

1 **Microbiome response differs among selected lines of** 2 **Sydney rock oysters to ocean warming and acidification**

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ABSTRACT

20 Oyster microbiomes are integral to healthy function and can be altered by climate change conditions.
21 Genetic variation among oysters is known to influence the response of oysters to climate change and
22 may ameliorate any adverse effects on oyster microbiome, however, this remains unstudied. Nine full-
23 sibling selected breeding lines of the Sydney rock oyster (*Saccostrea glomerata*) were exposed to
24 predicted warming (ambient = 24°C, elevated = 28 °C) and ocean acidification (ambient $p\text{CO}_2$ = 400,
25 elevated $p\text{CO}_2$ =1000 μatm) for four weeks. The haemolymph bacterial microbiome was characterised
26 using 16S rRNA (V3-V4) gene sequencing and varied among oyster lines in the control (ambient $p\text{CO}_2$, 24

27 °C) treatment. Microbiomes were also altered by climate change dependent on oyster lines. Bacterial α -
28 diversity increased in response to elevated $p\text{CO}_2$ in two selected lines, while bacterial β -diversity was
29 significantly altered by combinations of elevated $p\text{CO}_2$ and temperature in four selected lines. Climate
30 change treatments caused shifts in the abundance of multiple Amplicon Sequence Variants (ASVs)
31 driving change in the microbiome of some selected lines. We show that oyster genetic background may
32 influence the Sydney rock oyster haemolymph microbiome under climate change and that future
33 assisted evolution breeding programs to enhance resilience should consider the oyster microbiome.

INTRODUCTION

34 Microbiomes of marine organisms perform a range of beneficial services for their host (McFall-Ngai et
35 al. 2013; Person-Leary et al. 2020; de Vries et al. 2020). For example, microbiomes provide corals with
36 energy and nutrients (Yellowlees et al. 2008), enable fish to alter behaviour (Soares et al. 2019),
37 facilitate bio-luminescence in squid (Aprill 2017) and allow molluscs to resist infection (King et al. 2019a).
38 Microbiomes also allow organisms to exist in hostile environments – for instance, bivalve molluscs living
39 around hydrothermal vents rely on chemosynthetic bacteria to survive (Ponnudurai et al., 2017).
40 Despite their potential importance, little is known about how climate change may influence the
41 microbiome of marine species.

42 Climate change is increasing the frequency and severity of marine heatwaves, increasing average sea-
43 surface temperatures, and reducing the pH and availability of carbonate ions in oceans and estuaries
44 (Rhein et al. 2013; Scanes et al. 2020a). Over the coming decades, these changing ocean conditions will
45 threaten the survival and success of marine organisms, especially calcifying species such as molluscs that
46 produce shells and skeletons made from calcium carbonate (Hoegh-Guldberg et al. 2018). It has been
47 shown that for some calcifying organisms, climate change will decrease calcification, reproductive rate,
48 skew sex ratios, alter adult physiology, slow larval development, and increase larval mortality (Pörtner
49 2008; Ross et al., 2011; Parker et al. 2013; Gattuso et al. 2015; Thomsen et al. 2015; Parker et al. 2018).

50 Adaptation to changing ocean conditions, via the selection of more tolerant genotypes has been
51 identified as a potential source of resilience for some marine species (Parker et al. 2011; 2012; Wright et
52 al. 2014; Foo et al. 2016). For example, in the sea urchin *Centrostephanus rodgersii*, the tolerance of
53 early embryos to ocean warming and acidification was found to differ among genotypes, with some
54 genotypes, showing no effect of warming and only moderate effects of ocean acidification (Foo et al.
55 2016). In the Sydney rock oyster, *Saccostrea glomerata*, populations that were selectively bred for

56 enhanced growth/disease resistance, grew 65% faster and were better able to maintain their
57 extracellular pH under ocean acidification compared to the wild population (Parker et al. 2011; 2012).
58 This effect was also apparent among oyster breeding lines that were full-sibling families. Breeding lines
59 of the oyster *S. glomerata* were found to experience up to 50 % mortality when exposed to an
60 atmospheric heatwave of 50 °C in the laboratory, while there was less than 20 % mortality experienced
61 by other lines (Scanes et al. 2020b). The mechanisms underlying the increased resilience of some
62 genotypes to ocean warming and acidification have been linked to greater maternal provisioning (Foo et
63 al. 2012), changes in the acquisition, production and allocation of energy and changes in energy
64 metabolism (Melzner et al. 2009; Parker et al. 2012; 2015; Thomsen et al. 2015), as well as alterations in
65 genes involved in biomineralization, ion regulation and membrane regulation (Pespini et al. 2013).

66 Resilience to climate change may also be facilitated by microbes and the microbiome. Climate change is
67 likely to alter the microbiome of marine organisms (Scanes et al. 2021), but the level of this effect is
68 potentially dependant on the host's genotype, with this variation potentially driving responses to
69 disease (Wegener et al., 2013; Lokmer and Wegner 2015; Wendling et al. 2017; Qiu et al. 2019, King et
70 al., 2019b; Nguyen et al. 2020) or environmental stress (Aprill 2017; Glasl et al. 2019; Green et al.,
71 2019). Heritable genetic variation influencing the composition of microbiomes can therefore be an
72 integral part of an organism's response to stress. For example, studies on coral have shown that
73 microbiomes differ among coral genotypes, with heat resilient corals possessing a unique microbiome
74 (van Oppen et al. 2018; Glasl et al. 2019; Osman et al. 2020). Microbiomes can also aid in organisms
75 coping with adverse environmental conditions such as toxic pollutants (Aprill 2017). The
76 interdependence of genotype-specific microbiomes and an organism's response to climate change can
77 influence the resilience of marine organisms and play a decisive role in their survival.

78 To understand the influence of oyster genetic variation in the Sydney rock oyster (*S. glomerata*)
79 microbiome's response to climate change, we characterised the oyster haemolymph microbiome in nine
80 *S. glomerata* selected breeding lineages following exposure to ocean warming and acidification. Our
81 rationale for studying the impact of climate change on the host microbiome in oysters is two-fold.
82 Firstly, oysters have proven to be excellent models for studying both the role of microbiome (Lokmer et
83 al. 2015; 2016a; Wendling et al. 2017; King et al. 2019a) and climate change resilience (Parker et al.
84 2012; 2015; Wright et al. 2014; Scanes et al. 2020b). Secondly, oysters are also ubiquitous on shorelines
85 around the globe, including eastern Australia (Scanes et al, 2016), where they provide immense
86 ecological services through building three-dimensional reef structures that shape hydrodynamics and

87 wave attenuation, as well as providing habitat (Coen et al. 2007). In addition, oysters filter the water
88 column and cycle nutrients (Grabowski and Peterson 2007). Oyster aquaculture also provides income
89 and sustenance for communities across the globe, with over 5.5 million tonnes of live oysters with a
90 value of USD 12 billion produced each year (FAO 2018). In eastern Australia the Sydney rock oyster (*S.*
91 *glomerata*) comprises over 90 % of total aquaculture production value worth USD \$35 million annually
92 (NSW DPI 2020) to the state of New South Wales (NSW). The economic cost of climate change on global
93 aquaculture has been estimated to be USD100 billion (Narita et al. 2012). Understanding the
94 mechanisms underlying resilience to climate change, is essential if we are to future-proof these
95 ecologically, economically and culturally important organisms.

96 **MATERIALS AND METHODS**

Study organisms

97 As part of a wider oyster breeding program run by the NSW Department of Primary Industries, genetic
98 lineages of *S. glomerata* have been selectively bred for three-generations at the Port Stephens Fisheries
99 Institute (PSFI; Dove et al. 2020). The purpose of creating selectively bred lines was to build resilience to
100 disease and fast growth to improve aquaculture profitability, however, previous research has shown
101 that different oyster selected lines exhibit dissimilar responses to the effects of climate change (Parker
102 et al. 2011; 2012; 2015; Scanes et al. 2020b). Seventy two oysters were obtained from each of nine
103 distinct selected lines of *S. glomerata*. Each selected line used in this study was created at the PSFI in
104 November 2017 using the culture methods of Dove et al. (2019). Each selected line is created by the pair
105 mating of a male and female oyster from selected lines with known pedigree and traits. When juvenile
106 oysters reached a shell length of 1-2 mm they were transferred to purpose built bags (SEAPA Co.
107 Edwardstown South Australia, 600 x 250 x100 mm) and cultured on intertidal leases in Cromarty Bay,
108 Port Stephens (152° 4'0.69"E, 32°43'19.69"S) where they remained for approximately two years until the
109 beginning of the experiment. The nine selected lines from the breeding program were designated: A, B,
110 C, D, E, F, G, H and I, for this study.

111 **Oyster husbandry, acclimation and experimental exposure**

112 All seawater used in acclimation and experimental exposure was collected from Little Beach, Port
113 Stephens (152°9'30.00"E, 32°42'43.03"S), filtered through canister filters to a nominal 5 µm, and stored
114 onsite in 38,000 L polyethylene tanks as a stock of filtered seawater (FSW).

115 Approximately 72 individual *S. glomerata*, from each of the nine selected lines (A-I) were collected from
116 intertidal leases in Cromarty Bay, Port Stephens (152° 4'0.69"E, 32°43'19.69"S) in September 2019 for
117 experiments, meaning all oysters were 22 months old when experiments began. Once collected, oysters
118 were transported to the laboratory at PSFI and gently cleaned of any fouling organisms before being
119 placed into a single 1500 L fibreglass tank containing aerated FSW. Despite being the same age, the dry
120 tissue weight of oysters differed among selected lines. Mean (\pm SE; g) dry weight of selected lines was; A
121 = 0.7 ± 0.06 ; B = 0.77 ± 0.05 ; C = 0.65 ± 0.03 ; D = 0.58 ± 0.02 ; E = 0.64 ± 0.03 ; F = 0.68 ± 0.04 ; G = $0.7 \pm$
122 0.04 ; H = 0.59 ± 0.03 ; I = 0.49 ± 0.02 g.

123 Oysters were maintained in a single 1500 L fibreglass tank for two weeks to acclimate to laboratory
124 conditions at 24 °C and ambient $p\text{CO}_2$ (400 μatm). During all acclimation and experimental exposure,
125 tanks containing oysters received a full water change every second day. This involved removing all
126 oysters from the tank, gently rinsing them with freshwater to remove solid waste and un-eaten food,
127 placing them into a new tank of FSW pre-equilibrated to their $p\text{CO}_2$ and temperature treatment. During
128 all acclimation and experimental exposure, oysters were fed live algae cultured on site comprising of 25
129 % *Chaetoceros muelleri*, 25 % *Diacronema lutheri*, and 50 % *Tisochrysis lutea* at a rate of 1×10^9
130 cells/oyster/day.

131 Following acclimation, oysters from each selected line were divided among twelve 750 L polyethylene
132 tanks filled with 400 L FSW at a density of 54 oysters per tank, with each selected line represented by six
133 replicate individuals. Experimental treatments are described in detail in Table 1. Treatments consisted of
134 orthogonal combinations of two $p\text{CO}_2$ concentrations (ambient [400 μatm]; elevated [1000 μatm]) and
135 two temperature treatments (24 and 28 °C). Each combination was replicated across three tanks (Table
136 1). Treatments were selected to represent temperatures and $p\text{CO}_2$ concentrations predicted for 2080-
137 2100 by the IPCC (Collins et al. 2013) and reflect measured changes in estuary temperatures reported
138 from south eastern Australia (Scanes et al. 2020a). Oysters remained in experimental treatments for
139 four weeks. Oysters were checked daily for mortality. However, no dead oysters were found in any tanks
140 during the four-week exposure period.

141 The two $p\text{CO}_2$ levels used in this study (400 μatm , 1000 μatm) were based on the multi-model average
142 projection by the IPCC for 2100 (Collins et al. 2013). These $p\text{CO}_2$ levels corresponded to a mean ambient
143 pH_{NBS} of (8.18 ± 0.01) and at elevated CO_2 levels a mean pH_{NBS} of (7.84 ± 0.01). The elevated CO_2 level
144 was maintained using a pH negative feedback system (Aqua Medic, Aqacenta Pty Ltd, Kingsgrove, NSW,
145 Australia; accuracy ± 0.01 pH units). To determine the pH level corresponding to $p\text{CO}_2$ levels, total

146 alkalinity (TA) was quantified at each water change using triplicate Gran-titration (Gran 1952), and a CO₂
147 system calculation program (CO₂ SYS; Lewis et al. 1998), using the dissociation constants of (Mehrbach
148 et al. 1973), and the pH level corresponding with the desired *p*CO₂ level was calculated. Seawater
149 variables including; pH_{NBS}, total alkalinity, and salinity were measured each water change (Table 1), the
150 desired pH values corresponding with *p*CO₂ levels were then recalculated accordingly. Food grade CO₂
151 (BOC Australia) was bubbled directly into independent tanks via a CO₂ reactor to ensure proper mixing
152 and reduce pH. A pH probe connected to a controlling computer was placed within each tank. Each tank
153 set to elevated *p*CO₂ was controlled by its own independent pH controlling system. The pH values of
154 each tank were monitored daily and checked daily against another calibrated pH probe (NBS buffers,
155 WTW 3400i).

156 Once oysters were transferred to experimental tanks, the *p*CO₂ level and temperature were steadily
157 increased in elevated exposure tanks over one week until the experimental treatment level was
158 reached. Oysters were then exposed to their respective treatments for a further four weeks. This length
159 of time has previously been used to elicit a stable response in *S. glomerata* physiology to warming and
160 acidification (Scanes et al., 2017).

161 **Sampling procedure**

162 Following exposure to experimental conditions for four weeks, haemolymph was taken from three
163 replicate oysters from each oyster selected line, from each tank for microbial analysis. This amounted to
164 nine individuals from each selected line, in each treatment. To take haemolymph samples, each oyster
165 was opened using an autoclave sterilised shucking knife, ensuring that the pericardial cavity was not
166 ruptured. Excess fluid was tipped off the tissue surface and 200 – 300 µL of haemolymph was extracted
167 from the pericardial cavity using a new sterile 1 mL needled syringe (Terumo Co.). Samples from three
168 oysters were transferred to three new pre-labelled DNA/RNA free 1 mL tubes (Eppendorf Co.) and
169 immediately frozen at -80 °C where they were stored until DNA extraction.

170 **DNA extraction and sequencing**

171 Bacteria have been identified as the most important component of the microbiome found in oyster
172 haemolymph (Dupont et al., 2020). To characterise the bacterial microbiome of *S. glomerata*
173 haemolymph, 16S rRNA gene sequencing was used. DNA was extracted from 324 oyster haemolymph
174 samples (9 selected lines × 4 treatments × 3 replicate tanks × 3 replicate oysters per tank) using the
175 Qiagen DNeasy Blood and Tissue Kit (Qiagen Australia, Chadstone, VIC), according to the manufacturer's

176 instructions. Twelve samples did not contain enough haemolymph for successful DNA extraction,
177 however, all selected lines in all treatments were still represented. The bacterial microbiome of the
178 oyster haemolymph was characterised with 16S rRNA gene sequencing, using the 341F
179 (CCTACGGGNGGCWGCAG) and 805R (GACTACHVGGGTATCTAATCC) primer pair (Herlemann et al. 2011)
180 targeting the V3-V4 variable regions of the 16S rRNA gene with the following cycling conditions: 95 °C
181 for 3 min, 25 cycles of 95 °C for 30 s, 55 °C for 30 s and 72 °C for 30 s, and a final extension at 72 °C for 5
182 min. Amplicons were sequenced on the Illumina Miseq platform (2x300bp) following the manufacturer's
183 guidelines at the Ramaciotti Centre for Genomics, University of New South Wales. Raw data files in
184 FASTQ format were deposited in NCBI Sequence Read Archive (SRA) under Bioproject number
185 PRJNA663356.

186 **Sequence analysis**

187 Raw demultiplexed data was processed using the Quantitative Insights into Microbial Ecology (QIIME 2
188 version 2019.1.0) pipeline. Briefly, paired-end sequences were imported (qiime tools import), trimmed
189 and denoised using DADA2 (version 2019.1.0), which also removes chimeras (Callahan et al. 2016).
190 Sequences were identified at the single nucleotide threshold (Amplicon Sequence Variants; ASV) and
191 taxonomy was assigned using the classify-sklearn QIIME 2 feature classifier against the Silva v138
192 database (Quast et al. 2012). Sequences were further cleaned by removing ASVs with reads below 0.005
193 % ASVs of the summed relative abundance across all 324 samples. Sequences identified as chloroplasts
194 or mitochondria were also removed. Cleaned data were then rarefied at 6,500 counts per sample. Data
195 from QIIME 2 was then exported to R v.4.0.1 (R Core team) for analysis and statistical testing using the
196 packages, "Phyloseq" (McMurdie and Holmes 2013), "DESeq2" (Anders and Huber 2010) and "Vegan"
197 (Dixon, 2003).

198 **Data analysis**

199 To establish a baseline and compare bacterial microbiomes among oyster selected lines at control
200 conditions, a subset of data was created for only those samples taken from oysters in the ambient $p\text{CO}_2$,
201 24 °C treatment. Samples from this treatment contained 4086 unique ASVs. α -diversity metrics of the
202 ambient $p\text{CO}_2$, 24 °C treatment subset including Simpson's index of diversity and Pielou's evenness were
203 calculated and found not to be normally distributed as determined by Shapiro-Wilk normality tests. α -
204 diversity indices were then analysed among selected lines using the Kruskal- Wallis test. For significant
205 differences detected among selected lines, Kruskal-Wallis multiple comparison tests with Benjamini-

206 Hochberg adjusted p-values were done to determine the source of variation. To analyse β -diversity,
207 Jaccard and Unifrac distances were created from the ambient $p\text{CO}_2$, 24 °C treatment subset and
208 analysed using a single factor PERMANOVA with “Selected line” (A-I) as the fixed and orthogonal factor.

209 α -diversity metrics of the entire data set (all treatments) including Simpson’s index of diversity and
210 Pielou’s evenness, were calculated from rarefied sequence data using the “Phyloseq” package in R
211 software. α -diversity metrics were also calculated from non-rarefied data however the results were not
212 different to the rarefied data and therefore are not presented. Indices were then analysed using a 4-way
213 linear mixed model (LMM) with “Temperature” (24 or 28 °C), “CO₂” (ambient or elevated), and “Selected
214 line” (A-I) as fixed and orthogonal factors, and “Tank” (n=3) as a random factor. ANOVAs (Type II Wald
215 chi-square tests) were performed on models to determine difference among factors using the “lme4”
216 package (Bates et al. 2014), and pairwise comparisons among levels of significant factors or interactions
217 were made using estimated marginal means with Tukey’s methods in with the “emmeans” package.
218 Shapiro’s test was used to confirm a normal distribution and Pearson residual plots were used to check
219 model error.

220 Prior to beta diversity analysis, we removed low abundance ASVs to reduce the dataset complexity (<1
221 % total prevalence [Lokmer et al. 2016a; Callahan et al. 2016]), resulting in the removal of 1143 ASVs to
222 give 4837 ASVs. Unifrac and Jaccard distance matrices were then created from the entire dataset and
223 analysed using a three factor PERMANOVA using the Adonis method with “Temperature” (24 or 28 °C),
224 “CO₂” (ambient or elevated) and “Selected line” (A-I) as fixed and orthogonal factors. These distance
225 matrices were chosen because they are sensitive to low abundance taxa which have been previously
226 shown to drive differences in the *S. glomerata* microbiome (Scanes et al., 2021).

227 To determine which selected lines were significantly affected by CO₂ or Temperature treatments, two
228 factor PERMANOVA using the Adonis methods were run for each selected line on Unifrac and Jaccard
229 distances with “Temperature” (24 or 28 °C) and “CO₂” (ambient or elevated) as fixed and orthogonal
230 factors. This analysis was repeated for each selected line of oysters (Table S2). To determine significant
231 differences in the abundance of ASVs between levels of CO₂ and Temperature treatments, the “DESeq2”
232 package in R v4.0.1 was used. DESeq2 uses univariate Generalised Linear Models with a negative
233 binomial distribution and a Benjamini-Hochberg adjusted P value to compare abundances of ASVs
234 among treatments (Anders and Huber, 2010). The abundances of ASVs between levels of CO₂ and

235 temperature treatments were compared for each oyster selected line. ASVs identified by DESeq2 as
236 significant were searched on the NCBI BLAST database.

237 To explore significant PERMANOVA results, Principal Coordinate Analyses (PCoA) and Canonical Analysis
238 of Principal Coordinates (CAP, using Selected line, CO₂ or Temperature as constraining variables) plots
239 were created for distance matrices to allow for comparisons of constrained and unconstrained
240 ordination (Anderson and Willis 2003). Non-significant PERMANOVA factors were pooled in ordinations
241 to allow for comparison of levels within significant factors. ANOVA was used to assess the significance of
242 CAP constraints. Plots were created using the “Phyloseq” and “Vegan” packages (McMurdie and Holmes
243 2013). Homogeneity of multivariate dispersions were tested using “betadisper” function in the “Vegan”
244 package for all PERMANOVA analyses. Where factors with non-homogenous dispersions were identified,
245 they have been treated with necessary caution (Anderson 2001).

246 To determine the core bacterial microbiome, ASVs were identified that occurred in > 90% of samples
247 from a given grouping, these were calculated using the “AmpVis2” (Albertsen et al. 2015) package in R
248 v4.0.1 software.

249 To explore correlations among bacterial abundances, selected lines and experimental treatments,
250 microbial network analyses were used and figures were created with the “NetCoMi” package (Peschel et
251 al., 2021). Bacterial associations were calculated among selected lines that were “not-affected”
252 (selected lines A, B, C, F, H) and were significantly “affected” (D, E, G, I) by treatments of either elevated
253 pCO₂ or temperature. The SPRING (Yoon et al., 2019) method was used as the association method,
254 correlations were transformed to dissimilarities via the “signed” distance metric and the corresponding
255 (non-negative) similarities are used as edge (connecting line) weights. Node size indicated relative
256 abundance of that taxa, and eigenvector centrality was used for defining hubs (nodes with a centrality
257 value above the empirical 90% quantile). Data were filtered to only include the 85 most abundant taxa
258 to simplify correlation figures.

RESULTS

259 **Bacterial microbiome of different *S. glomerata* selected lines**

260 The microbiome differed among selected oyster lines in the control (ambient pCO₂, 24 °C) treatment
261 (Figure 1). The Simpson’s Index of selected line G was significantly lower than that of selected lines A, E
262 and F (Chi² = 18.49, P = 0.017) and there were significant differences in the bacterial community β –

263 diversity among selected lines (PERMANOVA; Jaccard- $F_{8,72}=1.1$, $P=0.003$; Unifrac- $F_{8,72}=1.24$, $P=0.006$;
264 Figure 2). The core microbes present in control oysters mostly belonged to the bacterial families
265 *Arcobacteraceae* and *Marinifiliaceae*, with five ASVs from *Arcobacteraceae* present in the core microbiota
266 of selected oyster line H (Figure 3).

267 **Bacterial microbiome of *S. glomerata* selected lines exposed to elevated $p\text{CO}_2$ and warming**

268 A total of 5980 bacterial Amplicon Sequencing Variants (ASVs) were identified across the entire dataset
269 of 312 samples. Across all selected lines, the bacterial microbiome of *S. glomerata* haemolymph was
270 predominantly composed of the classes Bacteroidia, Campylobacteria, Gammaproteobacteria and
271 Spirochaetia. The most abundant ASVs in oyster haemolymph were assigned to the genera *Marinifilum*,
272 *Halarcobacter* and *Neptuniibacter* which comprised 75% of the total bacterial microbiome (Figure 1,
273 Figure S2). Four core ASVs were identified, cumulatively comprising 12.6% relative abundance and all
274 members of the *Marinifilaceae*.

275 Overall within and between oyster diversity of hemolymph bacterial microbiomes differed among
276 selected lines when exposed to elevated $p\text{CO}_2$ and temperature ($\text{Chi}^2 = 16.43$, $P = 0.037$; Table S1).
277 Analyses of both α and β - diversity metrics revealed that elevated $p\text{CO}_2$ and temperature affected the
278 bacterial microbiome of *S. glomerata* differently according to oyster selected line. The bacterial α -
279 diversity of two selected lines, D and G, was significantly greater at elevated compared to ambient $p\text{CO}_2$
280 (LMM, $\text{CO}_2 \times$ Selected line; Simpson's Index; Figure 4). Pairwise tests also indicated the bacterial α -
281 diversity of selected lines B and C was significantly different to those of D, H and I, but not A, E, F and G
282 at elevated $p\text{CO}_2$ (Figure 4). There were no significant effects of treatments on Pielou's Index of
283 evenness.

284 The β -diversity of haemolymph bacterial microbiomes were also altered by both elevated $p\text{CO}_2$ and
285 temperature dependent on oyster selected lines. PERMANOVA analysis performed on both Unifrac and
286 Jaccard distances also indicated significant Selected line \times CO_2 and Selected line \times Temperature
287 interactions (Table 2).

288 There were significant effects of both CO_2 and temperature (Figure 5) on the bacterial microbiome of
289 selected lines D (PERMANOVA; CO_2 , $F_{1,32}=0.034$, $P=0.02$; Temperature, $F_{1,31}=0.031$, $P=0.023$) and E (CO_2 ,
290 $F_{1,32}=1.7$, $P=0.029$; Temperature, $F_{1,31}=1.8$, $P=0.014$). In addition, selected line G was significantly
291 affected by CO_2 ($F_{1,32}=0.3$, $P=0.04$) and selected line I, by Temperature ($F_{1,32}=1.4$, $P=0.05$). There were no
292 significant Temperature \times CO_2 interactions in PERMANOVA run for any selected lines.

293 The effects of CO₂ and temperature on selected line D were due to decreases in abundance of nine
294 ASVs, all belonging to the genus *Malaciobacter*. Significant effects (DESeq2 GLM Padj <0.01) of CO₂ and
295 temperature on selected lines E and G were driven by increases and decreases, respectively, in ASVs
296 belonging to the family *Spirochaetaceae* (Table 3). Significant effects of temperature on selected line I
297 were driven by decreases in two ASVs, one belonging to the genus *Mycoplasma* and the other to the
298 genus *Marinifilum*. PCoA plots showed few clear patterns with each selected line displaying overlapping
299 polygons representing treatments (Supplementary Figure S1). Canonical Analysis of Principal
300 Coordinates (CAP) showed that CO₂ and temperature treatments affected each selected line in a unique
301 way (Figure 5). Of the selected lines identified to be significantly affected by experimental treatments,
302 there was clear separation of the treatment groups (Figure 5). CAP constraints were found to be
303 significant (ANOVA P <0.05). Results from the NCBI BLAST database search of significant ASVs as
304 identified by DESeq2 can be found in Table 4.

305 Experimental treatments caused changes in the core haemolymph microbiome of oysters (Figure 3). Six
306 *Arcobacteraceae* ASVs were present in the core of oysters in the control treatment (Ambient CO₂; 24 °C),
307 five ASVs present in oyster selected line H and one present in oyster selected line D. However, the
308 presence of these ASVs was reduced to two in oyster selected line D in the Ambient pCO₂; 28 °C
309 treatment and were entirely lost from the core of oysters in both elevated pCO₂ treatments. In the
310 elevated pCO₂; 28 °C treatment there were increases in core microbes from the family *Marinifilaceae* for
311 oyster selected lines E, F, G, H, I. Elevated pCO₂ increased the number of core members of the family
312 *Nitrincolaceae*, with this ASV present in 7 selected lines in the elevated pCO₂; 24 °C treatment.

313 Network analysis showed different bacterial associations among the selected lines that were and were
314 not-affected by elevated pCO₂. The two networks shared similar properties, there were five node hubs
315 identified in both oyster selected line groups, two of which were shared between the family groups;
316 *Aurantivirga_8e5d* and *Anaerovoracaceae(f)_735d*. There were also similar numbers of clusters in the
317 not-affected and affected lines (11 and 10 respectively). The largest differences in closeness centrality
318 between the selected line groups belonged to the taxa *Anaerovoracaceae(O)_aef4*,
319 *Arcobacteraceae(f)_5970*, *Halarcobacter_faf8* (also from the family *Arcobacteraceae*),
320 *Spirochaetaceae_84ds* and *Spirochaetaceae_ca08*. There were similar overall properties but differing
321 importance of ASVs as nodes and hubs between the two selected line groups as indicated by network
322 analysis (Figure S3).

323 **DISCUSSION**

324 Relationships between microbiomes and the genotypes of their host has so-far been unstudied in
325 response to the ocean warming and acidification caused by climate change. Our study has shown that
326 compositional shifts in the hemolymph bacterial microbiome of oysters was dependent on the oyster
327 selected line, temperature, and ocean acidification. Elevated $p\text{CO}_2$ caused the strongest compositional
328 shifts in microbiomes of some selected lines with negligible changes in others. The selected line
329 dependent nature of microbiome shifts observed in this study may begin to help identify selected lines
330 in which resilient traits are connected to microbiome composition. This may open potential links to the
331 pattern of selected line dependent physiological responses seen in previous studies of climate change
332 and selected lines of *S. glomerata* (Parker et al. 2011; 2012; 2015; Scanes et al. 2020b). The wellbeing of
333 marine organisms is interconnected with their microbiome (McFall-Ngai et al. 2013; Person-Leary et al.
334 2020; de Vries et al. 2020).

335 **Effects of climate change on oyster bacterial microbiomes across selected lines**

336 This study supports other studies that have found that the oyster microbiome is influenced by host
337 genotypes in addition to the environment (Lokmer et al. 2016a; King et al. 2019b; Nguyen et al. 2020;
338 Dupont et al., 2020). We found that in the control treatment, the bacterial microbiome varied according
339 to oyster selected line, a result that has also been found in the Pacific oyster, *Crassostrea gigas* (Lokmer
340 et al. 2016a; King et al., 2019b) and *S. glomerata* (Nguyen et al., 2020), however none of these previous
341 studies investigated the microbiome in a laboratory environment. Recent research on *C. gigas* suggests
342 that the environment is a stronger determinant of bacterial microbiome rather than host genotype
343 (Dupont et al., 2020). In our study, oysters were all reared in the laboratory as larvae, then reared in
344 adjacent cages at the same site for 20 months before being transferred to the laboratory for this study.
345 By keeping oysters together their entire life, we have strong evidence that differences in the
346 microbiome can be driven by genetic variation among oysters (i.e. selected lines).

347 Marine organisms have been found to differ in their resilience to climate change dependent on
348 genotype in urchins (Foo et al. 2016), oysters (Parker et al. 2010; Wright et al. 2014; Scanes et al.
349 2020b), mussels (Thomsen et al. 2017) and seagrass (Reusch et al. 2005). Our study has now shown that
350 the genotype- dependent effects of climate change extend to the microbiome. Recent advances have
351 begun to unravel the reasons for greater resilience of some populations or selected lines of oysters to
352 climate change. Genotypes of oysters that can maintain their haemolymph pH (pH_e) often have greater
353 metabolic rates and more energy efficient filtration rates (Parker et al., 2015; Stapp et al. 2018) which
354 may enable resilience when exposed to climate change.

355 The findings from this study show that climate change-driven shifts in haemolymph bacterial
356 microbiome of oysters is governed in part by the host genotype. Genotypic differences in physiological
357 traits of oysters have been previously identified and suggested to contribute to the resilience of some
358 oyster selected lines (Parker et al. 2015, 2017; Stapp et al. 2018). Such physiological traits may also be
359 present in the selected lines which experienced small perturbations in microbiomes when exposed to
360 elevated $p\text{CO}_2$ or temperature. Our experiment was not equipped to determine whether a microbiome
361 resilient to warming and acidification improves resilience in oysters, or merely a side effect of an oyster
362 with resilient physiology. Further experiments would be needed to determine this.

363 The microbiome of oyster selected lines in this study responded differently to that of wild oysters
364 examined in a recent study (Scanes et al., 2021). Scanes et al., (2021) showed that the microbiome of *S.*
365 *glomerata* was significantly altered by both elevated $p\text{CO}_2$ and temperature at levels comparable to this
366 study. The microbiome of only two of nine selected lines in this study experienced similar effects as the
367 wild cohort of *S. glomerata*. Many of the same taxa were present in the haemolymph microbiome of
368 wild oysters and the selected lines from this study including *Marinifilaceae*, *Arcobacteraceae* and
369 *Spirochaetaceae* (Scanes et al., 2021). Direct comparisons between the microbiome of wild oysters and
370 those of the selected lines used here must be treated with caution because of the differing life histories
371 of wild oysters compared to hatchery reared selected lines and the likelihood that a cohort of wild-
372 caught oysters would contain multiple genotypes from an unknown number of parental combinations
373 (In et al., 2016).

374 This study identified elevated $p\text{CO}_2$ to be a stronger driver of shifts in microbiomes compared to
375 elevated temperature. Elevated $p\text{CO}_2$ increased the α - diversity of bacterial microbiome associated with
376 oysters from selected lines D and G, while some selected lines experienced significant decreases in
377 select ASVs. Reduced diversity of microbiomes is generally considered to be indicative of a shift towards
378 an “unhealthy” state in oysters (Green and Barnes 2010; Lokmer and Wegner 2015; de Lorgeril et al.
379 2019). There were shifts in the relative abundance of taxa dependant on genotypes, for example,
380 elevated $p\text{CO}_2$ and temperature reduced the presence of the bacterial family *Arcobacteraceae*, especially
381 from oyster selected line D and the core of oyster selected line H. *Arcobacteraceae* were also suggested
382 by network analysis as differing between the affected and not-affected selected lines. *Arcobacteraceae*
383 are identified as opportunistic pathogens and have been found in associated with disease causing
384 bacteria such as *Vibrio* (Lokmer and Wegner 2015; de Lorgeril et al. 2019). Several members of the
385 *Arcobacteraceae* have previously been found associated with oysters (Delgado 2020) and are

386 microaerophilic (Vandamme et al. 1991), so it follows that changes in the relative abundance of
387 *Arcobacteraceae* may be due to oysters increasing or decreasing their respiration and metabolic rates,
388 which are responses often reported in response to elevated $p\text{CO}_2$ (Lannig et al. 2010; Parker et al. 2012;
389 Scanes et al. 2017) and the direction of change (increase or decrease) can vary among genotypes (Parker
390 et al. 2012; Wright et al. 2014). Overall, *Arcobacteraceae* comprised a large proportion of the bacterial
391 microbiome of *S. glomerata* haemolymph which is consistent with previous work suggesting the
392 *Arcobacteraceae* are haemolymph-specific symbionts (Lokmer and Wegner 2015).

393 Previous studies have shown that temperature can influence the microbiome in other oyster species
394 such as *C. gigas* (Lokmer and Wegner 2015). However, the microbiome of wild *S. glomerata* was shown
395 to be more strongly influenced by elevated $p\text{CO}_2$ than temperature (Scanes et al., 2021). The strong
396 effects caused by exposure to elevated $p\text{CO}_2$ may be attributable to the internal physiological changes
397 elicited by elevated $p\text{CO}_2$ which are often greater than those caused by elevated temperature at the
398 levels used in this study (Parker et al. 2018). The warming and acidification scenarios used in this study
399 are not unusual for Australian estuaries (Scanes et al. 2020a) indicating that oysters may already be
400 experiencing such alterations to their microbiome in their natural habitat or on oyster farms.

401 There is evidence that the hemolymph microbiome of Pacific oysters, *C. gigas*, can vary according to
402 oyster genotype, however these inter-genotype differences can disappear with changing environmental
403 conditions (Wegener et al., 2013; Lokmer and Wegner 2015). Furthermore, environmental disturbance
404 can remove rare taxa from the microbiome (Lokmer and Wegner 2015). We found changes in the core
405 microbiome of all selected lines, with the strongest changes in the overall haemolymph microbiome
406 experienced by oyster selected lines D, E and G. Each selected line experienced changes in different
407 taxa; DESeq analysis showed that selected line D experienced decreases in the genus *Malaciobacter*, and
408 selected line E had increases in the ASV *Spirochaetaceae_cae5* at both elevated temperature and $p\text{CO}_2$.
409 Conversely, oyster selected line G experienced decreases in *Spirochaetaceae_cae5* when exposed to
410 elevated $p\text{CO}_2$ and temperature. BLAST searches of the NCBI database revealed that at least three ASVs
411 found to be significantly altered by our treatments have been previously identified in oysters.
412 *Spirochaetaceae_cae5* was identified in the tissues of *C. gigas* from Tasmania (Australia) in an
413 experiment that tested the effects of warming on dead oyster tissue (Fernandez-Piquer et al., 2012).
414 Two *Spirochaete* ASVs were identified by network analysis as different between the groups of not-
415 affected and affected selected lines. The prevalence of *Spirochaetes* in our study and that of others (e.g.
416 Fernandez-Piquer et al., 2012) suggests that *Spirochaetes* may be a common feature of the oyster

417 microbiome that respond to environmental change. Two ASVs identified as *Mycoplasma*;
418 *Mycoplasma_e553* and *Mycoplasma_d15b* matched ASVs in the BLAST search from previous studies by
419 Fernandez-Piquer et al., (2012) and Green and Barnes (2010). Green and Barnes (2010) found that
420 *Mycoplasma sp.* were absent in *S. glomerata* collected from the Pimpama River, Queensland (Australia)
421 that were infected with the protozoan *Marteilia sydneyi*. *M. sydneyi* is responsible for QX disease in *S.*
422 *glomerata*, a disease that causes widespread oyster death and significant economic damage during
423 outbreaks. We found *Mycoplasma_e553* to decrease in abundance at elevated $p\text{CO}_2$ indicating that this
424 ASV may be more abundant in healthy oysters. The functional role of these bacterial taxa remains
425 largely unknown and further testing would be required to discern whether changes in their abundance
426 would have downstream consequences for oyster health and survival particularly in the context of
427 climate change.

428 **Conclusions**

429 This study has shown that both elevated $p\text{CO}_2$ and temperature can cause significant shifts in
430 microbiome composition. With four of the nine selected lines showing microbiome shifts, the
431 microbiome's response to climate change is clearly dependent on the genetic background of the
432 respective oyster selected lines. Such shifts have previously been associated with an immune
433 compromised state (de Lorgeril et al. 2019), which may allow opportunistic bacteria to invade and result
434 in death. Microbiome stability across environments might therefore correlate with climate change
435 resilience, and preliminary findings suggest those selected lines with stable microbiomes may also be
436 physiologically resilient to climate change (Parker et al., Unpublished data). Further research on the
437 causal direction between physiology and microbiome composition will be needed to explore this
438 connection.

439 Oysters contribute invaluable ecological services such as providing habitat for fish and invertebrates
440 while filtering the water column (Grabowski et al. 2012). Oyster aquaculture also provides income and
441 sustenance for communities across the globe, with over 5.5 million tonnes of live oysters with a value of
442 USD \$12 billion produced each year, much of which is produced by sustenance farmers in developing
443 nations (FAO 2018). Producing invertebrates sustainably will be vital to meeting the ever-growing global
444 demand for food. We therefore, must identify, understand and build resilience in oysters to maintain
445 their economic, social and cultural value in the face of climate change. The response of oyster
446 microbiomes is dependent on oyster genotypes which may hold the key to enhancing the production
447 and survival of oysters in the face of intensifying climate change by assisted evolution and selective

448 breeding. Such approaches need to consider the microbiome if we are to future-proof marine organisms
449 and habitats to climate change.

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460 CONFLICT OF INTEREST

461 The authors have no conflicts of interest to declare.

462 FIGURE LEGENDS

463 **Figure 1.** Mean relative abundances of *S. glomerata* haemolymph microbial taxa at the Order level for
464 each oyster selected lines A-I as well as temperature (24 or 28 °C) and $p\text{CO}_2$ (ambient or elevated)
465 treatments at the end of four weeks experimental exposure. Each column represents a mean relative
466 abundance for that selected line and treatment combination.

467 **Figure 2.** Ordination plots of microbial microbiome in *S. glomerata* haemolymph using Constrained
468 Analysis of Principal Coordinates (CAP) on Jaccard (Panel A) and Unifrac (Panel B) distances of samples
469 for each oyster selected line (selected lines A-I) in the control (ambient $p\text{CO}_2$; 24 °C) treatment. Polygons
470 encircle oyster selected lines, colours and line type also identify oyster selected lines.

471 **Figure 3.** Presence/absence map of core microbial members that were present in >90% of samples from
472 an oyster selected line (A-I) in temperature (24 or 28 °C) and $p\text{CO}_2$ (Ambient or elevated) treatments.
473 Each row represents a unique ASV, and ASV rows are grouped into taxonomic families. Only ASVs that
474 were in the core of the control (ambient $p\text{CO}_2$; 24 °C) treatment are included to simplify this figure.

475 **Figure 4.** Boxplot (n=54) in the style of Tukey of Simpson's Index calculated from *S. glomerata*
476 haemolymph bacterial microbiome at ambient (white) or elevated (grey) $p\text{CO}_2$ treatments and each
477 oyster selected line (selected lines A-I). The upper and lower "hinges" correspond to the first and third
478 quartiles (the 25th and 75th percentiles) while whiskers extend from the hinge to the highest and lowest
479 value that is within $1.5 \times \text{IQR}$ of the hinge, data beyond the end of the whiskers are outliers and plotted
480 as points. Asterisks below red-outlined boxes indicate significant differences between $p\text{CO}_2$ treatments
481 for those selected lines as determined by Post-Hoc Tukey tests. Results were averaged over the factor
482 "Temperature" because this factor was not significant.

483 **Figure 5.** Ordination plots of microbial microbiome in *S. glomerata* haemolymph using Constrained
484 Analysis of Principal Coordinates (CAP) on Unifrac distances of samples for each oyster selected line.
485 **Panel A** Shows the effects of elevated $p\text{CO}_2$ on the selected lines identified as significantly affected by
486 PERMANOVA (selected lines D, E, G), the X-axis was constrained by CO_2 treatment. **Panel B** shows the
487 effects of elevated temperature on the selected lines identified as significantly affected by PERMANOVA
488 (selected lines D, E, I), the X-axis was constrained by temperature treatment. Polygons encircle
489 temperature and $p\text{CO}_2$ treatments for each oyster selected line, indicated by colours and shapes of
490 points. The percentages in parentheses represent the variability explained by significant axes.

491

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