the expenses of others (Frölicher and Laufkötter, 2018). The 2016 MHW that was associated with El Niño–Southern Oscillations resulted in an $8\,^\circ\text{C}$ increase in sea surface temperature leading to the mass coral bleaching of more than $90\,\%$ of the Great Barrier Reef (Hughes et al., 2017). While it is clear that MHWs can have severe consequences on a variety of systems and organisms, their effects on marine microbes and the biogeochemical processes that they mediate have rarely been investigated (Joint and Smale, 2017).

While there is evidence that increases in seawater temperature can lead to increased DMSP and/or DMS concentrations in phytoplankton (McLenon and DiTullio, 2012; Sunda et al., 2002), it is not clear how a shift in DMSP net production by phytoplankton under acute temperature stress will alter the composition and function of their associated microbiome and how, in turn, this will influence biogenic sulfur cycling processes within marine habitats. There is therefore a pressing need to understand the physiological and biogeochemical consequences of thermal stress on phytoplankton–bacteria interactions within the context of events such as MHWs. This is particularly important, given that a shift in the composition of the phytoplankton microbiome could potentially dictate atmospheric DMS fluxes depending on whether the bacterial community preferentially cleaves or demethylates DMSP (Todd et al., 2007; Kiene et al., 2000).

The aims of this study were to investigate how an acute increase in temperature $(+12\,^\circ\text{C})$, comparable to those associated with MHW events and leading to thermal stress in *A. minutum* could alter the physiological state and biogenic sulfur cycling dynamics of *A. minutum* and determine how these changes might influence the composition of the *Alexandrium* microbiome. We hypothesised that an abrupt increase in temperature would lead to physiological impairment (Falk et al., 1996; Robison and Warner, 2006; Iglesias-Prieto et al., 1992; Rajadurai et al., 2005) and oxidative stress (Lesser, 2006) in *A. minutum*, leading to an upregulation of DMSP, DMS and DMSO production (McLenon and DiTullio, 2012; Sunda et al., 2002) in this high DMSP producer, which could ultimately lead to a shift in the composition of the *A. minutum* microbiome.

2 Methods

2.1 Culturing and experimental design

Cultures of *Alexandrium minutum* (CS-324), isolated from southern Australian coastal waters (Port River, Adelaide, 11 November 1988, CSIRO, ANACC's collection) were grown in GSe medium at $18\,^\circ\text{C}$ and $50 \, \mu\text{mol photons m}^{-2}\,\text{s}^{-1}$ under a 12 : 12 light : dark cycle. One month before the start of each experiment, *A. minutum* cultures were acclimated over four generations to $20\,^\circ\text{C}$ (average summer temperature at Port River, IMOS) and $200 \, \mu\text{mol photons m}^{-2}\,\text{s}^{-1}$ using a 14 : 10 h light : dark cycle mimicking summer conditions. Light intensity was comparable to that used in Berdalet et al. (2011) for *A. minutum* and conveniently allowed the growth of other algae cultures using the same facilities. Cultures were grown to a cell concentration of $\sim 60\,000 \, \text{mL}^{-1}$ before cells were inoculated into fresh GSe medium. 6 d prior to the start of each experiment, 20 L of GSe medium was inoculated with a cell concentration of $1140 \, \text{mL}^{-1}$ (experiment 1, April 2016) and $680 \, \text{mL}^{-1}$ (experiment 2, June 2016) and aliquots of 500 mL were transferred into 40 individual 750 mL sterile tissue culture flasks. Culture flasks were incubated in four independent water

baths (10 flasks in each) and maintained under control conditions of 20 °C and 200 µmol photons m^{-2} s^{-1}. Temperature and light control was achieved using circulating water heaters (Julabo USA, Inc.) and programmable LED lights (Hydra FiftyTwo, AquaIllumination, USA). All cultures were mixed twice daily to keep cells in suspension by gentle swirling.

On day 1 (T_0), five culture flasks from each 20 °C water bath were transferred to four new water baths for exposure to experimental treatment temperatures (either 24 °C in experiment 1 or 32 °C in experiment 2) so that each control and experimental water bath contained five flasks. Experimental temperatures were carefully chosen based on preliminary experiments conducted at 24, 28, 30 and 32 °C, where only a 12 °C increase in temperature (32 °C treatment) led to a physiological stress response in this robust strain of *A. minutum* in culture. Although an increase in temperature of this magnitude might be rare in coastal marine systems, this presented a unique opportunity to investigate the consequences of MHW-induced thermal stress on this relevant phytoplankton. One culture flask from each tank was immediately sampled for baseline measurements of the following: DMS (2 mL), DMSP and DMSO (1 mL) concentrations, photochemical efficiency (3 mL), algal and bacterial cell counts (1 mL), ROS quantification (1 mL) and DNA extraction ($\sim$ 470 mL). The dissolved DMSP fraction was not determined because preliminary investigations showed that gravity filtration was too time consuming, potentially due to clogging of filters by the large *A. minutum* cells (30 µm diameter), leading to filtration artefacts for DMSP analysis, as previously mentioned by Berdalet et al. (2011). At 18:00 AEDT on day 1 (T_6), 12:00 on day 2 (T_{24}), 12:00 on day 5 (T_{96}) and 12:00 on day 6 (T_{120}) one flask from each of the eight water baths was removed from the incubation conditions and sampled as described above.

2.2 Photosynthetic efficiency measurements

Subsamples for measurement of photosynthetic efficiency were dark adapted for 10 min under aluminium foil and transferred to a quartz cuvette for pulse amplitude modulated (PAM) fluorometric analysis using a WATER PAM (Walz GmbH, Effeltrich, Germany). Once the base fluorescence (F_0) signal had stabilised (measuring light intensity 3, frequency 2 s), a saturating pulse (intensity 12, width 0.8 s) was used to measure the maximum quantum yield (F_V/F_M) of photosystem II (PSII). As base fluorescence is dependent on cell density, the photomultiplier gain was adjusted and recorded to maintain F_0 at a level of 0.2 a.u. before saturating the photosystem. Samples were kept in suspension during measurements via continuous stirring at minimal speed inside the quartz cuvette to avoid cells settling.

2.3 Microalgal and bacterial cell counts

Subsamples for bacterial cell counts were stained with SYBR Green at a final concentration of 1 : 10 000 and incubated in the dark for 15 min (Marie et al., 1997). Subsamples for microalgal cell counts and stained subsamples for bacterial cell counts were diluted 1 : 10 and 1 : 100 respectively into sterile GSe medium prior to analysis with a BD Accuri C6 Flow Cytometer (Becton Dickinson). Phytoplankton cells were discriminated using red autofluorescence and side scatter (SSC), whereas bacterial populations were discriminated and quantified using SYBR Green fluorescence and SSC.

2.4 Reactive oxygen species measurements

The presence of reactive oxygen species (ROS) was detected within cultures using the fluorescent probe 2,7-dichlorodihydrofluorescein-diacetate (CM-H2DCFDA; Molecular Probes), which binds to ROS and other peroxides (Rastogi et al., 2010). The reagent was thawed at room temperature for 10 min and activated using 86.5 µL of DMSO, with 5 µL of activated reagent added to each sample (final concentration 5 µM). Samples were vortexed for 5 s and incubated at room temperature for 30 min. Samples were then centrifuged at 2000 g for 2 min, the supernatant with reagent dye was discarded, and stained cells were resuspended in 1 mL of phosphate-buffered saline (PBS), prior to quantification of fluorescence by flow cytometry. Mean green fluorescence was quantified from cytograms of forward light scatter (FSC) against green fluorescence. A positive control ($+10$ µL of 30 % H_2O_2, final concentration 97 mM) and negative (no ROS added) control of PBS were run to ensure that detected cell fluorescence was completely attributable to the ROS probe.

2.5 Sulfur analysis by gas chromatography

The preparation of all blanks and samples used in the dilution steps described below were prepared with sterile (0.2 µM filtered and autoclaved) phosphate-buffered saline (PBS, salinity 35 ppt) to avoid cell damage from altered osmolarity and to maintain similar physical properties as seawater during headspace analysis by gas chromatography. Aliquots for DMS analysis were transferred into 14 mL headspace vials that were immediately capped and crimped using butyl rubber septa (Sigma Aldrich Pty; 27232) and aluminium caps (Sigma Aldrich Pty; 27227-U) respectively. DMSP aliquots were 1 : 1 diluted with sterile PBS, and DMSP was cleaved to DMS by adding 1 pellet of NaOH to each vial, which was immediately capped and crimped. Samples were incubated for a minimum of 30 min at room temperature to allow for the alkaline reaction and equilibration to occur prior to analysis by gas chromatography (Kiene and Slezak, 2006).

DMS and DMSP samples were analysed by 500 µL direct headspace injections using a Shimadzu gas chromatograph

(GC-2010 Plus) coupled with a flame photometric detector (FPD) set at 180 °C with instrument grade air and hydrogen flow rates set at 60 and 40 mL min^{-1}, respectively. DMS was eluted on a capillary column (30 m × 0.32 mm × 5 µm) set at 120 °C using high purity helium (He) as the carrier gas at a constant flow rate of 5 mL min^{-1} and a split ratio of 5. A six-point calibration curve and PBS blanks were run by 500 µL direct headspace injections prior to subsampling culture flasks using small volumes of concentrated DMSP. HCl standard solutions (certified reference material WR002, purity 90.3 ± 1.8 % mass fraction, National Measurement Institute, Sydney, Australia) were diluted in sterile PBS to a final volume of 2 mL. Detection limit was 50 nM for 500 µL headspace injections. Concentrations obtained in vials treated with NaOH accounted for both DMS and DMSP. Consequently, the DMSP concentration in each sample was obtained by subtracting the corresponding DMS concentration.

Following DMS and DMSP analysis, alkaline samples used for DMSP analysis were uncapped and left to vent overnight under a fume hood. On the next day, samples were purged for 10 min with high purity N$_2$ at an approximate flow rate of 60 mL min^{-1} to remove any remaining DMS produced from the alkaline treatment. Samples were then neutralised by adding 80 µL of 32 % HCl, and DMSO was converted to DMS by adding 350 µL of 12 % TiCl$_3$ solution to each vial, which were then immediately capped and crimped (Kiene and Gerard, 1994; Deschaseaux et al., 2014b). Vials were then heated in a water bath at 50 °C for 1 h and cooled down to room temperature prior to analysis by 500 µL direct headspace injections on the GC-FPD as described above. A five-point calibration curve was run prior to DMSO analysis using DMSO standard solutions (Sigma Aldrich Pty; D2650) diluted in PBS to a final volume of 2 mL and converted to DMS with TiCl$_3$ in the same manner as the experimental samples. PBS blanks treated with NaOH and TiCl$_3$ were also run along with the calibration curves. All dimethylated sulfur compounds were normalised to cell density, which best reflects biogenic production.

2.6 DNA extraction

Following sub-sampling for the physiological and biogenic sulfur measurements described above, the remaining 400 mL within each culture flask was filtered onto a 47 mm diameter, 0.22 µm polycarbonate filter (Millipore) with a peristaltic pump at a rate of 80 rpm to retain cells for DNA analysis. The filters were subsequently stored in cryovials, snap frozen with liquid nitrogen and stored at −80 °C until extraction. DNA extraction was performed using a bead-beating and chemical lysis based DNA extraction kit (PowerWater DNA Isolation Kit, MoBio Laboratories) following the manufacturer's instructions. DNA quantity and purity were checked for each sample using a Nanodrop 2000 (Thermo Fisher Scientific, Wilmington, DE, USA). Three replicate samples

with the highest DNA quantity and purity from the control and treatment tanks, collected at the beginning (T_0) and end (T_{120}) of the experiment, were subsequently sequenced.

2.7 16S rRNA amplicon sequencing and bioinformatics

To characterise the bacterial assemblage structure (microbiome) of *A. minutum* cultures, we employed 16S rRNA amplicon sequencing. We amplified the V1–V3 variable regions of the 16S rRNA gene using the 27F (AGAGTTTGATCMTGGCTCAG, Lane, 1991) and 519R (GWATTACCGCGGCKGCTG, Turner et al., 1999) primer pairing, with amplicons subsequently sequenced using the Illumina MiSeq platform (Ramaciotti Centre for Genomics; Sydney, NSW, Australia) following the manufacturer's guidelines. Raw data files in FASTQ format were deposited in the National Center for Biotechnology Information (NCBI) Sequence Read Archive (SRA) under the study accession number PRJNA486692.

Bacterial 16S rRNA gene sequencing reads were analysed using the QIIME pipeline (Caporaso et al., 2010; Kuczynski et al., 2012). Briefly, paired-end DNA sequences were joined, de novo operational taxonomic units (OTUs) were defined at 97 % sequence identity using UCLUST (Edgar, 2010) and taxonomy was assigned against the SILVA v128 database (Quast et al., 2012; Yilmaz et al., 2013). Chimeric sequences were detected using usearch61 (Edgar, 2010) and together with chloroplast OTUs were filtered from the dataset. Sequences were then aligned, filtered and rarefied to the same depth to remove the effect of sampling effort upon analysis.

2.8 Statistical analysis

Repeated measures analysis of variance (rmANOVA) models were fitted to the data to quantify the effects of temperature and time (fixed factors) on all response variables measured in this experiment (cell density, F_V/F_M, ROS, DMS, DMSP and DMSO concentrations) using IBM SPSS Statistics 20. Assumptions of sphericity were tested using Mauchly's test. In cases where this assumption was violated, the degrees of freedom were adjusted using the Greenhouse–Geisser correction factor. Bonferroni adjustments were used for pairwise comparisons. Each variable was tested for the assumption of normality and log, ln or square root transformations were applied when necessary.

For sequencing data, alpha diversity parameters of the rarefied sequences and jackknife comparison of the weighted sequence data (beta diversity) were calculated in QIIME (Caporaso et al., 2010). A two-way PERMANOVA with Bray–Curtis similarity measurements was performed on abundance data of taxonomic groups that contained more than 1 % of total generated OTUs (represent 90.23 % of the data) using PAST (Hammer et al., 2008). In addition, PAST was used to perform non-metric multidimensional scaling (nMDS) analysis and to isolate the environmental parameters (normalised

Table 1. Output of repeated measures analysis of variance (rmANOVA) for algal (CELLS$_A$) and bacterial (CELLS$_B$) cell abundance, photosynthetic efficiency (F_V/F_M), oxidative stress (ROS), dimethyl sulfoniopropionate (DMSP), dimethyl sulfide (DMS) and dimethyl sulfoxide (DMSO) concentrations as a function of temperature (24 or 32 °C) and time. Numbers in bold indicate significant data based on the level of significance $p < 0.05$. df1 indicates the numerator df, and df2 indicates the denominator df.

Parameters		24 °C – mild thermal stress			32 °C – acute thermal stress		
		Temperature	Time	Temperature × time	Temperature	Time	Temperature × time
CELLS$_A$	F	4.04	**335**	**4.16**	**27.47**	**237.62**	**8.28**
	df1	1	**4**	**4**	**1**	**2.04**	**2.04**
	df2	6	**24**	**24**	**6**	**12.26**	**12.26**
	p	0.91	**< 0.001**	**0.01**	**< 0.001**	**< 0.001**	**0.005**
CELLS$_B$	F	2.13	**52.2**	1.35	**32.56**	**199.8**	**22.26**
	df1	1	**1.29**	1.29	**1**	**4**	**4**
	df2	6	**7.74**	7.74	**6**	**24**	**24**
	p	0.2	**< 0.001**	0.3	**0.001**	**< 0.001**	**< 0.001**
F_V/F_M	F	**0.42**	**33.43**	**6.9**	**48.79**	**12.58**	**13.11**
	df1	**1**	**4**	**4**	**1**	**1.19**	**1.19**
	df2	**6**	**24**	**24**	**5**	**5.93**	**5.93**
	p	**0.54**	**< 0.001**	**0.001**	**0.001**	**0.01**	**0.01**
ROS	F	**37.26**	**6.3**	**5.88**	**33.23**	**8.85**	**8.41**
	df1	**1**	**4**	**4**	**1**	**2.32**	**2.32**
	df2	**6**	**24**	**24**	**6**	**13.9**	**13.9**
	p	**0.001**	**0.001**	**0.002**	**0.001**	**0.003**	**0.003**
DMSP	F	0.79	**31.16**	0.95	3.03	**15.18**	3.17
	df1	1	**1.56**	1.56	1	**4**	**4**
	df2	6	**9.35**	9.35	6	**24**	**24**
	p	0.41	**< 0.001**	0.4	0.13	**< 0.001**	**0.03**
DMS	F	**51.5**	**38.73**	2.01	5.08	**30.77**	**5.23**
	df1	**1**	**2.14**	2.14	1	**4**	**4**
	df2	**6**	**12.87**	12.87	6	**24**	**24**
	p	**< 0.001**	**< 0.001**	0.17	0.07	**< 0.001**	**0.004**
DMSO	F	**36.56**	**26.64**	**7.21**	4.68	**14.74**	**7.14**
	df1	**1**	**4**	**4**	1	**4**	**4**
	df2	**6**	**24**	**24**	6	**24**	**24**
	p	**0.001**	**< 0.001**	**0.001**	0.07	**< 0.001**	**0.001**

as follows: (x mean)/SD) that contributed the most to the differences between groups using the Bray–Curtis similarity measure. SIMPER analysis performed with the White t test was used to identify the taxonomic groups that significantly contributed the most to the shift in bacterial composition in *A. minutum* cultures over time and between temperature treatments.

3 Results

### 3.1 Algal growth and physiological response

A. minutum cell abundance exponentially increased over time in both the control (20 °C) and 24 °C temperature treatment, but a significantly faster growth rate ($p = 0.001$, t test) occurred at 24 °C (2.66 ± 0.01 d^{-1}; average ± SE) compared

to the 20 °C control (2.57 ± 0.01 d^{-1}), resulting in significantly greater cell abundance at 96 h ($p = 0.007$) and 120 h ($p < 0.001$) (rmANOVA, Table 1, Fig. 1a). On the other hand, the 32 °C treatment resulted in decreased growth rates (2.40 ± 0.02 d^{-1} versus 2.58 ± 0.02 d^{-1}; t test) and significantly lower cell abundance, relative to the 20 °C control, at all time points from 6 h after the start of the experiment ($p \leq 0.03$; rmANOVA, Table 1, Fig. 1b). *A. minutum* abundance demonstrated a marked decline on day 5 in the 32 °C treatment.

No significant difference in the maximum quantum yield (F_V/F_M) of *A. minutum* cultures occurred between 20 and 24 °C until 120 h after the start of the experiment, where a significantly lower F_V/F_M occurred in the 24 °C treatment ($p = 0.01$; rmANOVA, Table 1, Fig. 2a). In contrast, F_V/F_M was significantly lower in *A. minutum* cultures maintained at
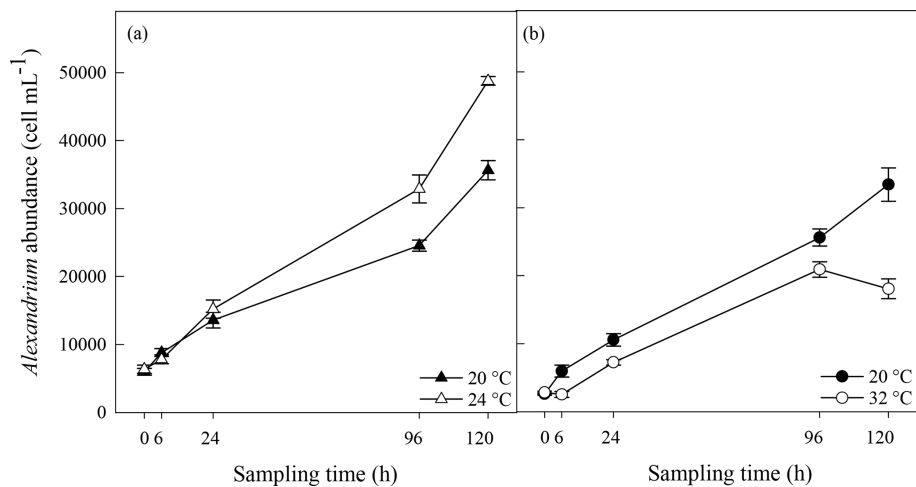
Figure 1. Algal cell abundance in *A. minutum* cultures in experiment 1 (20 and 24 °C) **(a)** and experiment 2 (20 °C and 32 °C) **(b)**; average ± SE, $n = 4$.
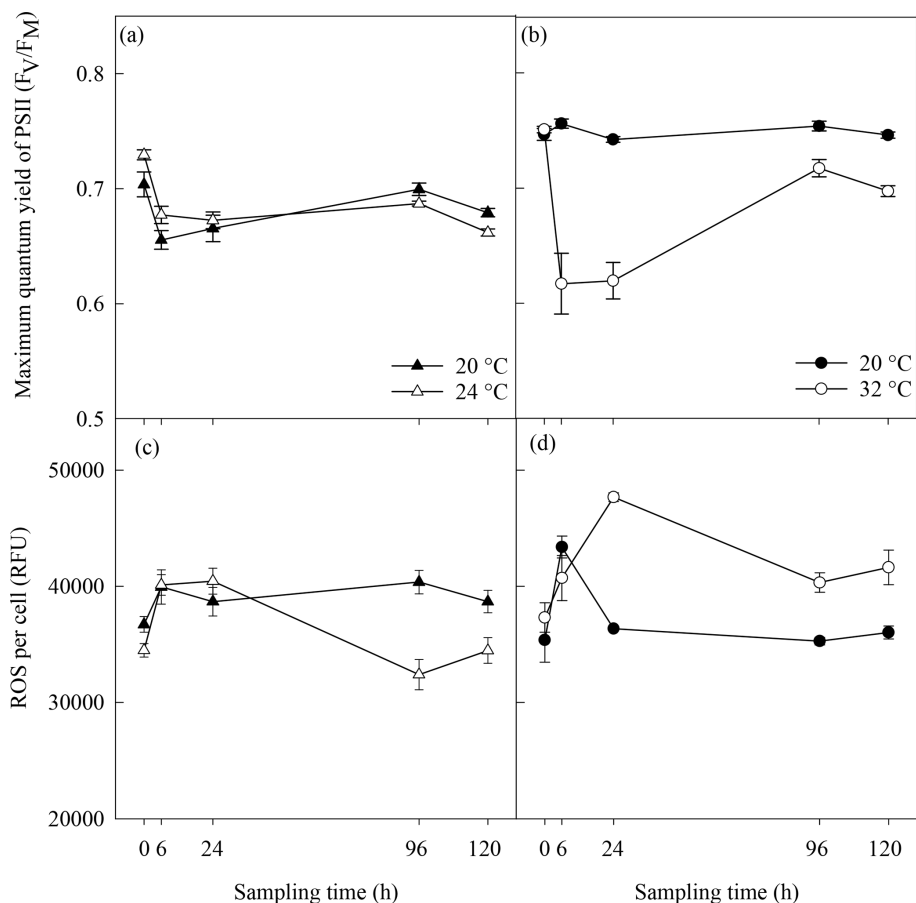
Figure 2. Photosynthetic efficiency **(a, b)** and reactive oxygen species (ROS) **(c, d)** in *A. minutum* cultures in experiment 1 (20 and 24 °C) **(a, c)** and experiment 2 (20 and 32 °C) **(b, d)**; average ± SE, $n = 4$.
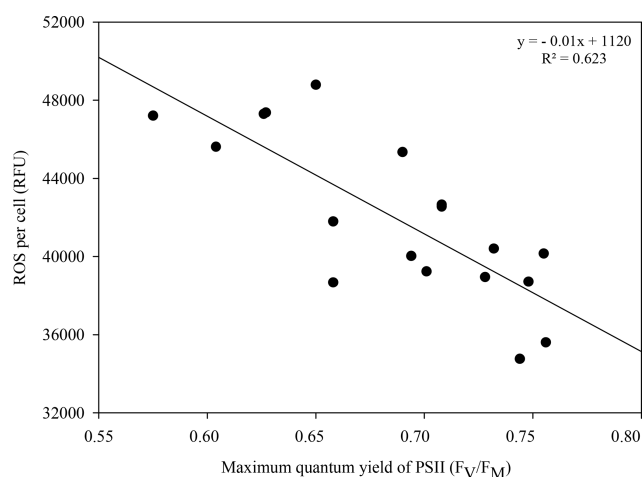
Figure 3. Correlation between the photosynthetic efficiency and reactive oxygen species (ROS) in *Alexandrium minutum* under the 32 °C thermal stress treatment; $n = 18$.

32 °C compared to the 20 °C control at all time points from 6 h after the start of the experiment ($p \leq 0.01$; rmANOVA, Table 1, Fig. 2b). However, on days 5 and 6, the F_V/F_M of cultures kept at 32 °C recovered to values (0.72 ± 0.008) close to those of the control (0.75 ± 0.004) (Fig. 2b), although it remained significantly lower than at 20 °C ($p < 0.01$ and $p < 0.001$ on days 5 and 6 respectively).

3.2 Reactive oxygen species (ROS)

Significantly lower concentrations of ROS were measured at 24 °C than at 20 °C at 96 h ($p = 0.003$) and 120 h ($p = 0.03$) (rmANOVA, Table 1, Fig. 2c). In contrast, significantly greater concentrations of ROS were measured at 32 °C than at 20 °C 24 h ($p < 0.001$), 96 h ($p = 0.001$) and 120 h ($p = 0.01$) after the start of the experiment (rmANOVA, Table 1, Fig. 2d). In line with the recovery in measured F_V/F_M, ROS concentrations in cultures kept at 32 °C started to decline to values closer to those of the control on days 5 and 6 of the experiment (Fig. 2d). A significant negative correlation between F_V/F_M levels and ROS concentrations was observed under the 32 °C temperature treatment ($R^2 = 0.623$; $p = 0.02$, $n = 18$; Fig. 3).

3.3 Biogenic sulfur dynamics

Biogenic concentrations of DMSP, DMS and DMSO ranged from 424 ± 35 to 1629 ± 170 fmol cell^{-1}, from 13 ± 1.02 to 87 ± 5 fmol cell^{-1} and from 9 ± 1.41 to 94 ± 24 fmol cell^{-1}, respectively, over both experiments (Fig. 4). Concentrations of all three sulfur compounds slowly decreased over time in all *A. minutum* cultures regardless of the temperature treatment. No significant difference in DMSP concentration was recorded between 20 and 24 °C throughout the experiment ($p > 0.05$; rmANOVA, Table 1, Fig. 4a), whereas signifi-

cantly less DMSP was measured in cells at 32 °C than in the 20 °C control at 96 h ($p = 0.02$; rmANOVA, Table 1, Fig. 4b).

Significantly lower DMS concentrations were measured at 24 °C compared to 20 °C at 24 h ($p < 0.001$) and 120 h ($p = 0.002$) (rmANOVA, Table 1, Fig. 4c). In contrast, DMS was significantly higher at 32 °C than 20 °C 6 h after the start of the experiment ($p = 0.008$; rmANOVA, Table 1, Fig. 4d). A similar pattern was observed for DMSO, which, relative to the controls, was significantly lower at 24 °C 24 h after the start of the experiment ($p = 0.001$; rmANOVA, Table 1, Fig. 4e) and significantly greater at 32 °C after 6 and 24 h ($p < 0.05$, Fig. 4f).

3.4 Bacterial abundance and composition

Bacterial cell abundance exponentially increased over time at both 20 and 24 °C (Fig. 5a). Bacterial abundance was significantly greater at 24 °C than at 20 °C 120 h after the start of the experiment ($p = 0.05$; rmANOVA, Table 1, Fig. 5a). However, no significant difference ($p > 0.05$, t test) in bacterial growth rate was observed between 20 °C (4.15 ± 0.05 d^{-1}) and 24 °C (4.18 ± 0.01 d^{-1}). In contrast, bacterial growth rate was significantly lower at 32 °C than in the 20 °C control (4.05 ± 0.01 d^{-1} versus 4.23 ± 0.02 d^{-1}; $p < 0.001$, t test) (Fig. 5b), resulting in significantly lower bacterial cell densities at 24 h ($p = 0.002$), 96 h ($p = 0.002$) and 120 h ($p < 0.001$) relative to the control (rmANOVA, Table 1, Fig. 5b).

The composition of the initial (T_0) *A. minutum* microbiome was consistent across all samples, but then diverged significantly with time and between temperature treatments (Fig. 6a–b; Bray–Curtis similarity measurement, Shepard plot stress $= 0.0587$). A significant temporal shift in bacterial composition occurred at both 20 and 32 °C, with dissimilarities in community composition between T_0 and T_{120} of 27 % and 42 % occurring respectively (SIMPER analysis). Notably, bacterial communities at 32 °C differed significantly (two-way PERMANOVA; $p < 0.05$) to 20 °C at T_{120}, with 32 % dissimilarity in community composition. These differences were primarily driven by increased relative abundance of bacterial operational taxonomic units (OTUs) within the *Oceanicaulis* (17 %), Phycisphaeraceae *SM1A02* (8.8 %) and *Balneola* (4.9 %) along with a decline in the relative abundance of OTUs matching *Maribacter* (24 %), *Marinoscillum* (4.7 %) and *Seohaeicola* (2.7 %) (Rhodobacteraceae family) in the 32 °C treatment (White test, Fig. 6c), with all taxa cumulatively contributing to 63 % of the OTU differences between temperature treatments at T_{120} (SIMPER analysis). In the 32 °C treatment, differences in microbiome composition between T_0 and T_{120} were aligned with the elevated levels of ROS, while in the control (20 °C) the community shift was principally aligned with differences in bacterial and algal cell abundance (Fig. 6a; MDS analysis). Similarly, the elevated concentration of ROS as well as the lower F_V/F_M, lower al-
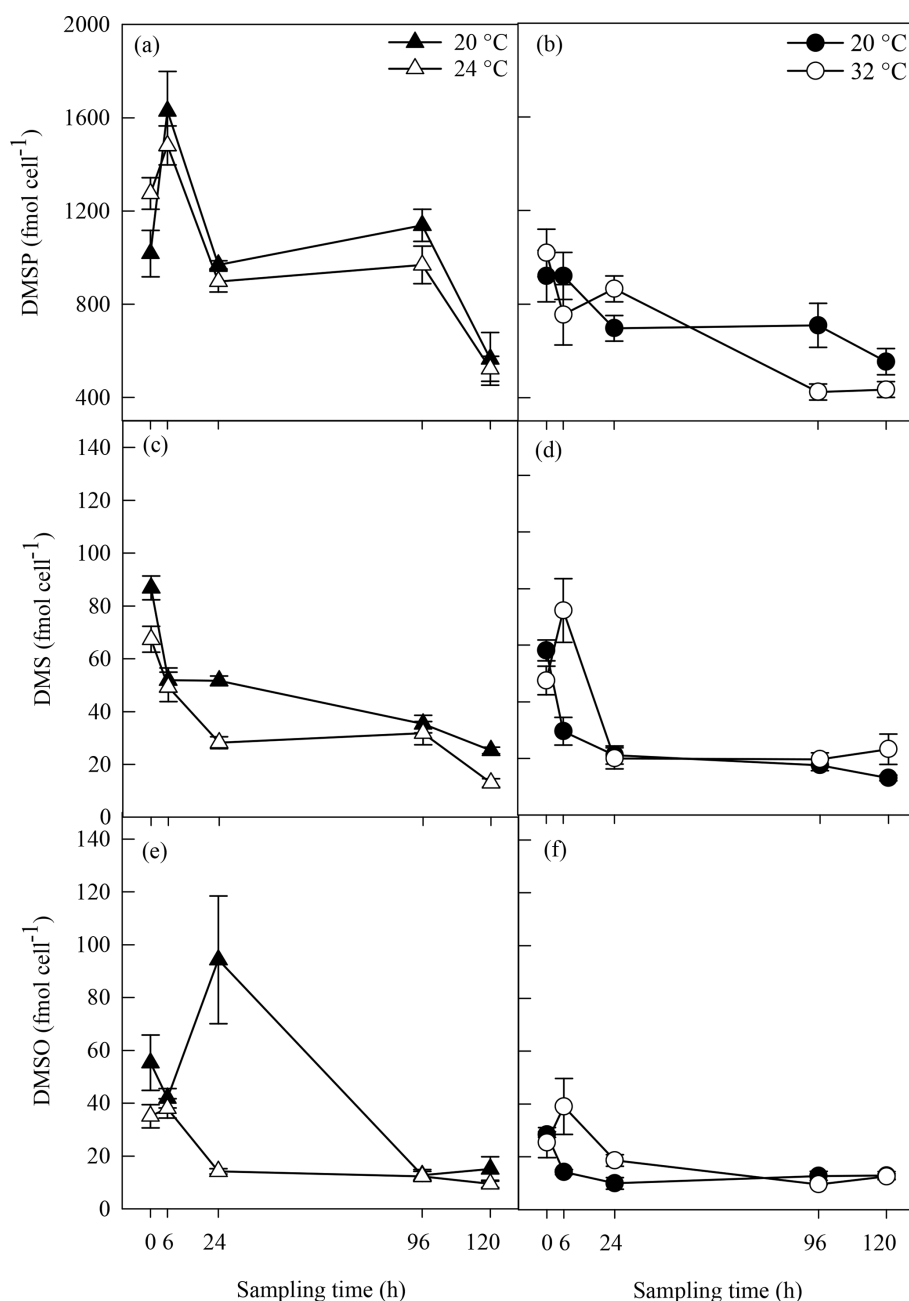
Figure 4. DMSP **(a, b)**, DMS **(c, d)** and DMSO **(e, f)** concentrations in *A. minutum* cultures in experiment 1 (20 and 24 °C) **(a, c, e)** and experiment 2 (20 and 32 °C) **(b, d, f)**; average ± SE, $n = 4$. Variability in between the two 20 °C controls is probably a consequence of experiments 1 and 2 being conducted at different times (April and June), whereby changes in the physiological state of the culture at each time led to different dimethylated sulfur profiles.

gal and bacterial cell abundance and lower DMSP, DMS and DMSO concentrations at 32 °C were aligned with the differences in microbiome composition between the temperature treatments (Fig. 6b; MDS analysis).

4   Discussion

Climate change induced shifts within marine ecosystems are predicted to fundamentally alter the physiology of planktonic organisms and the biogeochemical transformations that they mediate (Finkel et al., 2009; Tortell et al., 2008; Hallegraeff, 2010). Rising seawater temperatures are one of the major impacts of climate change on marine ecosystems (Harley
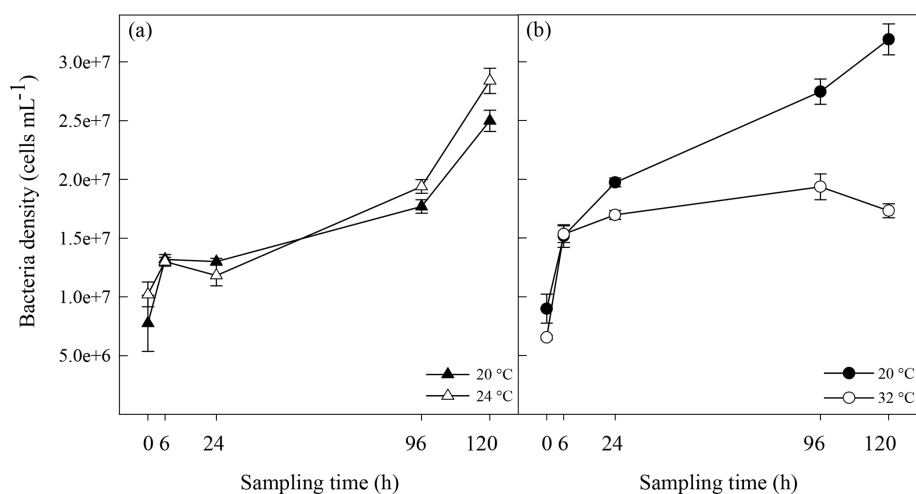
Figure 5. Bacterial cell abundance in *A. minutum* cultures in experiment 1 (20 and 24 °C) **(a)** and experiment 2 (20 and 32 °C) **(b)**; average ± SE, $n = 4$.
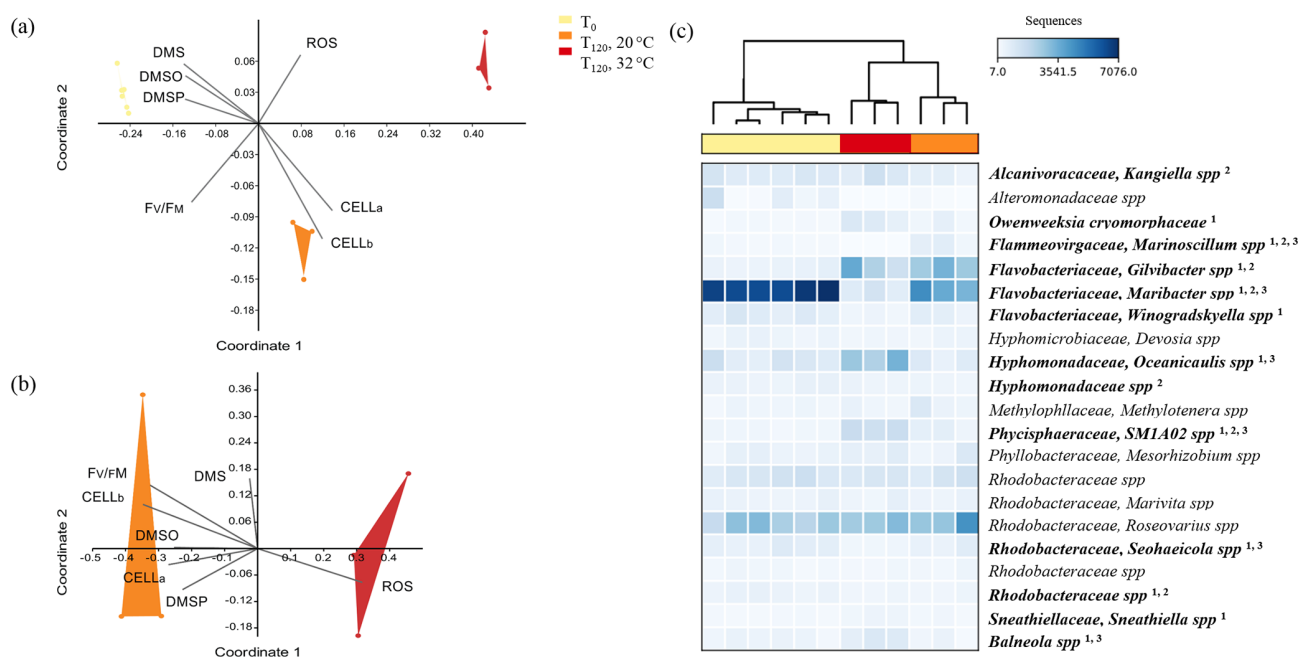
Figure 6. Multi-dimensional scaling (MDS) of the three phylogenetic groups defined by 16S rRNA sequencing of the bacteria population associated with *A. minutum* cultures grown under control conditions (20 °C) and acute thermal stress (32 °C) at T_0 and T_{120} **(a)** and MDS excluding the T_0 control **(b)**. Vectors represent the factors that most likely drove the shift in bacterial composition between groups. The taxonomic groups that significantly contributed to the difference in bacterial composition between T_0 and T_{120} at 32 °C ([1]), between T_0 and T_{120} at 20 °C ([2]) and between 32 and 20 °C at T_{120} ([3]) appear in bold next to the heatmap **(c)**, with scaling based on relative abundance.

et al., 2006), and they can be manifested both as long-term gradual increases (IPCC, 2007, 2013) or intense episodic marine heatwaves (Frölicher and Laufkötter, 2018; Hobday et al., 2016). Although less examined to date than chronic temperature increases, MHWs are predicted to become more frequent and severe (Oliver et al., 2018) and have been proposed as a mechanism for triggering toxic algal blooms (Ummenhofer and Meehl, 2017). Against this backdrop of changing environmental conditions, microbial production and cycling of dimethylated sulfur compounds could be particularly relevant because they simultaneously play a role in the stress response of marine phytoplankton (Berdalet et al., 2011; Deschaseaux et al., 2014a; Sunda et al., 2002; Wolfe et al., 2002; Stefels and van Leeuwe, 1998) and have been predicted to have biogeochemical feedback effects that are relevant for local climatic processes (Charlson et al., 1987).

This study investigated the biogenic sulfur cycling dynamics of *A. minutum*, and its microbiome, in response to an extreme, short-term thermal stress event, akin to the marine heatwave events occurring with increasing frequency within coastal habitats (Oliver et al., 2018). Indeed, MHWs have been defined as an abrupt and ephemeral increase in temperature of at least 3 to 5 °C above climatological average that lasts for at least 3 to 5 d (Hobday et al., 2016). Large increases in temperature of about 8 °C above the monthly climatological average led to red tides of exceptional density in San Francisco Bay (Cloern et al., 2005). While a 12 °C increase in temperature constitutes an extreme scenario of MHWs even for coastal habitats, this experimental temperature was selected after preliminary investigations with the intention to induce thermal stress in this particularly robust strain of *A. minutum* in culture.

A. minutum has been targeted in this study as (1) an ecologically relevant phytoplankton responsible for some of the most harmful algal blooms (Anderson et al., 2012) and (2) as biochemically relevant for containing the highest DMSP concentrations ever reported in marine dinoflagellates (Caruana and Malin, 2014). However, it is to be noted that DMSP concentrations reported in this study were a degree of magnitude lower (0.42 ± 0.04 to 1.63 ± 1.70 pmol cell^{-1}) than that previously reported for *A. minutum* (14.2 pmol cell^{-1}; Caruana and Malin, 2014; Jean et al., 2005). This is potentially because this culture of *A. minutum* had been isolated from free-living *A. minutum* for a long time (since 1988) or because culturing conditions failed to mimic the exact same biochemical conditions in which this strain of *A. minutum* usually grow. This biochemical difference could potentially reflect that this strain of *A. minutum* in culture was more robust than free-living dinoflagellates of the same species, thereby potentially justifying the need of a 12 °C increase in temperature to induce thermal stress.

4.1   Effects of thermal stress on *A. minutum* growth, physiology and ROS production

A mild increase in temperature (4 °C) resulted in faster algal growth and lower oxidative stress, indicating that 24 °C was close to a temperature optimum for this strain of *Alexandrium*. This is perhaps not surprising considering that *Alexandrium* species are capable of growing under a wide range of temperatures from 12 to 25 °C (Laabir et al., 2011). In contrast, an extreme increase in temperature (12 °C) resulted in a rapid and clear cascade of physiological responses, indicative of an acute thermal stress response in *A. minutum*. Overall, *A. minutum* cells exposed to 32 °C immediately exhibited slower growth relative to the 20 °C control, suggesting that a 12 °C increase in temperature rapidly led to either an increase in cell death rate or a decrease in cell division (Rajadurai et al., 2005; Veldhuis et al., 2001). The slower growth rate at 32 °C was coupled with a drop in photosynthetic efficiency and an increase in ROS concen-

trations, which are both common stress responses to thermal stress in marine algae (Lesser, 2006; Falk et al., 1996; Robison and Warner, 2006; Iglesias-Prieto et al., 1992). In fact, these two physiological responses are often interconnected as increased ROS production generally occurs in both the chloroplast and mitochondria of marine algae exposed to thermal stress, causing lipid peroxidation and ultimately leading to a loss in thylakoid membrane integrity (Falk et al., 1996) and a decrease in the quantum yield of PSII (Lesser, 2006). This was reflected in the negative correlation observed between the maximum quantum yield of PSII and ROS concentrations.

Although photosynthetic efficiency remained impaired and ROS concentrations remained high under 32 °C until the end of the experiment, both biomarkers of stress started to return to values closer to those of the 20 °C control by days 5 and 6 of the experiment. This was most likely at the expense of a decline in algal abundance since slow growth often coincides with concurrent cellular repair and photosystem activity recovery (Robison and Warner, 2006). The differential physiological response between 24 and 32 °C indicates that although cultures of this strain of *A. minutum* appear to be highly resistant to temperature changes, an abrupt increase in temperature of 12 °C simulating an extreme case of marine heatwave led to a clear stress response. The physiological pattern at 32 °C also suggested an acclimation period necessary for such an abrupt shift in temperature, especially since recovery (in F_V/F_M and ROS levels) was observed towards the end of the experiment.

4.2   Biogenic sulfur cycling as a response to thermal stress in *A. minutum*

Biogenic organic compounds are key compounds in the stress response of phytoplankton, with evidence they can be used in response to changes in temperature (Van Rijssel and Gieskes, 2002; Stefels, 2000). An upregulation of the biogenic sulfur yield was expected as a stress response to increased temperature in *A. minutum*, through either an increase in cellular DMSP concentrations, or an increase in DMS via the cleavage of DMSP (McLenon and DiTullio, 2012; Berdalet et al., 2011; Wolfe et al., 2002; Sunda et al., 2002). No significant change in DMSP concentrations was observed between the control and 24 °C treatment, where, as described above, physiological responses converged to indicate that 24 °C was in fact a more optimal growth temperature for this organism. This temperature optimum was generally associated with lower DMS and DMSO concentrations than in the 20 °C control, although this was only evident 24 h after the start of the experiment. Since algal stress responses often result in increased cellular sulfur concentrations in dinoflagellates (McLenon and DiTullio, 2012; Berdalet et al., 2011), it is perhaps not surprising that DMS and DMSO concentrations were lower under what appear to have been more optimal growth temperature conditions.

In contrast to the lower DMS and DMSO concentrations observed at 24 °C compared to the 20 °C control, exposure to 32 °C resulted in spikes in DMS and DMSO 6 h after the start of the experiment, which accompanied decreased algal growth and an impaired photosystem II. Although sporadic, the increases in DMS and DMSO observed in the 32 °C treatment may have resulted from enhanced intracellular DMSP cleavage by phytoplankton (Del Valle et al., 2011) or enhanced DMSP exudation from phytoplankton cells during cell lysis (Simó, 2001), resulting in an increasing pool of dissolved DMSP made readily available to both bacteria and phytoplankton DMSP lyases (Riedel et al., 2015; Alcolombri et al., 2015; Todd et al., 2009, 2007). However, it is notable that lower DMSP concentrations in the 32 °C treatment than in the control only occurred on day 4, whereas the spikes in DMS and DMSO were evident at the outset of the experiment (6 h). Since this decrease in DMSP at 96 h was not coupled with an increase in DMS, this could alternatively be indicative of a decrease in methionine synthase activity (McLenon and DiTullio, 2012) or assimilation of DMSP sulfur by bacterioplankton for de novo protein synthesis (Kiene et al., 2000), with this demethylation pathway often accounting for more than 80 % of DMSP turnover in marine surface waters. The spike in DMSO measured 6 h after the increase in temperature to 32 °C most likely indicated rapid DMS oxidation by ROS under thermal stress (Sunda et al., 2002; Niki et al., 2000). At that time however, we found no evidence for ROS build-up in *A. minutum* cultures, possibly because ROS concentrations were kept in check by sufficient DMS synthesis and active DMS-mediated ROS scavenging (Lesser, 2006; Sunda et al., 2002). In contrast, 24 h after the start of the experiment, increased ROS coincided with an abrupt decline in DMS and DMSO, perhaps suggestive of serial oxidation via active ROS scavenging of both DMS to DMSO and DMSO to methane sulfinic acid (MSNA) (Sunda et al., 2002), although it is always difficult to confidently link DMS(O) and ROS dynamics unless using tracing techniques. It is also to be noted that measuring standing stocks may constitute a limitation to the capturing of subtle changes in DMS, DMSP and DMSO over time.

The only previous study that has examined sulfur responses to stress exposure in *A. minutum* examined the effect of physical turbulence by shaking *A. minutum* cultures for up to 4 d (Berdalet et al., 2011). While the authors of that study also observed slower cell growth as a response to stress exposure, in contrast to our study, cellular DMSP concentrations increased by 20 %. Here, a drop in DMSP concentration was observed at 96 h between the control and temperature treatment. Therefore, even though DMSP concentrations were quantified with a similar approach as in Berdalet et al. (2011) (no filtration of the samples with the assumption that particulate DMSP concentrations overrule dissolved DMSP and DMS concentrations), it seems that heat stress and turbulence triggered a dissimilar sulfur response to stress in *A. minutum*.

Overall, a 12 °C increase in temperature led to lower photosynthetic efficiency, increased oxidative stress and slower cell growth in this robust strain of the red-tide mediating dinoflagellate *A. minutum*. This physiological stress response was coupled with a differential biogenic sulfur cycling as shown by spikes in DMS and DMSO as well as lower DMSP concentrations, most likely translating ROS scavenging and DMSP uptake by bacterioplankton respectively. Because the turnover of DMS, DMSP and DMSO in biological systems can occur very quickly (Simó et al., 2000), DMS and DMSO concentrations can change rapidly, which sometimes makes it difficult to clearly establish cause–effect relationships between physiological stress and the biogenic sulfur response.

4.3 A shift in *A. minutum*-associated bacteria composition triggered by thermal stress

In light of DMSP and related biogenic sulfur compounds constituting an important source of carbon and sulfur to phytoplankton-associated bacteria (Kiene et al., 2000), it follows that any shift in biogenic sulfur concentrations could influence the microbiome composition of *A. minutum*. However, it is undeniable that a shift in the microbial community could also be driven by a range of physiological and biochemical parameters that were not measured in this study. Nevertheless, the most pronounced temporal shift in the composition of the bacterial community associated with *A. minutum* occurred in the 32 °C treatment. This shift was primarily characterised by a statistically significant increase in the relative abundance of OTUs classified as members of the *Oceanicaulis*, Phycisphaeraceae and *Balneola* and a significant decrease in OTUs classified as members of the *Maribacter*, *Marinoscillum* and *Seohaeicola*.

To predict any potential roles of these key OTUs in biogenic sulfur cycling processes, we screened the genomes of members of these groups using BLAST for four genes commonly involved in DMSP catabolism: *dmdA*, GenBank ID CP000031.2 (Howard et al., 2006); *dddP*, GenBank ID KP639186 (Todd et al., 2009); *tmm*, GenBank ID JN797862 (Chen et al., 2011) and *dsyB*, GenBank ID KT989543 (Curson et al., 2017). A BLAST query of the sequences in the NCBI nucleotide collection (nr/nt) database revealed that previously sequenced members of the genera *Maribacter* (taxid:252356, 357 sequences), *Oceanicaulis* (taxid:153232, 36 sequences), *Marinoscillum* (taxid:643701, 23 sequences), *Seohaeicola* (taxid:481178, 18 sequences) and *Balneola* (taxid:455358, 44 sequences) did not possess any homologs of these sulfur cycling genes. While no homologs were found in the genus *SM1A02*, perhaps because very little genomic information is available for this genus, a close phylogenetic relative to *SM1A02* (99 % query cover, 80 % identical, E value = 0.0), and also a member of the Phycisphaeraceae family (*P. mikurensis* 10 266; GenBank accession numbers AP012338.1), possessed significant homologs to all four query genes involved in DMSP metabolism: *dmdA* (92 %

identical, E value < 0.001), $dddP$ (87 % identical, E value $= 0.003$), tmm (82 % identical, E value $= 0.002$) and $dsyB$ (92 % identical, E value < 0.001). It is thus possible that the spike in DMS and DMSO concentrations in the early stage of the 32 °C heat treatment was a consequence of (or contributed to) the preferential recruitment of Phycisphaeraceae *SM1A02*.

Some members of the Rhodobacteraceae family such as several members of the *Roseobacter* genus and *Rhodobacter sphaeroides* are known to possess homologs of either or both *dmdA* and *ddd* genes, which are responsible for DMSP demethylation and DMSP-to-DMS cleavage respectively (Howard et al., 2006; Curson et al., 2008). However, none of the available reference genomes for *Seohaeicola*, a member of the *Rhodobacteracea*, possessed any homologs of targeted biogenic sulfur cycling. Similarly, members of the *Maribacter*, which was the main contributor to the difference in microbiome structure between the control and thermal stress treatment, are known not to possess DMSP or DMS transformation pathways (Kessler et al., 2018). Hence, the decline of this taxa in the heat stress treatments, where an upshift in biogenic sulfur availability occurred, is perhaps not surprising. However, this change in microbial abundance could have also been triggered by a range of other parameters that were not measured in this study.

Ultimately, the rapid changes in DMS and DMSO concentrations were potentially caused by (or led to) a shift in microbiome composition towards the preferential growth of sulfur-consuming bacteria (e.g. Phycisphaeraceae *SM1A02*) at the expense of other types of bacteria (e.g. *Seohaeicola*). Alternatively, the observed shifts in microbiome structure may have occurred independently to the biogenic sulfur cycling processes and were instead related to either temperature itself or other metabolic shifts in the heat-stressed *A. minitum*. Notably, the temporal shift in bacterial composition under thermal stress was associated with increased cellular ROS at the end of the experiment, indicating a potential link to oxidative stress.

5 Conclusion

Abrupt and intense increases in seawater temperatures associated with MHWs are predicted to become more frequent and intense (Oliver et al., 2018) and have the potential to influence the structure of coastal microbial assemblages and the nature of the important biogeochemical processes that they mediate. Here, we hypothesised that a very acute increase in temperature, mimicking an extreme scenario of coastal MHWs, would trigger both a physiological and biochemical stress response in the DMSP-producing dinoflagellate *A. minutum*. This response was indeed observed following a 12 °C increase in temperature, with evidence for impaired photosynthetic efficiency, oxidative stress, spikes in DMS and DMSO concentrations, a drop in DMSP concen-

tration and a shift in the composition of the *A. minutum* microbiome. These patterns are indicative of a profound shift in the physiological state and biochemical function of this ecologically relevant dinoflagellate in the context of MHWs and suggest that extreme thermal stress has the potential to not only influence the composition and interactions of coastal microbial food webs, but reshape sulfur budgets in coastal waters.

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