

Elsevier required licence: © <2020>. This manuscript version is made available under the CC-BY-NC-ND 4.0 license <http://creativecommons.org/licenses/by-nc-nd/4.0/>
The definitive publisher version is available online at <https://doi.org/10.1016/j.cep.2020.108112>

1 **Comparison of dual stage ultrafiltration and hybrid**
2 **ultrafiltration-forward osmosis process for harvesting microalgae**
3 **(Tetraselmis sp) biomass**

4 MhdAmmar Hafiz^a, Alaa H. Hawari^{a,*}, Probir Das^b, Ali Altaee^c, Adnan Alhathal Alanezi^d

5
6
7 ^aDepartment of Civil and Architectural Engineering, Qatar University, P.O. Box 2713, Doha,
8 Qatar.

9 ^b Algal Technology Program, Center for Sustainable Development, College of Arts and
10 Sciences, Qatar University, Qatar

11 ^cSchool of Civil and Environmental Engineering, University of Technology in Sydney, 15
12 Broadway, Ultimo, NSW 2007, Australia

13 ^dDepartment of Chemical Engineering Technology, College of Technological Studies, The
14 Public Authority for Applied Education and Training (PAAET), P.O. Box 42345, Shuwaikh
15 70654, Kuwait

16
17
18
19
20 *Corresponding author: Dr. Alaa H. Hawari, Department of Civil and Architectural
21 Engineering, College of Engineering, Qatar University, 2713 Doha, Qatar.
22 Email: a.hawari@qu.edu.qa
23

24
25
26
27

28

29

30

31 **Abstract**

32 In this study, a hybrid ultrafiltration - forward osmosis (FO) process was investigated
33 for harvesting of marine microalgae (*Tetraselmis* sp). FO was applied as a post-treatment
34 process after the ultrafiltration (UF) process to obtain higher harvesting efficiency while
35 consuming less energy. The UF-FO process was tested using a pilot-scale UF and bench-
36 scale FO setups. The FO process assessed the impact of different flow rates, membrane
37 orientation and feed solution concentration on the process performance. A maximum algal
38 harvesting concentration factor (CF) of 7.0 was achieved using ultra-filtrated algae as feed
39 solution in the FO membrane operating in the FO mode at 2.5 LPM flowrate for 48 hours
40 operation time. The total energy consumption decreased by 46% using a hybrid UF-FO
41 process instead of dual-stage UF. The FO process was inefficient for the harvesting of raw
42 microalgae culture with a concentration ≤ 1 g/l. However, The FO process was an energy
43 efficient post-treatment process after ultrafiltration for further harvesting of microalgae cells.

44

45 **Keywords:** Harvesting of Algae; Forward osmosis; Ultrafiltration; Brine reject; Membrane
46 fouling.

47

48 1. Introduction

49 In recent years, microalgal biomass is increasingly considered as a promising
50 alternative raw material source for biofuel production. Microalgae can be used to produce
51 multiple high-value items such as food supplements, cosmetics and pharmaceutical products.
52 However, algae cells are available in a very diluted culture medium, where their density is
53 similar to water. Harvesting of microalgae is the separation of algae cells from the culture
54 solution. Harvesting of microalgae is considered the most challenging constraint to an
55 industrial scale production process because it stipulate 50% of the total energy of the biomass
56 production process (Lei et al., 2015). Harvesting is usually done using conventional processes
57 such as sedimentation, centrifugation, chemical flotation, electrophoresis and coagulation and
58 flocculation (Buckwalter et al., 2013; Mo et al., 2015). Recently, ultrafiltration (UF) (Shao et
59 al., 2015; Zhang et al., 2013) and microfiltration (MF) (Simstich et al., 2012) have been
60 utilized in harvesting microalgae due to their high separation efficiency and ease of use.
61 However, these pressure-driven membrane filtration processes are amenable to fouling,
62 which increases energy costs (Mo et al., 2015; Rickman et al., 2012). Thus, there is a vital
63 need for a sustainable algae harvesting process to overcome the drawbacks of the existing
64 technologies. Forward osmosis (FO) is an evolving membrane filtration process that has the
65 potential to minimize the total cost of harvesting microalgae by utilizing osmotic pressure
66 differences to concentrate microalgae. In the FO process, fresh water transports from the feed
67 to the draw solution side through a semi-permeable membrane (Noffsinger et al., 2009).
68 Various studies showed the effectiveness of the FO process in harvesting microalgae due to
69 its high efficiency and low energy consumption [reference, I suggest “Separation and
70 Purification Technology, Volume 2042 October 2018Pages 154-161]. Table 1 summarizes
71 previous studies on the harvesting of microalgae by the forward osmosis process.

72

Table 1. Previous studies on harvesting microalgae using forward osmosis

Process	Type of microalgae	Membrane	Mode	Feed Volume (L)	DS	Concentration Factor (CF)	Ref.
Forward osmosis	Chlorella vulgaris	Aquaporin-based polyether sulfone (PES)	AL-FS	0.5	Sea water	4	(Munshi et al., 2018)
Forward osmosis	Scenedesmus obliquus	Cellulose triacetate (CTA)	AL-FS	1	Brine solution	3	(Larronde-Larretche & Jin, 2016)
Forward osmosis	Chlorella vulgaris	Cellulose triacetate (CTA)	AL-FS	1	Brine solution	3	(Larronde-Larretche & Jin, 2017)
Forward osmosis	Chlorella vulgaris	Thin film composite (TFC)	AL-FS	0.5	Urine	1.7	(Volpin et al., 2019)
Electrically-facilitated forward osmosis	Chlorella vulgaris	Thin film composite (TFC)	AL-FS	0.05	Brine solution	1.5	(Son et al., 2017)

74

75 Previous studies revealed the influence of membrane orientation and the availability
76 of spacer have significant on the performance of concentrating microalgae by the FO process
77 (Honda et al., 2015). Active layer (AL) facing FS has resulted in a stable flux and better flux
78 recovery by physical cleaning (Honda et al., 2015). Removing the feed spacer improved the
79 harvesting of algal biomass by 27%, using a saline solution (70,000 ppm) DS in the FO
80 process operating in the FO mode (Larronde-Larretche & Jin, 2016). In another study, the
81 FO process was used for dewatering of *Scenedesmus acuminates* suspensions utilizing a
82 polyamide thin film composite (TFC) with an enhanced surface shearing force (Ye et al.,
83 2018). Shear force provided by mechanical stirring improved the harvesting of microalgae
84 and the average water flux increased by 57.5% at a stirring speed of 1000 rpm using 23.0 g/L
85 of microalgal suspension FS and 2 mol/L of $MgCl_2$ DS (Ye et al., 2018). **Previous studies**
86 **have investigated the performance of different draw solutions in the FO process, such as**
87 **seawater (Nguyen et al., 2013),** brine from desalination plants (Thabit et al., 2019),
88 electrolytes (i.e., $MgCl_2$ and NaCl) and thermolytic salt ammonium bicarbonate (Achilli et
89 al., 2010). Desalination brine is a promising draw solution for the concentration of algae cells
90 because of abundancy. In addition, the desalination brine has high osmotic pressure, around
91 54 bars, as a result it induces high osmotic pressure gradient in the process (Singh, 2015).

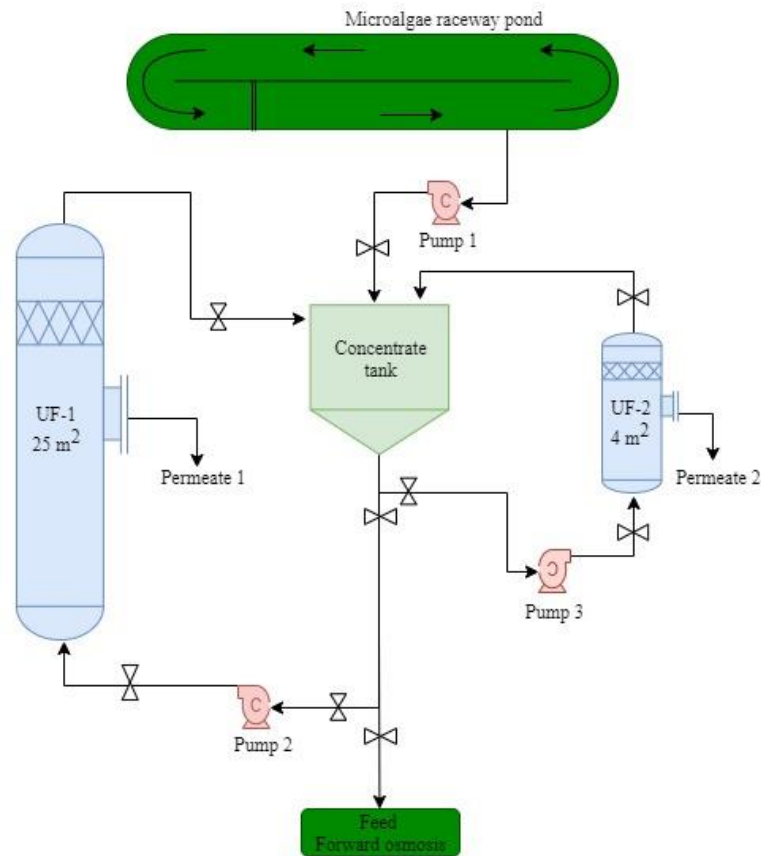
92 The diluted brine can be discharged back to the sea with a minimal impact on the
93 environment.

94 This study evaluates the performance of a hybrid ultrafiltration forward osmosis
95 system for the concentration of algal suspension. The FO process will be used as a post-
96 harvesting process after the ultrafiltration process for microalgae. The effects of different
97 flowrates, feed solution concentration and membrane orientation on harvesting microalgae by
98 the FO process was investigated. The performance of forward osmosis was evaluated by
99 measuring water flux, concentration factor (CF) and reverse solute flux. In addition, the
100 energy consumption of the hybrid system was evaluated.

101 2. Materials and setup

102 2.1 Ultrafiltration setup

103 The ultrafiltration (UF) unit shown in Figure 1 was constructed near the 250 L
104 microalgae raceway tank. The unit consisted of two hollow fiber ultrafiltration modules (UF-
105 1 and UF-2). The hollow fiber within the UF modules was made of polyacrylonitrile. UF-1
106 and UF-2 units have an effective membrane surface area of 25 and 4 m², respectively. The
107 modules were connected to tanks and pumps through a system of interconnected (PVC) pipe
108 fittings and valves. Pump 1 was used for transferring microalgae culture from the raceway
109 tank to the feed tank (concentrate tank), whereas pump 2 and pump 3 served as feed pumps
110 for transferring the culture to the inlet of ultrafiltration modules. The power rating for pump 1
111 was 1.6 kW, whereas the power rating for pump 2 and pump 3 was 1.6 and 0.75 kW,
112 respectively. In the preliminary step, the microalgae culture was concentrated to a desired
113 volume by UF-1 module. The concentrate obtained from the UF-1 module was further
114 concentrated using the UF-2 module. The concentrated microalgae culture obtained from the
115 UF-2 module was collected and stored in a concentrate tank prior to be used as a feed in the
116 FO process.



117

118

Figure 1. Schematic diagram of the ultrafiltration pilot-scale membrane test skid

119

120 2.3 Microalgal cells and growth media

121

122

123

124

125

126

127

128

129

130

131

132

133

This study used a halotolerant *Tetraselmis* sp. (Das et al., 2019). Guillard's f/2 medium was used for growing *Tetraselmis* sp. The used nutrients were of analytical grade except for nitrogen, therefore, commercial-grade urea with 46% of nitrogen was added. Seawater samples were collected from Al Thakhira beach in Qatar and sterilized with a commercial grade 0.02% bleach and kept for one day before using. The used biomass consisted of algae mainly because the algae samples were collected from an open raceway pond where the bacteria growth is minimal. Algal culture samples were checked under a microscope on a regular basis where little or no bacteria contamination was observed. The initial characteristics both algal and ultra-filtered concentrated algae are summarized in Table 2.

134

Table 2. Initial characteristics of the original algae and the ultra-filtered algae.

Parameter	Original Algae	Ultra-filtered Algae	Standard method
Temperature (°C)	21.2±0.1	22.1±0.1	APHA 2550 Temperature
pH	8.2 ± 0.03	8±0.02	APHA 4500-H+ B. Electrometric Method
Salinity (ppt)	67.68±0.01	64.5±0.02	APHA 2520 B. Electrical Conductivity Method

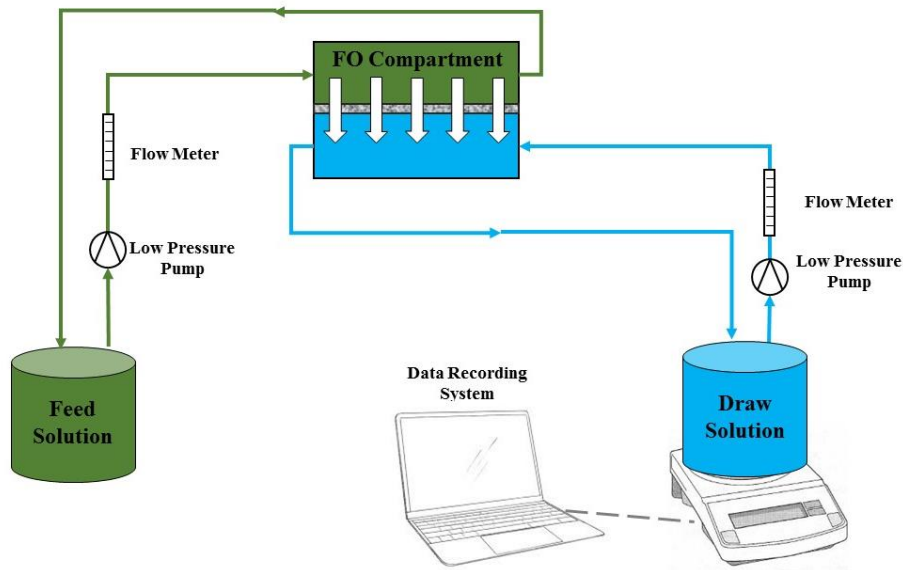
135

136

2.3 Forward osmosis setup

137 A schematic diagram of the FO setup is shown in Figure 2. CF042 Delrin flat sheet
138 forward osmosis membrane cell was purchased from Sterlitech. The cell dimensions are 12.7
139 x 8.3 x 10 cm, and the active inner dimensions are 4.6 x 9.2 cm and a slot depth of 0.23 cm.
140 The membrane separated the feed and draw solutions inside the cell. The feed and draw
141 solutions were supplied by two tanks with a capacity of 6 L each. The feed and draw solution
142 were circulated through the membrane cell using two Cole-Parmer gear pumps. The flow rate
143 of the feed and the draw solutions were measured using two flow meters (Sterlitech Site Read
144 Panel Mount Flow Meter). A digital balance (ICS-241 Mettler Toledo) was used to measure
145 the water flux in the FO system. A known quantity of 2 L was used for the feed and draw
146 solution at the beginning of the experiment. The solutions going out from the FO cell were
147 recycled back into the same tanks. Each experiment was running for 2800 min. After 24h of
148 operation, the membrane was washed using distilled water for 30 min to retrieve the initial
149 water flux. A commercial FO membrane (TFC FO membrane, FTSH2O (USA)) was used.
150 Two feed solutions (FS) were used in the FO process namely; microalgae culture with a
151 concentration of 0.43 g/l and ultra-filtered microalgae with a concentration of 15.7 g/l. The
152 draw solution (DS) in the FO system was a concentrated brine from a desalination plant, TDS
153 ~81 g/L. Brine samples were collected from a thermal desalination plant located in Doha,
154 Qatar. The characteristic of brine is summarized in Table 3.

155



156

157

Figure 2. Schematic diagram of the forward osmosis lab-scale membrane test skid

158

159

Table 3.Characteristics of brine reject from the desalination process (draw solution)

Parameter	Value	Standard Method
TDS (ppm)	81392 ± 5	APHA 2540 C. Total Dissolved Solids Dried at 180 °C
pH	9.03	APHA 4500-H+ B. Electrometric Method
Turbidity (NTU)	0.25 ± 0.1	APHA 2130 B. Nephelometric Method
EC (mS/cm)	93.6 ± 0.2	APHA 2510 B. Conductivity
Chloride (ppm)	$35,266 \pm 0.2$	APHA 4110 Determination of anions by ion chromatography
Bromide (ppm)	120.3 ± 0.2	
Sulfate (ppm)	5032.2 ± 0.2	
Sodium (ppm)	20982.5 ± 0.2	APHA 3120 Determination of metals by plasma emission spectroscopy
Potassium (ppm)	728.6 ± 0.2	
Calcium (ppm)	723.2 ± 0.2	
Magnesium (ppm)	2511.2 ± 0.2	

160

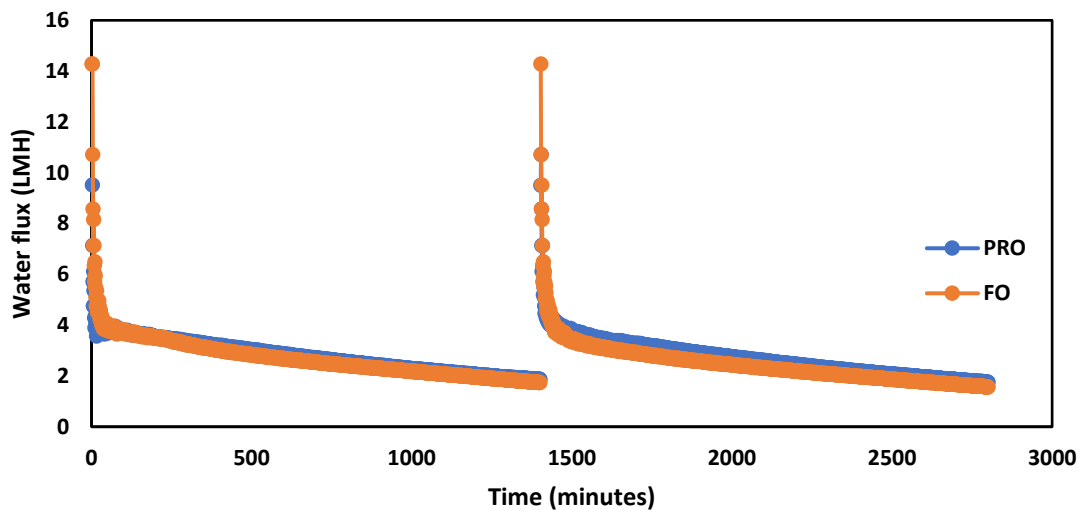
161 3. Results and discussions

162 3.1 Effect of membrane orientation on water flux

163 The water flux (J_w) in the FO process was calculated from the following equation:

$$J_w = \left(\frac{V_p}{A_m \times t} \right) \quad (1)$$

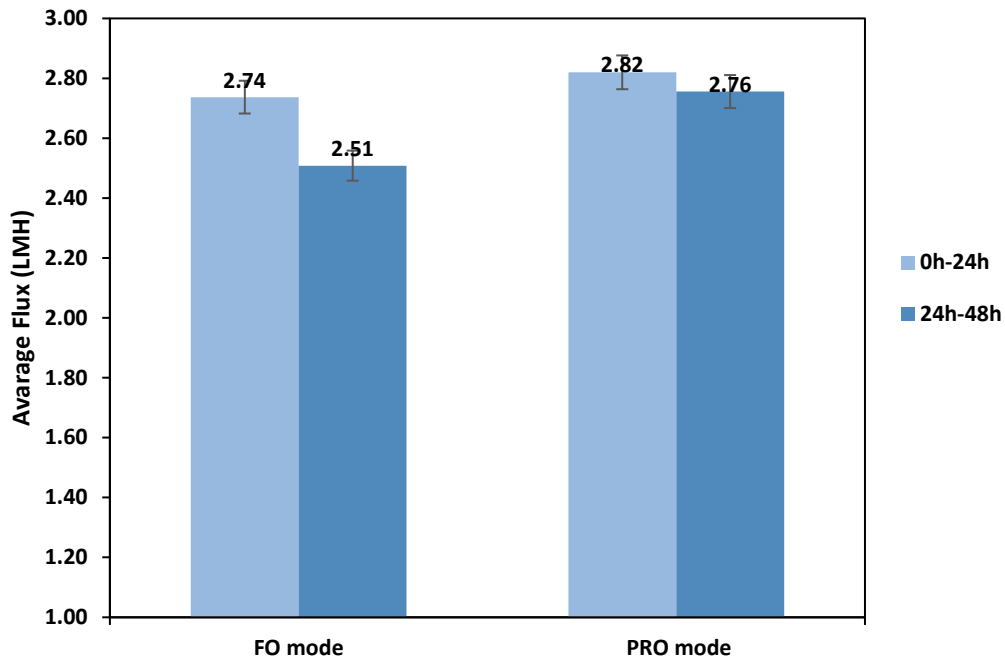
164 Here, V_p is the volume of the permeate (L), A_m is the area of the membrane (m^2), t is the
165 operating time (h). Figure 3 shows the change of water flux with time in the PRO (AL-DS)
166 and FO (AL-FS) modes. Figure 3 shows a gradual decrease in the water flux over time due to
167 membrane fouling, dilution of the DS, and concentration polarization. Algae cells contain
168 lipids, proteins, and carbohydrates, which could accumulate on the surface of the membrane
169 and cause fouling. In the PRO mode, the initial water flux was 9.52 LMH and decreased by
170 80% after 24 h. After washing the membrane with distilled water for 30 minutes, the water
171 flux was retrieved to 9.51 LMH (i.e. almost 100%), then decreased by 83% at the end of the
172 experiment. In the FO mode, the initial water flux was 14.3 LMH and decreased by 88% after
173 24 h. After washing with distilled water for 30 minutes, the retrieved water flux was 14.2
174 LMH (i.e. almost 100%), then decreased by 89% at the end of the experiment. Although the
175 water flux decreased during the FO process in both modes, but it 100% retrieved after
176 washing the membrane with distilled water. This reveals the ability of the FO systems to
177 tolerate biofouling and recover water flux easily after washing. In order to compare the
178 performance of PRO and FO modes, the average water flux was calculated.



179

180 *Figure 3. Water flux using algae as the feed solution and brine as the draw solution in FO mode (active layer*
 181 *facing feed solution) and PRO mode (active layer facing draw solution)*

182 Figure 4 shows the average water flux for PRO and FO modes using a flow rate of 2.5
 183 LPM for both the feed solution and the draw solution. In the FO mode, the average water flux
 184 during the first 24 hours was 2.74 LMH and declined by 9.2% in the second 24 hours after
 185 washing. In the PRO mode, the average water flux was 2.82 LMH in the first 24 hours and
 186 declined by 2.2% in the second 24 hours after washing. The average water flux in the PRO
 187 mode was higher than the FO mode, which is in good agreement with previous studies
 188 (Honda et al., 2015; Mi & Elimelech, 2008; Wang et al., 2010). The reason is that the effect
 189 of dilutive concentration polarization at the draw solution side can be mitigated in the PRO
 190 mode. The denser and smoother surface of the active layer can improve the fluid shear stress
 191 around the membrane surface to reduce diffusion of salt into the membrane and accumulation
 192 of salt on the membrane, therefore, reduce both external concentration polarization and
 193 internal concentration polarization (Mi & Elimelech, 2008; Valladares Linares et al., 2013;
 194 Zhao et al., 2011). As a result, the PRO mode was selected to further evaluate the
 195 performance of the FO process.



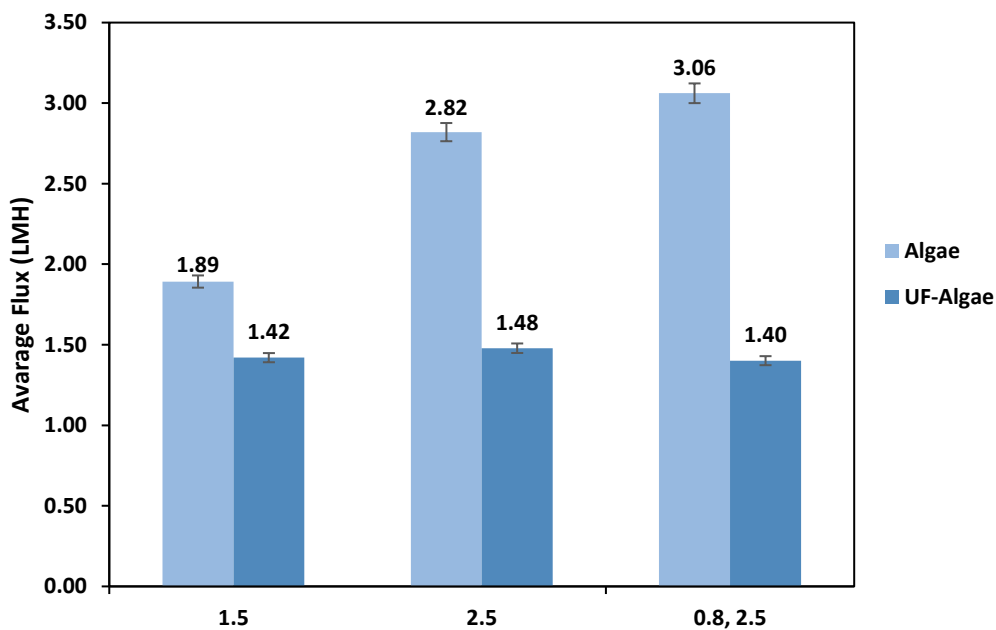
196

197 *Figure 4. Average water flux using algae as the feed and brine as the draw solution in FO and PRO modes*

198 **3.2 Performance of the hybrid ultrafiltration- forward osmosis system**

199 FO performance was evaluated using two different feeds namely: microalgae without
 200 ultrafiltration (i.e. concentration of 0.43 g/l) and concentrated ultrafiltered microalgae (i.e.
 201 concentration of 15.7 g/l). In the FO process, three different flow rates were used 1.5 LPM
 202 for DS and FS (DS:1.5LPM - FS:1.5LPM), 2.5 LPM for DS and FS (DS:2.5LPM –
 203 FS:2.5LPM) and 2.5 LPM for FS and 0.8 LPM for DS (DS:0.8LPM – FS:2.5LPM). It is
 204 obvious from Figure 5 that the average water flux increased with increasing the flow rate of
 205 the DS and FS while using microalgae as a feed solution. The average water flux increased
 206 from 1.89 LMH to 2.82 LMH as the flow rates of the draw and the feed solutions increased
 207 from 1.5 LPM to 2.5 LPM, respectively. The highest average water flux was 3.06 LMH
 208 obtained using a flow rate of 0.8 LPM (DS) and 2.5LPM (FS). The increase of the water flux
 209 with the increase of the DS and FS flow rate is due to the minimized concentration
 210 polarization effect at higher flow rates (McCutcheon & Elimelech, 2006). Concentration
 211 polarization plays a major role in decreasing the osmotic effect across the FO membrane
 212 which would decrease the membrane flux (Devia et al., 2015). As the flow rate increase, the
 213 turbulence around the membrane surface increases, which in return reduces the effect of
 214 concentration polarization and increases the mass transfer coefficient. Furthermore, the
 215 increase of turbulence flow at high flow rates would reduce fouling materials deposition on
 216 the membrane surface. **Explain the applied pressure of 0.5 bar.** As shown in Figure 5, the

217 average water flux was almost constant when using concentrated algae as the feed solution.
 218 The average water flux was around 1.43 LMH at the different studied flow rates for the feed
 219 and draw solutions. This could be due to the constant concentration polarization effect when
 220 using concentrated algae as the feed solution. It can be also seen from Figure 5 that at a flow
 221 rate of 1.5 LPM for both the feed and the draw solution, the average water flux decreased by
 222 25% when using the concentrated microalgae compared to the unfiltered microalgae. When
 223 the flow rate increased to 2.5 LPM for both the feed and the draw solution, the average water
 224 flux decreased by 48% when using the concentrated microalgae compared to the unfiltered
 225 microalgae. At a flow rate of 2.5 LPM (FS) and 0.8 LPM (DS), the average water flux
 226 decreased by 55% when using the concentrated microalgae compared to the unfiltered
 227 microalgae. This is due to the fact that the concentrated microalgae has higher density and
 228 salinity. The higher density caused more fouling of the membrane and the higher salinity
 229 decreased the osmotic pressure gradient (i.e. the driving force) in the FO process.



230

231 *Figure 5. Average water flux of the FO process using microalgae and ultra-filtered microalgae at different flow*
 232 *rates*

233 *3.2.1 Concentration Factor (CF)*

234 The main factor used to assess the harvesting process of microalgae is the
 235 concentration factor (*CF*). The *CF* was calculated using the following equation:

$$CF = \frac{C_f}{C_i} \quad (2)$$

236 Here, C_f and C_i are the final and initial concentrations of the algae biomass, respectively.
 237 Indirectly, the magnitude of CF depends on water flux, membrane area, feed volume and
 238 operating period. CF tends to increase with higher water flux, longer harvest durations, less
 239 feed volume, and larger membrane area. Figure 6a shows the concentration factor for forward
 240 osmosis using unfiltered microalgae as a feed with three different flow rates. It can be seen
 241 from Figure 6a that the CF slightly increased as the flow rate increased. The CF increased by
 242 4.9% as the flow rate increased from 1.5 LPM to 2.5 LPM for both the feed and the draw
 243 solutions. The highest CF was almost 1.23 obtained using a flow rate of 2.5 LPM (FS) & 0.8
 244 LPM (DS), which is attributed to the highest water flux. The total concentration factor of the
 245 hybrid process was also calculated by adding the concentration factors of each process
 246 (Figure 6(b)). In the first pass ultrafiltration process, the CF was 4.76. During the second pass
 247 ultrafiltration process, the concentration factor decreased significantly to 1.12. In the FO
 248 process, the concentration factor was 1.05 using the concentrated microalgae as feed and a
 249 flow rate of 1.5 LPM for both the feed and the draw solutions. The CF increased to 1.11
 250 using a flow rate of 2.5 LPM (FS) & 0.8 LPM (DS). The overall concentration factor of the
 251 dual pass ultrafiltration process was 5.88 and the total concentration factor increased to 7.0
 252 using forward osmosis. In ultrafiltration, the harvesting efficiency decreased significantly by
 253 76.4% when the concentrated algae was the feed solution instead of microalgae. In the FO
 254 process, the harvesting efficiency slightly decreased by 9% when using the concentrated
 255 algae instead of microalgae. Single-pass ultrafiltration has been proven an effective
 256 technology for harvesting microalgae, given that the CF of ultrafiltration was four times
 257 higher than forward osmosis. However, forward osmosis and second pass UF showed similar
 258 efficiency for harvesting microalgae. To evaluate the feasibility of each process, the energy
 259 consumption was calculated.

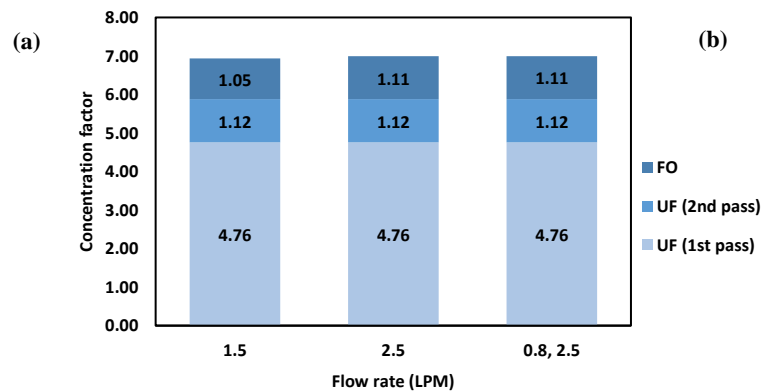
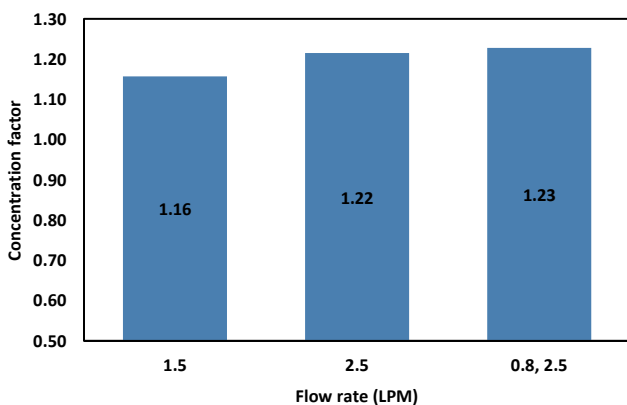


Figure 6. Concentration factor of FO and UF processes at different flow rates using different feed: (a) Forward osmosis using microalgae as feed (b) Microalgae UF 1st pass (concentration = 0.43 g/l), UF 2nd pass (concentration = 15.7 g/l) and FO (concentration = 15.7 g/l).

260 3.2.2 Specific Energy Consumption

261 The specific energy consumption of the UF process and FO process have been
 262 calculated using equation 3 (Shrivastava & Stevens, 2018) and equation 4 (Lambrechts &
 263 Sheldon, 2019) , respectively:

$$E_{s-UF} = \left(\frac{P}{n \times \%R} \right) \quad (3)$$

$$E_{s-FO} = \left(\frac{(Q_s)(\rho)(g)}{36 \cdot (Q_p)} \right) \quad (4)$$

264

265 Here, P is the applied feed pressure (bar), *n* is the pump efficiency, %R is the recovery rate,
 266 *Q_s* is the flowrate of feed and draw solution (m³/h), *Q_p* is the flow rate of the permeate
 267 (m³/h), *ρ* is the density (g/L) and *g* is the gravitational acceleration (m/s²). Figure 7a shows
 268 the specific energy consumption of the FO process using microalgae as the feed and different
 269 flow rates. In the FO process, the specific energy consumption increased from 0.56 Kwh/m³
 270 to 0.72 Kwh/m³ at a flow rate of 1.5 LPM and 2.5 LPM, respectively. As shown in Equation
 271 4, the specific energy is a function of FS and DS flow rate, and the permeate flow rate. The
 272 low specific energy consumption at such a low flow rate was due to the low pumping power
 273 requirements. As the flow rate increased, the specific energy increased because the pumping
 274 power increased and the permeate flow rate remained almost constant. Figure 7b shows the
 275 total energy consumption of the UF-FO hybrid process. The energy consumption of the first
 276 pass ultrafiltration was 4.87 Kwh/m³. In the second pass ultrafiltration, energy consumption
 277 increased by 5.6%. The energy consumption of FO was 0.88 – 1.25 Kwh/m³ depending of the
 278 flow rate of FS and DS. The energy consumption of FO increased as the flow rate increased
 279 due to the higher pumping power requirements. The energy consumption of ultrafiltration
 280 was four times higher than forward osmosis. The total specific energy consumption was
 281 calculated by adding the energy consumption of each process (Figure (b)). The energy
 282 consumption of dual-pass UF was 10 Kwh/m³. However, the energy consumption of the
 283 hybrid UF-FO process was 6.12 Kwh/m³ using a flow rate 2.5 LPM. While achieving the

284 same harvesting efficiency, the energy consumption can be reduced by 46% using a hybrid
285 UF-FO process instead of dual-stage UF.
286

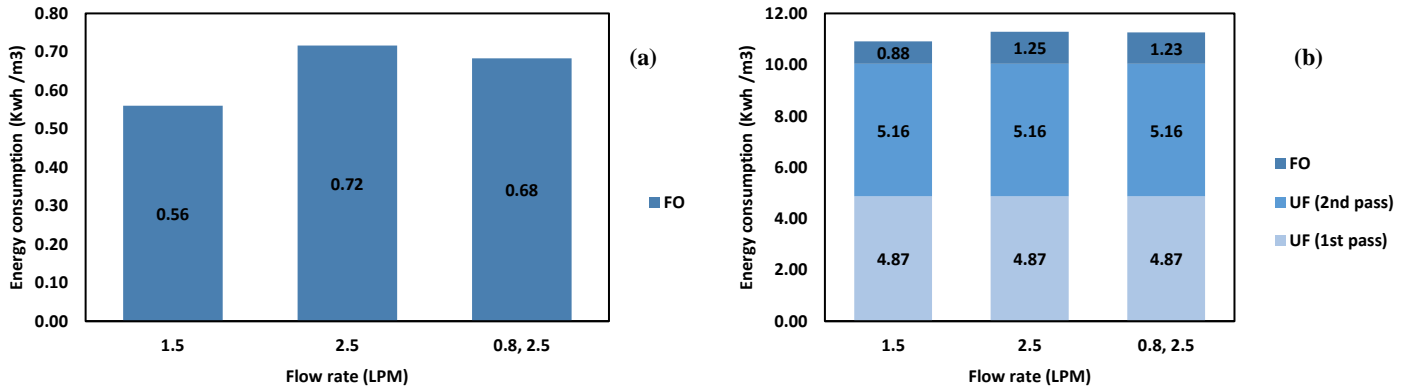


Figure 7. Energy consumption of FO and UF processes at different flow rates using different feed: (a) Microalgae culture (biomass concentration = 0.43 g/l) (b) Microalgae UF 1st pass (biomass concentration = 0.43 g/l), UF 2nd pass (biomass concentration = 15.7 g/l) and FO (biomass concentration = 17.6 g/l)

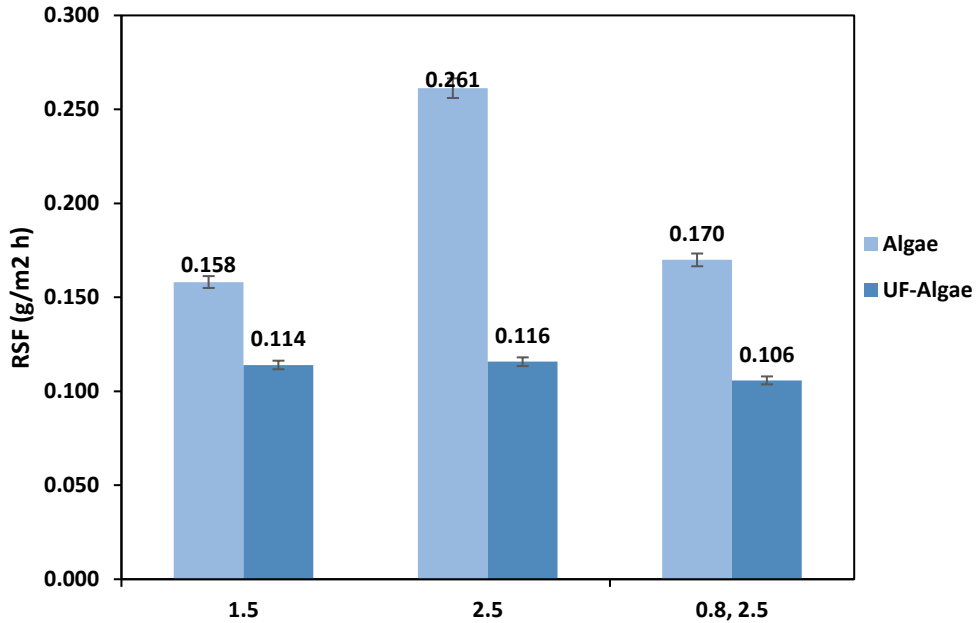
3.2.3 Reverse solute flux

287 Reverse solute flux (J_s) is the back diffusion of the draw solute across the FO
 288 membrane to the feed solution. RSF must be considered in the FO studies because it might
 289 contaminate the feed solution. RSF is calculated using the following equation.

$$J_s = \frac{(C_f V_f) - (C_o V_o)}{t \times A_m} \quad (3)$$

290 Here, C_f and V_f are the concentration and feed volume at the end of the experiment, C_o and
 291 V_o are the concentration and feed volume at the beginning of the experiment. It is obvious from
 292 Figure 7, while using microalgae as a feed solution, the RSF flux increased as the flow rate
 293 increased. The RSF increased from 0.158 g/m².h to 0.261 g/m².h as the flow rates of the draw
 294 and the feed solutions increased from 1.5 LPM to 2.5 LPM, respectively. At flow rate of 2.5
 295 LPM (FS) - 0.8 LPM (DS), the RSF was 0.170 g/m². h. When using ultrafiltered microalgae
 296 as the feed solution, the RSF was around 0.111 g/m². h at all studied flow rates. **Compare**
 297 **with the initial concentration and no contamination of the feed. If the concentration of feed**
 298 **solution is 67 ppt in table 1 (that is 67000 ppm), then the effect of salt back diffusion from**
 299 **DS to FS would be negligible isn't it?**

300



301

302 *Figure 7. Reverse solute flux of FO process using microalgae and ultra-filtered microalgae at different flow*
 303 *rates*

304 4. Conclusions

305 In this study, a hybrid ultrafiltration – forward osmosis process was investigated for
 306 the enhancement of marine microalgae harvesting. Forward osmosis (FO) was used as a post-
 307 harvesting process after ultrafiltration (UF) to attain higher harvesting efficiency and reduce
 308 the energy consumption. The FO process exhibited high resistance to fouling where the water
 309 flux was completely retrieved after washing the membrane with distilled water for 30
 310 minutes. The average water flux obtained using PRO mode was higher than FO mode. In
 311 general, the increase of flow rate increased the average water flux. The increase in flow rate
 312 from 1.5 to 2.5 LPM enhanced the average water flux by 33 %, while using microalgae as
 313 feed. However, the average water flux was unaffected by the flow rate while using pre-
 314 concentrated algae as feed solution, where the average water flux was around 1.40 LMH at
 315 different flow rates.

316 Concentration factors of 4.76 and 1.12 were obtained using first pass ultrafiltration
 317 and second pass ultrafiltration, respectively. Applying the FO process increased the
 318 concentration factor by 1.05 – 1.23, depending on the feed solution and the flow rates of the
 319 feed and the draw solutions. The energy consumption of ultrafiltration was four times higher
 320 than forward osmosis, where the energy consumption of the second pass ultrafiltration and
 321 forward osmosis were 5.26 and 1.25 Kwh/m³, respectively. While the harvesting efficiency of

322 forward osmosis and the second pass ultrafiltration were almost the same with a CF of almost
323 1.12. Forward osmosis was found to be an energy efficient post-harvesting process for
324 microalgae. A maximum total algal harvesting CF of 7.0 was obtained using ultrafiltrated
325 algae as feed, FO mode, 2.5 LPM flowrate and 48 hours operation time. It was found
326 that the energy consumption can be reduced by 46% using a hybrid UF-FO process compared
327 to a dual-stage UF.

328

329 **Acknowledgments**

330 The statements made herein are solely the responsibility of the authors. This research
331 is made possible by NPRP award (NPRP10-0117-170176) from Qatar National Research
332 Fund (QNRF). In addition, the authors wish to thank Qatar Foundation for the financial
333 support provided to one of the co-authors through a graduate sponsorship research award
334 (GSRA6-1-0509-19021).

335

336 **References**

- 337 Achilli, A., Cath, T.Y., Childress, A.E. 2010. Selection of inorganic-based draw solutions for forward
338 osmosis applications. *Journal of Membrane Science*, **364**(1), 233-241.
- 339 Buckwalter, P., Embaye, T., Gormly, S., Trent, J.D. 2013. Dewatering microalgae by forward osmosis.
340 *Desalination*, **312**, 19-22.
- 341 Devia, Y.P., Imai, T., Higuchi, T., Kanno, A., Yamamoto, K., Sekine, M. 2015. Effect of Operating
342 Conditions on Forward Osmosis for Nutrient Rejection Using Magnesium Chloride as a Draw
343 Solution. *International Journal of Environmental, Chemical, Ecological, Geological and*
344 *Geophysical Engineering*. pp. 690-696.
- 345 Honda, R., Rukapan, W., Komura, H., Teraoka, Y., Noguchi, M., Hoek, E.M.V. 2015. Effects of
346 membrane orientation on fouling characteristics of forward osmosis membrane in
347 concentration of microalgae culture. *Bioresource Technology*, **197**, 429-433.
- 348 Lambrechts, R., Sheldon, M.S. 2019. Performance and energy consumption evaluation of a fertiliser
349 drawn forward osmosis (FDFO) system for water recovery from brackish water. *Desalination*,
350 **456**, 64-73.
- 351 Larronde-Larretche, M., Jin, X. 2016. Microalgae (*Scenedesmus obliquus*) dewatering using forward
352 osmosis membrane: Influence of draw solution chemistry. *Algal Research*, **15**, 1-8.
- 353 Larronde-Larretche, M., Jin, X. 2017. Microalgal biomass dewatering using forward osmosis
354 membrane: Influence of microalgae species and carbohydrates composition. *Algal Research*,
355 **23**, 12-19.
- 356 McCutcheon, J.R., Elimelech, M. 2006. Influence of concentrative and dilutive internal concentration
357 polarization on flux behavior in forward osmosis. *Journal of Membrane Science*, **284**(1), 237-
358 247.
- 359 Mi, B., Elimelech, M. 2008. Chemical and physical aspects of organic fouling of forward osmosis
360 membranes. *Journal of Membrane Science*, **320**(1), 292-302.

- 361 Mo, W., Soh, L., Werber, J.R., Elimelech, M., Zimmerman, J.B. 2015. Application of membrane
362 dewatering for algal biofuel. *Algal Research*, **11**, 1-12.
- 363 Munshi, F.M., Church, J., McLean, R., Maier, N., Sadmani, A.H.M.A., Duranceau, S.J., Lee, W.H. 2018.
364 Dewatering algae using an aquaporin-based polyethersulfone forward osmosis membrane.
365 *Separation and Purification Technology*, **204**, 154-161.
- 366 Nguyen, N.C., Chen, S.-S., Yang, H.-Y., Hau, N.T. 2013. Application of forward osmosis on dewatering
367 of high nutrient sludge. *Bioresource Technology*, **132**, 224-229.
- 368 Noffsinger, J., Giustino, F., Louie, S.G., Cohen, M.L. 2009. Origin of superconductivity in boron-doped
369 silicon carbide from first principles. *Physical Review B*, **79**(10), 104511.
- 370 Rickman, M., Pellegrino, J., Davis, R. 2012. Fouling phenomena during membrane filtration of
371 microalgae. *Journal of Membrane Science*, **423-424**, 33-42.
- 372 Shao, P., Darcovich, K., McCracken, T., Ordorica-Garcia, G., Reith, M., O'Leary, S. 2015. Algae-
373 dewatering using rotary drum vacuum filters: Process modeling, simulation and techno-
374 economics. *Chemical Engineering Journal*, **268**, 67-75.
- 375 Shrivastava, A., Stevens, D. 2018. Chapter 2 - Energy Efficiency of Reverse Osmosis. in: *Sustainable*
376 *Desalination Handbook*, (Ed.) V.G. Gude, Butterworth-Heinemann, pp. 25-54.
- 377 Simstich, B., Beimfohr, C., Horn, H. 2012. Lab scale experiments using a submerged MBR under
378 thermophilic aerobic conditions for the treatment of paper mill deinking wastewater.
379 *Bioresource Technology*, **122**, 11-16.
- 380 Singh, R. 2015. Chapter 3. Hybrid Membrane Systems – Applications and Case Studies, pp. 179-281.
- 381 Son, J., Sung, M., Ryu, H., Oh, Y.-K., Han, J.-I. 2017. Microalgae dewatering based on forward osmosis
382 employing proton exchange membrane. *Bioresource Technology*, **244**, 57-62.
- 383 Thabit, M.S., Hawari, A.H., Ammar, M.H., Zaidi, S., Zaragoza, G., Altaee, A. 2019. Evaluation of forward
384 osmosis as a pretreatment process for multi stage flash seawater desalination. *Desalination*,
385 **461**, 22-29.
- 386 Valladares Linares, R., Li, Z., Abu-Ghdaib, M., Wei, C.-H., Amy, G., Vrouwenvelder, J.S. 2013. Water
387 harvesting from municipal wastewater via osmotic gradient: An evaluation of process
388 performance. *Journal of Membrane Science*, **447**, 50-56.
- 389 Volpin, F., Yu, H., Cho, J., Lee, C., Phuntsho, S., Ghaffour, N., Vrouwenvelder, J.S., Shon, H.K. 2019.
390 Human urine as a forward osmosis draw solution for the application of microalgae dewatering.
391 *Journal of Hazardous Materials*, **378**, 120724.
- 392 Wang, Y., Wicaksana, F., Tang, C.Y., Fane, A.G. 2010. Direct Microscopic Observation of Forward
393 Osmosis Membrane Fouling. *Environmental Science & Technology*, **44**(18), 7102-7109.
- 394 Ye, J., Zhou, Q., Zhang, X., Hu, Q. 2018. Microalgal dewatering using a polyamide thin film composite
395 forward osmosis membrane and fouling mitigation. *Algal Research*, **31**, 421-429.
- 396 Zhang, W., Zhang, W., Zhang, X., Amendola, P., Hu, Q., Chen, Y. 2013. Characterization of dissolved
397 organic matters responsible for ultrafiltration membrane fouling in algal harvesting. *Algal*
398 *Research*, **2**(3), 223-229.
- 399 Zhao, S., Zou, L., Mulcahy, D. 2011. Effects of membrane orientation on process performance in
400 forward osmosis applications. *Journal of Membrane Science*, **382**(1), 308-315.