

1 **The individual and synergistic indexes for assessments of heavy metal**  
2 **contamination in global rivers and risk: A review**

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34 **Abstract**

35           This article provides an overview of heavy metal contamination in rivers and  
36 assessment methods of their contamination and effects. According to literature, rivers  
37 with heavy metal contamination in surface water are mainly found in developing  
38 countries in Asia, Africa, and Latin America and the Caribbean area, while rivers with  
39 heavy metal contamination in sediments are mostly found in Europe. The increase in  
40 heavy metal contamination in rivers has led to the adoption of individual and synergistic  
41 assessment methods. Individual methods are useful in assessing the contamination and  
42 effects for a single heavy metal, while synergistic methods assess the combined  
43 contamination and effects of several heavy metals present in surface water and  
44 sediments. These two approaches have been commonly used together in recent studies  
45 to overcome the limitations of each other and provide a more comprehensive  
46 assessment. The developments, equations, advantages, limitations, and future  
47 perspectives of these methods are discussed in this review. Calculating indexes are  
48 simple, easy-to-implement, and effective methods to provide early alerts for the  
49 environmental changes and the adverse impacts on ecosystems and human health.  
50 However, calculating indexes still have limitations due to the lack of background  
51 concentrations of heavy metals in the study area. Therefore, this issue should be  
52 addressed to overcome the limitations of these methods in the future. This review  
53 provides a useful reference for future studies on heavy metal contamination in global  
54 rivers and the assessment methods for heavy metal contamination and effects.

55

56 **Keywords:** Ecological risk assessment; Heavy metal pollution; Human health risk  
57 assessment; Individual indexes; Synergistic indexes.

## 58 **1. Introduction**

59 Heavy metals released from the various sources eventually end up in the  
60 environment, especially in surface waters and sediments of rivers [1]. Rivers in highly  
61 industrialized regions, especially with several metal-related industries, often have  
62 higher concentrations of heavy metals in the water and sediments than in other regions  
63 [2..]. Since heavy metals are generally stable and non-degradable, they accumulate  
64 more in sediments of rivers over time, causing serious pollution [3]. Therefore, the  
65 assessment of heavy metal contamination in rivers is essential.

66 Various heavy metals are released into the environment during the  
67 manufacturing process, such as As, Cd, Cr, Cu, Pb, Ni, Zn, and Hg [2..]. Several heavy  
68 metals in trace amounts are necessary for the growth of organisms, such as Cu, Fe, and  
69 Zn [4]. However, most heavy metals are toxic and can cause an imbalance in the  
70 ecosystem and even the death of organisms [4, 5]. For human health, heavy metal  
71 contamination may pose both carcinogenic and non-carcinogenic risks when they enter  
72 the human body. Some common effects of heavy metal contamination on human health  
73 are memory loss, mental confusion, allergies, fatigue, high blood pressure, skin rashes,  
74 and joint stiffness [6]. Hence, the assessment of heavy metal contaminations and their  
75 effects on the ecosystem and human health is attracting much attention from scientists  
76 worldwide.

77 Various methods have been developed to assess heavy metal contamination  
78 levels in rivers and their effects on the ecosystem and human health [7, 8, 9]. These  
79 methods are based on heavy metal concentrations in the environment, regulatory  
80 limit, background concentrations, and toxicity values. The indexes are usually divided  
81 into two groups: individual and synergistic indexes [10..]. The individual indexes  
82 were developed to assess the contamination level and the impact of a single heavy

83 metal on the ecosystem and humans. These indexes include the geo-accumulation  
84 index ( $I_{geo}$ ), contamination factor ( $CF$ ), enrichment factor ( $EF$ ), potential  
85 contamination index ( $PCI$ ), hazard quotient ( $HQ$ ), lifetime cancer risk index ( $CR$ ), and  
86 modified hazard quotient ( $mHQ$ ) [9, 7, 11, 12, 13, 14••, 15, 16]. The synergistic  
87 indexes were developed to evaluate the total contamination of several heavy metals in  
88 the environment and their cumulative risks on humans and the ecosystem. Some  
89 common synergistic indexes are the modified degree of contamination ( $mCd$ ),  
90 pollution load index ( $PLI$ ), metal index ( $MI$ ), heavy metal pollution index ( $HPI$ ),  
91 degree of contamination ( $DC$ ), contamination severity index ( $CSI$ ), potential  
92 ecological risk index ( $PERI$ ), ecological contamination index ( $ECI$ ), cumulative cancer  
93 risk ( $CCR$ ), and total hazard index ( $HI$ ) [17, 18, 8, 19, 7, 15, 16]. Nowadays,  
94 combining individual and synergistic indexes to provide detailed and comprehensive  
95 evaluation is a popular direction in heavy metal contamination assessment [10••, 20•,  
96 21].

97       The use of indexes to assess heavy metal contamination has been applied to the  
98 most contaminated rivers in the world. For instance, indexes have been used to assess  
99 heavy metal contamination of the Tamirabarani River in India, the Korotoa River in  
100 Bangladesh, the Yangtze River in China, the Tigris River in Turkey, and the To Lich  
101 River in Viet Nam [22, 23, 21, 24, 25]. Several new indexes, such as  $mHQ$ ,  $ECI$ ,  $PLI$ ,  
102 and  $mCd$ , were also developed and applied in recent studies [14••, 19]. However, there  
103 is still a lack of a comprehensive summary of available assessment indexes in a review.  
104 The assessments on the advantages and limitations of index methods are limited.  
105 Therefore, this study presents an overview of available indexes commonly used to  
106 assess heavy metal contaminations and their effects on the ecosystem and humans. The  
107 advantages, limitations, and future perspectives of index methods are also discussed in

108 this review. The characteristics of heavy metals and their contamination in worldwide  
109 rivers are summarized as background information. This review provides an overall  
110 picture of the indexes used in heavy metal contamination assessment. It can be used as a  
111 useful reference for scientists in this field.

## 112 **2. Data collection**

113 The bibliographic databases of Web of Science, Google Scholar, and PubMed  
114 were collected for the article review. Web of Science is an online subscription that  
115 provides a comprehensive citation search and provides the full text of scientific articles  
116 worldwide. Google Scholar is a freely accessible web search engine that includes  
117 academic journals, books, conference papers, technical reports, and other scholarly  
118 peer-reviewed literature. PubMed is a free search developed by the National Center for  
119 Biotechnology Information (NCBI) and provides the full text of scientific articles and  
120 online books worldwide. The research articles and other literature included in this study  
121 were collected using different keywords. In total, 35 search terms (Table S1) were  
122 employed with different combinations. With no restriction of time, the literature review  
123 ended in December 2020 with approximately 100 documents selected after screening,  
124 including published journal articles, conference papers, books, and book chapters, as  
125 shown in the reference list. The information gathered from these previous studies was  
126 then incorporated into this study's tables, figures, and text. On that basis, the  
127 evaluations, discussions, conclusions, and future research perspectives for this area are  
128 provided.

## 129 **3. Heavy metal contamination in surface water and sediments**

### 130 **3.1. Sources and transport of heavy metals**

131 Heavy metals in surface water and sediments of rivers may originate from  
132 natural sources (rock weathering and volcanic eruptions) and human-made activities

133 (mining, metal processing, and agricultural activities) [26]. Industrial activities have  
134 become the main sources of heavy metal contamination in rivers [10••]. Different  
135 industrial activities can release different kinds of heavy metals into the environment  
136 (Table S2). Mining areas often have a high level of heavy metal contamination (As,  
137 Cd, Hg, and Pb) due to their release during the mining process [27]. Metallurgy,  
138 electroplating, and other metal-surface-processing can release various heavy metals  
139 into the environment such as As, Cd, Cr, Cu, Pb, Ni, Zn, and Hg during the  
140 manufacturing process [2••].

141         The transport of heavy metals in the environment is shown in Fig. S1. Rivers  
142 are often the final receptors for pollutants when released into the environment [1].  
143 Heavy metals from natural and human-made activities enter rivers through waste  
144 discharge, leaching, and runoff [28•]. When heavy metals dissolve in water, they  
145 often carry a positive charge [5]. Heavy metals can combine with other anions to form  
146 heavy metal compounds in surface water. Heavy metals and their compounds in  
147 surface water can accumulate in sediments, tissues of aquatic organisms and enter the  
148 human body through the food chains [3, 29]. Subsequently, they may harm the  
149 ecosystem and human health.

### 150 **3.2. Effects of heavy metal contamination on the ecosystem and human health**

151         Heavy metal contamination may adversely affect the ecosystem and biodiversity  
152 of the receiving environment [29, 3]. Heavy metals in aquatic systems are often  
153 suspended or insoluble before accumulating in sediments and organisms [3]. This  
154 accumulation is irreversible and takes place over a long period. The accumulation of  
155 heavy metals in aquatic organisms' tissues may cause the death of organisms because  
156 of their toxicity, leading to the imbalance and the destruction of aquatic ecosystems  
157 [5].

158 Fishes, one of the main aquatic organisms in the food chain, usually accumulate  
159 large amounts of heavy metals in their tissues [5]. Therefore, they are commonly used  
160 in estimating ecological and human health risks [30]. Heavy metals, such as As, Cr,  
161 Cd, Zn, Pb, Hg, Cu, and Ni, are common toxic contaminants for fishes. Previous  
162 studies showed that heavy metals could alter biochemical and physiological functions  
163 in tissue and blood and cause cancer in some fish species [3, 5]. References for acute  
164 (LC<sub>50</sub>) and chronic (NOEC or LOEC) toxicity of heavy metals to certain aquatic  
165 organisms, especially fish, are listed in Table S3.

166 Although heavy metals make an important and essential contribution to  
167 metabolism in the human body, they become toxic when they cannot be metabolized  
168 and accumulate in soft tissues [4]. Chronic toxicities of heavy metals to human health  
169 have been studied in previous researches (Table S3). The exposure pathways of heavy  
170 metals to humans are very diverse, including food and water consumption, dermal  
171 contact, and inhalation of polluted air. In the human body, most heavy metals are  
172 transported through the bloodstream and distributed in the tissues [5]. Because of their  
173 high degree of toxicity, some heavy metals, such as As, Cd, Cr, Pb, and Hg, are  
174 prioritized. These heavy metals are considered systemic toxicants that can damage  
175 many organs with less exposure than other metals. International Agency for Research  
176 on Cancer and the U.S. Environmental Protection Agency also classify them as  
177 human carcinogens [4]. Some popular effects of heavy metal contamination on human  
178 health are memory loss, mental confusion, allergies, fatigue, high blood pressure, skin  
179 rashes, and joint stiffness [6].

### 180 **3.3. An overview of heavy metal contamination in global rivers**

181 In recent decades, heavy metal contamination has become a global  
182 environmental problem [31••]. Rivers in urban areas, which receive wastewater,

183 reflect the extent of heavy metal contamination in the environment [1]. Worldwide  
184 heavy metal contamination in surface water and sediments of rivers are summarized  
185 in Table 1a and b, respectively. Generally, the world's major river systems are  
186 severely contaminated with heavy metals, especially in Asia, Africa, Europe, and  
187 Latin America and the Caribbean area [31••]. Typical examples have shown that  
188 many rivers in the world are “dead rivers”.

189 Heavy metal contamination in the surface water is concentrated mainly in  
190 developing countries in Asia and Africa, where there are high industrial activities and  
191 a lack of contamination control measures. As shown in Table 1a, the concentration  
192 ranges of heavy metals in the surface water in Asia and Africa rivers were: As:  
193 0.00097 - 0.05535 mg L<sup>-1</sup>; Cd: 0.00008 - 0.1 mg L<sup>-1</sup>; Cr: 0.0013 - 1.11; Cu: 0.000061  
194 - 1.05 mg L<sup>-1</sup>; Pb: 0.00058 - 7.5 mg L<sup>-1</sup>; Ni: 0.00495 - 0.21 mg L<sup>-1</sup>; Zn: 0.00038 -  
195 21.71 mg L<sup>-1</sup>; and Hg: 0.0002 - 0.0004 mg L<sup>-1</sup>. Rivers in Asia with heavy metal  
196 contamination in surface water were concentrated mainly in China, Bangladesh, and  
197 India, e.g., the Yangtze and Pearl River in China [32, 33], the Buriganga and Bangshi  
198 River in Bangladesh [34, 35], the Gomti and Kali River in India [36, 37]. Meanwhile,  
199 the Challawa River in Nigeria and the Nairobi River in Kenya are heavy metal  
200 contaminated rivers in Africa [38, 39]. Most of these rivers have received wastewater  
201 from industrial activities, the most contributing source for heavy metal contamination.  
202 This has led to higher heavy metal concentrations in the surface water of Asia and  
203 Africa rivers than in other regions. For instance, the highest heavy metal  
204 concentrations in the surface water of rivers in Asia were Cd, Cr, Cu, Pb, Ni, and Zn  
205 with concentrations 3.0-17.3; 3.5 - 17.7; 1.0 - 4.4; 1.1 - 4.8; 3.1 - 23.2; and 1.1 - 3.5  
206 times higher than that in rivers in Europe, South, and North America, respectively  
207 [31••]. The heavy metal concentrations in the surface water of rivers in Asia and



208 Africa also exceeded their respective standards. The concentrations were 0.02 - 12.02  
209 times higher than the permissible values of the freshwater toxicity reference values,  
210 the US EPA human health ambient water quality criteria, and the WHO drinking  
211 water quality guidelines [40, 41].

212 Besides, heavy metal pollution in surface water of some typical rivers in Latin  
213 America and the Caribbean area also reached alarming levels higher than the  
214 respective standards [42, 43, 44]. For instance, As, Cd, Cr, Cu, and Ni concentrations  
215 in the surface water of the San Pedro River in Mexico were 1-22 times higher than the  
216 freshwater toxicity reference values and 0.1-16 times higher than the WHO drinking  
217 water quality guidelines [45, 41]. Concentrations of heavy metals in the surface water  
218 of these rivers were also higher than in North and South America, as shown in Table 1  
219 a. The heavy metal pollution in rivers in Latin America and the Caribbean area mainly  
220 comes from mining activities [42, 43, 44].

221 Meanwhile, rivers with heavy metal contamination in sediments were mainly  
222 concentrated in developed areas such as European countries [31••], as shown in Table  
223 1b. This region includes countries that have gone through a period of vigorous  
224 industrial development. This has led to the high accumulation of heavy metals in the  
225 sediments of rivers in Europe. These rivers, such as the Odra River in Poland, the  
226 Tinto River in Spain, the Tigris River in Turkey, the Tees River in the UK, and the  
227 Danube River in central and eastern Europe, have alarmingly high heavy metal  
228 concentrations in their sediments. The dominant heavy metals in the sediments were  
229 As (95.33 mg kg<sup>-1</sup> dry wt. in the Odra River) [46], Cd and Cr (32.9 and 556.5 mg kg<sup>-1</sup>  
230 dry wt., respectively in the Danube River) [47], Cu, Ni, and Zn (2860.25; 534.58; and  
231 5,280 mg kg<sup>-1</sup> dry wt., respectively in the Tigris River) [24], and Pb (13,400 mg kg<sup>-1</sup>  
232 dry wt. in the Tinto River) [48]. The As, Cd, Cr, Cu, Pb, Ni, and Zn concentrations in

233 sediments of rivers in Europe were 1.7 – 24.9; 1.1 – 4.0; 1.2 – 4.1; 3.9 – 28.9; 1.5 –  
234 26.5; 14.5 – 48.5; and 2.7 – 29.8 times higher than that in rivers in Asia and Africa,  
235 respectively. Their concentrations were 0.52 – 46.85 times higher than the National  
236 Oceanic and Atmospheric Administration (NOAA)'s effects range low (ERL) and  
237 effects range median (ERM), and the freshwater toxicity reference value (TRV) of  
238 USEPA [45, 49].

239 Besides, some rivers in Latin America and the Caribbean area also have severe  
240 heavy metal contamination in sediment, as shown in Table 1b. For instance, the  
241 Rimac River in Peru had high concentrations of As (1,543 mg kg<sup>-1</sup> dry wt.), Cd (31  
242 mg kg<sup>-1</sup> dry wt.), Cr (71 mg kg<sup>-1</sup> dry wt.), Cu (796 mg kg<sup>-1</sup> dry wt.), Pb (2,281 mg kg<sup>-1</sup>  
243 dry wt.), and Zn (8,076 mg kg<sup>-1</sup> dry wt.) [27]. Long-time mining activities in this area  
244 are the cause of high heavy metal concentrations in the sediments.

#### 245 **4. Heavy metal contamination assessment**

246 The increase of heavy metal pollution in rivers and their toxicity has led to  
247 concerns about quantifying their contamination and effects. The assessment methods  
248 for heavy metal contamination and their effects have been developed and used by  
249 scientists recently [10••, 14••]. In general, contamination assessments are normally  
250 conducted by the calculation of contamination indexes. These calculations are  
251 performed based on the heavy metal concentration in surface water and sediments and  
252 their background values in the environment, or their permissible values according to  
253 specific standards. Besides, the levels of the adverse impact on the ecosystem are also  
254 indicated through the calculation of these indexes. The assessment indexes can be  
255 divided into two groups, individual indexes and synergistic indexes (Table S4).

##### 256 **4.1. Individual indexes**

257 *Geo-accumulation index ( $I_{geo}$ )*

258 The  $I_{geo}$  was first introduced by Muller et al. [9] and has been widely used in the  
259 assessment of heavy metal abundance in sediments [22, 50]. To classify the  
260 contamination levels of heavy metals in sediments,  $I_{geo}$  is calculated based on the Eq.  
261 (1):

$$262 \quad I_{geo} = \log_2 \left( \frac{C_{s_i}}{1.5C_{b_i}} \right) \quad (1)$$

263 where  $C_{s_i}$  is the measured concentration of heavy metal “ $i$ ” in sediments,  $C_{b_i}$  is  
264 the background value of heavy metal “ $i$ ”, and the coefficient 1.5 is the correction  
265 coefficient due to lithogenic effects of background values. The average shale values  
266 [51] and heavy metal concentration in the continental crust [52, 53] are normally used  
267 as background values. The heavy metal contamination degree in sediments can be  
268 divided into seven levels based on  $I_{geo}$  values, as shown in Table 2.

### 269 ***Contamination factor (CF)***

270 The contamination factor,  $CF$ , was proposed by Hakanson [7] to indicate the  
271 contamination level of individual heavy metals in sediment. Recent studies have also  
272 used  $CF$  to evaluate the heavy metal contamination level in surface water.  $CF$  is  
273 calculated by the ratio of heavy metal concentration ( $C_{s_i}$ ) and its background  
274 concentration ( $C_{b_i}$ ), as shown in the Eq. (2).

$$275 \quad CF_i = \frac{C_{s_i}}{C_{b_i}} \quad (2)$$

276 The background concentrations are usually referred from Turekian, Wedepohl  
277 [51], Taylor [52], and Rudnick, Gao [53]. Besides, the local standards have also been  
278 used as background concentrations in recent studies [28•]. Contamination levels of  
279 heavy metals in the environments classified based on the  $CF$  values are shown in  
280 Table 2. Heavy metals are considered contaminated when  $C_{s_i}$  is higher than  $C_{b_i}$  and  
281 uncontaminated when  $C_{s_i}$  is lower than  $C_{b_i}$ .

282            ***Enrichment factor (EF)***

283            The enrichment factor, *EF*, was proposed by Sinex, Helz [11] and is used to  
284 evaluate the human-made effects on heavy metal contamination in sediment. *EF* is  
285 also a good index to differentiate between heavy metal sources from natural and  
286 human-made activities. This index is calculated based on the normalization of heavy  
287 metal concentrations in sediments to background concentrations, as shown in Eq. (3).  
288 Many heavy metals are used for normalization, such as Fe, Al, Mn, Sc, Li, or Zr, due  
289 to their high natural abundance and lower probability of being enriched by  
290 anthropogenic activities [11]. The classification of *EF* is presented in Table 2.

291            
$$EF_i = \frac{[C_{s_i}/C_n]_s}{[C_{s_i}/C_n]_b} \quad (3)$$

292            where  $C_{s_i}$  is the concentration of heavy metal *i*;  $C_n$  is the background  
293 concentration of the normalizing metal; *s* is the study sample; *b* is the background.

294            ***Potential contamination index (PCI)***

295            *PCI* is a new index developed by Dauvalter, Rognerud [12] based on the  
296 method of Hakanson [7] and is used to estimate the potential contamination of heavy  
297 metals in sediments. *PCI* has three contamination levels: low, moderate, severe, or  
298 very severe contamination (Table 2). *PCI* is calculated by the ratio of the maximum  
299 concentration of heavy metal "i" ( $C_{s_{i-max}}$ ) and its background concentration ( $C_{b_i}$ ), as  
300 shown in Eq. (4).

301            
$$PCI_i = \frac{C_{s_{i-max}}}{C_{b_i}} \quad (4)$$

302            ***Hazard quotient (HQ)***

303            *HQ* has been used to estimate the potential hazards on the ecosystem of individual  
304 heavy metals in water and sediments [13]. *HQ* is estimated based on the ratio of the

305 heavy metal concentration ( $C_{s_i}$ ) in the environment and the environmental quality  
306 criteria ( $EQC$ ) as shown in Eq. (5).

$$307 \quad HQ_i = \frac{C_{s_i}}{EQC} \quad (5)$$

308 Normally, aquatic life criteria and sediment quality guideline (SQG) was used for  
309 water and sediment, respectively [13, 28•]. The ecosystem hazard levels based on  $HQ$   
310 are classified from "no adverse effects" to "high hazard" (Table 2).

### 311 ***Modified hazard quotient (mHQ)***

312 A newly developed index,  $mHQ$ , has been proposed to evaluate the adverse  
313 effects of individual heavy metals in sediments on the ecosystem [14••]. This new  
314 method allows the assessment of adverse effects by comparing the concentrations of  
315 individual metals in sediments with their probable effect level (PEL), threshold effect  
316 level (TEL), and severe effect level (SEL) values. The PEL, TEL, and SEL are used  
317 as reference values for contamination.

$$318 \quad mHQ = \left[ C_{s_i} \left( \frac{1}{TEL_i} + \frac{1}{PEL_i} + \frac{1}{SEL_i} \right) \right]^{1/2} \quad (6)$$

319 The proposed classification of  $mHQ$  for the individual heavy metal is presented  
320 in Table 2.

321 In short, the individual indexes were developed and widely applied early. These  
322 methods are uncomplicated and easy to calculate to provide contamination levels for  
323 each heavy metal. Individual indexes have been widely used to assess heavy metal  
324 contamination in typically contaminated rivers. For instance,  $CF$  and  $I_{geo}$  have been  
325 applied in a lot of previous studies to assess the heavy metal contamination level of  
326 different contaminated rivers in the world, for instance, Tamirabarani River (India),  
327 Korotoa River (Bangladesh), Yangtze River (China), Tigris River (Turkey), and To  
328 Lich River (Viet Nam) [22, 23, 21, 24, 25]. However, the individual indexes can only

329 be applied to a single element and thus may not be sufficient in assessing contamination  
330 since heavy metals are more likely to have synergistic effects on the environment  
331 [10..]. The limitations of the individual indexes have led to the development of  
332 synergistic indexes that have been widely applied to assess water and sediment quality.

#### 333 **4.2. Synergistic indexes**

##### 334 ***Degree of contamination (DC)***

335 *DC* was proposed by Hakanson [7] to estimate the levels of synergistic  
336 contamination of all heavy metals in the environments. *DC* is the sum of the  
337 contamination factors (*CF*) of “*n*” heavy metals present in the contamination area.

$$338 \quad DC = \sum_{i=1}^n CF_i \quad (7)$$

339 The synergistic contamination of heavy metals in water and sediments is  
340 classified into four levels: low to a very high degree of contamination, as shown in  
341 Table 2.

##### 342 ***Modified degree of contamination (mCd)***

343 The modified degree of contamination (*mCd*) index is based on *CF* and  
344 indicates synergistic contamination of heavy metals. The *mCd* index was introduced  
345 by Abraham [16] and calculated using Eq. (8):

$$346 \quad mCd = \frac{1}{n} \sum_{i=1}^n CF_i \quad (8)$$

347 Based on *mCd*, heavy metal contamination of a study site can be classified into  
348 seven synergistic contamination levels (Table 2).

##### 349 ***Pollution load index (PLI)***

350 The *PLI* is calculated as the “*n*” square root of the multiplication of the *CF*  
351 values for “*n*” heavy metals in a specific site [17].

$$352 \quad PLI = (CF_1 \times CF_2 \times CF_3 \times \dots \times CF_n)^{1/n} \quad (9)$$

353 Like *DC* and *mCd*, *PLI* can estimate the synergistic effects of all metals in  
354 surface water or sediments of rivers. The study site is assessed to have heavy metal  
355 pollutant load when the *PLI* value is higher than 1. Contrariwise, when the *PLI* value  
356 is less than 1, there is no heavy metal pollution load at the study site (Table 2).

### 357 ***Nemerow pollution index ( $PI_{Nemerow}$ )***

358 Another approach to determine the synergistic contamination level of heavy  
359 metals at a study area is to use the Nemerow pollution index ( $PI_{Nemerow}$ ) [54]. Similar  
360 to *mCd*, this index is also calculated based on the mean value of the contamination  
361 factors (*CF*), as shown in Eq. (10). However, a weighted average is given based on  
362 the maximum contamination factor ( $CF_{max}$ ) of a single metal to emphasize the effect  
363 of that single metal on the synergistic contamination degree. The  $PI_{Nemerow}$  index is a  
364 weighted multi-factor index for environmental quality assessment. It can help to  
365 highlight the main heavy metal in the study area.

$$366 \quad PI_{Nemerow} = \sqrt{\frac{CF_{average}^2 + CF_{max}^2}{2}} \quad (10)$$

367 where  $CF_{average}$  is the mean value of the contamination factors of heavy metals;  
368  $CF_{max}$  is the highest value of the contamination factors of heavy metals in the study  
369 area.

370 Classification of pollution levels based on  $PN_{Nemerow}$  is shown in Table 2. The  
371 assessment thresholds of this index are quite low compared to other indexes, which  
372 may lead to inaccurate pollution assessment results [55]. Besides, the complex  
373 behavior of sediments is not considered when the pollution level is calculated based  
374 on *CF*. Therefore, an improved method, the modified pollution index (*MPI*), is  
375 proposed based on enrichment factor (*EF*).

### 376 ***Modified pollution index (MPI)***

377 Proposed by Brady et al. [55], the *MPI* is calculated based on *EF* and enables  
378 the sediment quality assessment to take into account the complex behavior of heavy  
379 metals (Eq. (11)).

$$380 \quad MPI = \sqrt{\frac{EF_{average}^2 + EF_{max}^2}{2}} \quad (11)$$

381 Six pollution levels are classified based on *MPI* (Table 2). The thresholds of  
382 sediments quality assessment have been adjusted to provide more accurate pollution  
383 levels.

#### 384 ***Metal index (MI)***

385 *MI* proposed by Caeiro et al. [18], is often used to estimate the synergistic  
386 contamination of heavy metals in river water, canal water, drinking water, and  
387 sediment using Eq. (12).

$$388 \quad MI = \sum_{i=1}^n \frac{C_{s_i}}{UAC_i} \quad (12)$$

389 where  $C_{s_i}$  is the measured concentration of heavy metal “*i*” in sediment;  $UAC_i$  is  
390 the upper allowable concentration of heavy metal “*i*” in water and sediment quality  
391 guidelines. *MI* is categorized into six classes, as shown in Table 2.

#### 392 ***Heavy metal pollution index (HPI)***

393 *HPI* is an index commonly used to evaluate the aggregate effect of heavy metals  
394 on the overall water quality. The arithmetic quality average method was used to  
395 develop *HPI* by Mohan et al. [8]. There are two steps to developing this index (1)  
396 constituting a rating scale for each selected parameter to give the weightage and (2)  
397 choosing the contaminant parameter on which the index is based. The unit weightage  
398 for each heavy metal is given based on its relative influence on water quality for the  
399 consumer, resulting from the creation of values inversely proportional to the standard



400 value of the corresponding metal [8, 56•]. *HPI* is calculated by the following Eq. (13)  
 401 & (14):

$$402 \quad HPI = \frac{\sum_{i=1}^n W_i Q_i}{\sum_{i=1}^n W_i} \quad (13)$$

$$403 \quad Q_i = \sum_{i=1}^n \frac{|C_{S_i} - I_i|}{S_i - I_i} \times 100 \quad (14)$$

404 where “*n*” is the number of heavy metals,  $C_{S_i}$  is the concentration,  $I_i$  is the  
 405 expected limit,  $S_i$  is the permissible limit,  $Q_i$  is the sub-index, and  $W_i$  is the unit  
 406 weightage ( $W_i = 1/S_i$ ) of heavy metal “*i*” in surface water.

#### 407 ***Contamination severity index (CSI)***

408 The contamination severity index (*CSI*) is a new index for assessing heavy  
 409 metal contaminated-sediments developed by Pejman et al. [19]. It is based on the  
 410 effects range low (ERL) and effects range median (ERM) values [49]. The proposed  
 411 equation of *CSI* is shown in Eq. (15) & (16).

$$412 \quad CSI = \sum_{i=1}^n W_t \left[ \left( \frac{C_{S_i}}{ERL_i} \right)^{\frac{1}{2}} + \left( \frac{C_{S_i}}{ERM_i} \right)^2 \right] \quad (15)$$

$$413 \quad W_t = \frac{L_{f_i} \times E_v}{\sum_{i=1}^n (L_{f_i} \times E_v)} \quad (16)$$

414 where  $W_t$  is the weighted value for heavy metals,  $L_{f_i}$  is the factor loading  
 415 associated with heavy metal “*i*”,  $E_v$  is the eigenvalue. There are nine contamination  
 416 levels classified based on *CSI*, as shown in Table 2.

#### 417 ***Ecological contamination index (ECI)***

418 The ecological contamination index (*ECI*) is a new and reliable index that can  
 419 assess the cumulative ecological risk of heavy metal contamination in sediments  
 420 [14••]. *ECI* is an aggregative experimental approach based on modified hazard  
 421 quotient (*mHQ*), as shown in the following equation:

422 
$$ECI = B_n \sum_{i=1}^n mHQ_i \quad (17)$$

423 where  $B_n$  is the reciprocal of the derived eigenvalue of heavy metal  
 424 concentrations. The heavy metal contamination levels are classified into seven levels,  
 425 shown in Table 2.

426 ***Potential ecological risk index (PERI)***

427 The *PERI* was proposed by Hakanson [7] to assess the cumulative ecological  
 428 risk due to heavy metal contamination in the study area. *PERI* is calculated based on  
 429 the contamination factors ( $CF_i$ ) and the toxic-response factors ( $T_i$ ) of the heavy  
 430 metals, as shown in Eq. (18).

431 
$$PERI = \sum_{i=1}^n T_i \times CF_i \quad (18)$$

432 The potential ecological risk can be classified into four levels: low, moderate,  
 433 high, and significantly high risk (Table 2).

434 ***Modified risk assessment code (mRAC)***

435 The *mRAC* index is proposed to evaluate heavy metal contamination based on  
 436 their toxicity and bioavailability in sediments [57, 14••]. The toxicity and  
 437 bioavailability of heavy metals are important characteristics for determining the risk  
 438 information of heavy metals associated with sediments [14••]. The *mRAC* is estimated  
 439 by Eq. (19).

440 
$$mRAC = \frac{\sum_{i=1}^n T_i RAC_i}{\sum_{i=1}^n T_i} (\%) \quad (19)$$

441 where  $T_i$  is the toxic response factor,  $RAC_i$  is the percentage concentration of  
 442 heavy metal “*i*” that can exchange and combine with the carbonate fraction, “*n*” is the  
 443 number of heavy metals. The classification of *mRAC* is presented in Table 2.

444 ***The mean probable effects level quotient (mPEL<sub>Q</sub>) and the mean effect range***  
 445 ***median quotient (mERM<sub>Q</sub>)***

446 The  $mPEL_Q$  and  $mERM_Q$  were developed based on the probable effect level  
447 ( $PEL$ ) and the effect range median ( $ERL$ ) values, respectively [14••, 58, 59]. These  
448 indexes are used to estimate the possible adverse biological effects of multiple heavy  
449 metals present in the sediment. Classifications of adverse biological effects based on  
450  $mPEL_Q$  and  $mERM_Q$  are shown in Table 2. Their calculations are presented in Eq.  
451 (20) and (21).

$$452 \quad mPEL_Q = \frac{\sum_{i=1}^n (C_{s_i}/PEL_i)}{n} \quad (20)$$

$$453 \quad mERM_Q = \frac{\sum_{i=1}^n (C_{s_i}/ERM_i)}{n} \quad (21)$$

454 where  $PEL_i$  and  $ERM_i$  are the probable effect level and effects range median,  
455 respectively;  $C_{s_i}$  is the concentration of heavy metal “ $i$ ” in the sediment; “ $n$ ” is the  
456 number of heavy metals present in the sediments.

457 In general, the synergistic indexes can provide an overall assessment of heavy  
458 metal contamination because they can indicate the combined contamination level of all  
459 heavy metals present in surface water and sediments. This is the advantage of  
460 synergistic indexes compared to individual indexes. With their advantages, synergistic  
461 indexes have been widely used in recent studies to evaluate the synergistic  
462 contamination level of heavy metals in surface water and sediments. For instance,  $mCd$   
463 and  $PN$  calculations were used to assess the synergistic effects of heavy metals in  
464 water and sediments of the Yellow River in China and the Houjing River in Taiwan  
465 [60, 10••]. The  $HPI$  index has been used for synergistic assessment of heavy metal  
466 contamination in many different rivers such as the Swarnamukhi River Basin in India  
467 [56•], the Bogacayi River in Turkey [61], the Uglješnica River in Serbia [62], the  
468 Wen-Rui Tang River in China [63]. Besides, several synergistic indexes were also

469 commonly used in previous studies to evaluate the synergistic effect of heavy metal  
470 contamination on aquatic ecosystems such as *PERI*, *mRAC*, *ECI*, and *RI* [7, 14••, 57].

471 However, to assess the contamination level of each heavy metal, the individual  
472 indexes still play an important role when the synergistic index cannot solve this  
473 problem. The advantages of the individual indexes can overcome the disadvantages of  
474 the synergistic indexes and vice versa if used in combination [10••]. Therefore, using  
475 these two index groups in combination is widely used in recent studies on heavy metal  
476 contamination assessment. For instance, *HPI*, *CF*, and *I<sub>geo</sub>* were used to assess the  
477 heavy metal contamination levels of Cu, Ni, Fe, and Mn in the Swarnamukhi River  
478 Basin in India [56•]. The contamination of heavy metals in the Korotoa River in  
479 Bangladesh was evaluated using the *PLI*, *I<sub>geo</sub>*, and *EF* [23]. Heavy metal  
480 contamination in the Yangtze River's surface water was estimated by calculating *EF*,  
481 *I<sub>geo</sub>*, and *DC* [21]. The assessment from individuals to synergistic heavy metal  
482 contamination in sediments of the Tigris River in Turkey was also conducted using  
483 *EF*, *CF*, *PLI*, and *I<sub>geo</sub>* [24].

#### 484 **5. Human health risk assessment indexes**

485 Because of the rapid increase in the effects of heavy metal pollution on human  
486 health in recent times, the human health risk assessment of heavy metal exposure is  
487 essential. Many adverse effects of heavy metals on human health, including  
488 carcinogenic and non-carcinogenic risks, have been reported [64]. Various methods  
489 for estimating human health risks have been developed, including carcinogenicity and  
490 non-carcinogenicity for both individual and cumulative effects [32]. Carcinogenic  
491 risks are usually estimated by the lifetime cancer risk index (*CR*), and the cumulative  
492 cancer risk (*CCR*) [65], as shown in Eq. (22) & (23). While the hazard quotient (*HQ*)  
493 and the total hazard index (*HI*) are used to assess non-carcinogenic risks [64, 65], as

494 shown in Eq. (24) & (25). There are no adverse non-carcinogenic effects on human  
 495 health if  $HQ$  and  $HI$  values are  $\leq 1$  and possible negative health risks if  $HQ$  and  $HI$   
 496 values are  $> 1$ . The carcinogenic risk is negligible if  $CR_i$  and  $CCR$  values are  $\leq 10^{-6}$ ,  
 497 unacceptable if they are  $> 10^{-4}$ , and acceptable if these values are between  $10^{-6}$  and  $10^{-4}$ .

$$499 \quad CR_i = ADD_i \times CSF_i \quad (22)$$

$$500 \quad CCR = \sum_{i=1}^n CR_i \quad (23)$$

$$501 \quad HQ_i = \frac{ADD_i}{RfD_i} \quad (24)$$

$$502 \quad HI = \sum_{i=1}^n HQ_i \quad (25)$$

503 where: “ $n$ ” is the number of heavy metals;  $ADD_i$  is the average daily dose;  $CSF_i$   
 504 is the cancer slope factor;  $RfD_i$  is the reference dose for each heavy metal “ $i$ ”.

505 The average daily dose ( $ADD$ ) is usually estimated based on the exposure  
 506 pathway of heavy metals to humans. The main exposure pathways of heavy metals in  
 507 rivers to humans are food ingestion, water ingestion, and dermal contact while  
 508 swimming, and the equations to estimate  $ADD$  are Eq. (26), (27), & (28), respectively  
 509 [29, 15, 66].

510 *Food ingestion:*

$$511 \quad ADD_{FI} = U \times C_{food} \times \frac{EF_{FI} \times ED_{FI}}{AT \times BW} \text{ (mg kg}^{-1}\text{d}^{-1}\text{)} \quad (26)$$

512 where:  $ADD_{FI}$  is the average daily dose of food ingestion;  $U$  is the food  
 513 ingestion rate;  $C_{food}$  is the concentration of heavy metals in food ( $\text{mg kg}^{-1}$ );  $EF_{FI}$  is the  
 514 exposure frequency ( $EF_{FI} = 365 \text{ d yr}^{-1}$ );  $ED_{FI}$  is the exposure duration;  $AT$  is the  
 515 averaging time ( $AT = 365 \text{ d yr}^{-1} \times ED_{FI}$ );  $BW$  is the body weight.

516 *Dermal contact while swimming:*

$$517 \quad ADD_{DC} = SA \times PC \times CF \times C_{water} \times \frac{ET_{SW} \times EF_{SW} \times ED_{SW}}{AT \times BW} \text{ (mg kg}^{-1}\text{d}^{-1}\text{)} \quad (27)$$

518 where:  $ADD_{DC}$  is the average daily dose for dermal contact during swimming;  
519  $SA$  is the skin area available for contact;  $PC$  is the permeability constant of heavy  
520 metals ( $\text{cm h}^{-1}$ );  $CF$  is the unit conversion factor ( $CF = 1 \text{ L (1000 cm}^3\text{)}^{-1}$ );  $C_{water}$  is the  
521 heavy metal concentration in surface water;  $ET_{SW}$  is the exposure time for swimming;  
522  $EF_{SW}$  is the exposure frequency for swimming;  $ED_{SW}$  is the exposure duration.

523 *Incidental water ingestion while swimming:*

$$524 \quad ADD_{WI} = CR \times C_{water} \times \frac{ET_{SW} \times EF_{SW} \times ED_{SW}}{AT \times BW} \text{ (mg kg}^{-1}\text{d}^{-1}) \quad (28)$$

525 where  $ADD_{WI}$  is the average daily dose for incidental water ingestion during  
526 swimming;  $CR$  is the contact rate while swimming.

527 Individual and cumulative human health risk assessments of heavy metal  
528 contamination in the global rivers have been widely presented in previous studies [32,  
529 64, 43]. For instance, human health risks due to fish consumption in the Yangtze  
530 River, China, were estimated to pose adverse health effects to adults ( $HI = 2.17$ ) [32].  
531 Water consumption from the Pardo River, Brazil, is a health concern for the local  
532 population due to the non-carcinogenic risks exceeding the maximum recommended  
533 level [64]. Conclusions of the level of risk have been drawn based on the risk  
534 quantification to provide accurate assessments. Combining the individual indexes  
535 ( $CR_i, HQ_i$ ) and the cumulative index ( $CCR, HI$ ) in different exposure scenarios can  
536 comprehensively assess heavy metal effects on human health.

## 537 **6. Advantages and limitations of indexes and future perspectives**

538 The use of indexes in assessing heavy metal contamination and its adverse  
539 effects on the ecosystem and human health is widely adopted for various advantages.  
540 For example, indexes are simple, easy-to-implement, and effective methods to  
541 provide preliminary assessments of the adverse impacts of contamination on  
542 ecosystems and human health. Calculating indexes may help to quantify the

543 magnitude of heavy metal contamination levels. This means that large data sets will  
544 be represented in a simpler way that minimizes data volume. While complex  
545 information is simplified, the index results can be easily communicated to the public.

546       However, using indexes to assess the contamination level and adverse effects  
547 also present certain limitations. The calculation formulas of the indexes are built  
548 based on some characteristics of the contaminants, such as heavy metal abundance in  
549 the environment and normalization of heavy metal concentrations to background  
550 concentrations. The calculation formulas of the indexes are also based on the different  
551 types of samples (surface water or sediments). These bases oversimplify the  
552 complexity of heavy metal pollution in the environment. In other words, the overall  
553 assessment cannot be provided by an index. Besides, the lack of background values  
554 for each locality leads to less accurate results. Choosing which background values and  
555 standards to use for specific situations also often confuses decision-makers.

556       Therefore, future perspectives for overcoming current limitations in this  
557 research area are proposed as follows:

- 558 a). Research on the quantification of the heavy metal background concentrations  
559 needs to be promoted for each local area to provide the input data for the index  
560 calculation methods. This helps the determination of pollution and risk levels of  
561 heavy metals more accurately for each locality.
- 562 b). Standards, regulations, thresholds, and toxicity values for heavy metals need to be  
563 continuously researched, developed, and updated to improve the reliability of the  
564 heavy metal pollution assessment based on the index calculation.
- 565 c). To assess pollution or risk levels for several pollutants in multiple media, new  
566 developments for multitasking and multipurpose indexes that can integrate several  
567 purposes (e.g., assessing pollution or risk levels for the diversity of pollutants) and

568 can use in different environments (e.g., water, soil, and air), are highly  
569 recommended in the near future.  
570 d). Nowadays, with the development of computer science, the integration of the  
571 indexes into models is an effective direction. This direction can help to limit errors  
572 during the calculation process and save time.

## 573 **7. Conclusions**

574 In this study, basic information on heavy metals, their effects on ecosystems and  
575 humans, and contamination status in rivers worldwide have been summarized.  
576 Contamination and the effect levels of heavy metals on the environment are  
577 increasing, especially in Asia, Africa, and Europe. This leads to more attention from  
578 scientists on methods to assess the heavy metal contamination and effect levels.  
579 Individual and synergistic indexes have been developed that are simple and efficient  
580 assessment methods. The individual indexes, including  $I_{geo}$ ,  $CF$ ,  $EF$ ,  $PCI$ ,  $HQ$ , and  
581  $mHQ$  were used for each heavy metal. The synergistic indexes, including  $DC$ ,  $mCd$ ,  
582  $PLI$ ,  $PI_{Nemerow}$ ,  $MPI$ ,  $MI$ ,  $HPI$ ,  $CSI$ ,  $ECI$ ,  $PERI$ ,  $mRAC$ ,  $mPEL_Q$ ,  $mERM_Q$  were  
583 employed for all heavy metals in the environment. The indexes can quantify the  
584 contamination levels and the effects of heavy metals in rivers, contributing to having a  
585 more accurate assessment. However, the lack of background values for the specific  
586 areas limits these methods. This review provides comprehensive information on  
587 heavy metal contaminations and their assessment methods. These results will be a  
588 useful reference for future studies, especially for index calculation methods.

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596

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598 The authors declare that they have no known competing financial interests or personal  
599 relationships that could have appeared to influence the work reported in this paper.

600

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1002

1003 **List of Tables**

1004

1005 Table 1. Heavy metal contamination in (a) surface water and (b) sediments of global  
1006 rivers

1007 Table 2. Classifications of heavy metal contamination and adverse effect levels on  
1008 ecosystem based on assessment indexes

1009 Table 1. Heavy metal contamination in (a) surface water and (b) sediments of global rivers

1010 **a)**

Region	Heavy metal concentration in surface water (mg L <sup>-1</sup> )	As	Cd	Cr	Cu	Pb	Ni	Zn	Hg	References
Asia	Yangtze River, China	0.00097	0.0004	0.0013	0.0028	0.002	-	0.031	0.00004	[32]
	Wen-Rui Tang River, China	-	0.00098	0.00532	0.0209	0.00423	-	0.0721	0.00003	[63]
	Beijiang River, China	0.01953	0.00043	-	0.003	0.00224	-	0.01636	0.00002	[33]
	Gomti River, India	-	0.1	-	0.02	0.02	0.04	0.07	-	[36]
	Beas River, India	-	0.005	0.031	0.004	0.081	-	0.22	-	[67]
	Kali River, India	-	0.06	0.06	-	0.13	-	21.71	-	[37]
	Ganga, India	-	0.02141	0.0502	0.03128	0.09578	0.04052	0.05041	-	[68]
	Buriganga River, Bangladesh	-	0.059	0.114	0.239	0.119	0.015	0.33	-	[34]
	Bangshi River, Bangladesh	0.024	0.007	0.093	1.05	0.108	0.035	3.32	-	[35]
	To Lich River, Vietnam	-	-	0.0029	0.0045	0.0081	-	0.0511	-	[25]
	Soan River, Pakistan	-	-	0.01	0.02	0.65	-	0.015	-	[69]
	Haraz River, Iran	0.05535	0.00265	-	0.01325	0.0044	-	0.05275	-	[70]
Asia rivers and lakes (1970 – 2017)	-	0.01071	0.12804	0.03759	0.03605	0.09162	0.20805	-	[31••]	
Africa	Ismailia Canal, Egypt	-	0.00045	-	0.007	0.018	-	0.015	-	[71]
	Nile River, Egypt	-	0.0045	-	0.007	0.018	0.01	0.015	-	[71]
	Challawa River, Nigeria	-	-	0.924	0.39	0.84	0.21	2.227	-	[38]
	Nairobi River, Kenya	-	-	1.11	0.05	7.5	-	0.48	-	[39]
	Mkuju River, Tanzania	0.009785	0.00008	0.01008	0.0095	0.03935	0.00495	0.03247	-	[72]
	Nil River, Algeria	-	0.00032	-	0.000061	0.00058	-	0.00038	-	[73]
	Africa rivers and lakes (1970 – 2017)	-	0.003	0.03694	0.03417	0.03405	0.02973	0.05918	-	[31••]
Europe	Mala Welna River, Poland	-	0.003	0.009	0.089	0.04	0.015	0.115	-	[74]
	Guadalquivir River, Spain	-	0.000015	-	0.00264	0.000178	-	0.00158	-	[75]
	Odiel River, Spain	4.686	0.589	0.18	122	1.985	4.429	466	-	[76]

	Tigris River, Turkey	0.0024	0.0014	< 0.005	0.1650	0.0003	0.0720	0.0370	-	[24]
	Bogacayi River, Turkey	0.00043	0.00023	0.0032	0.00092	0.00048	0.00347	-	-	[61]
	Gironde Estuary, France	-	0.00005	-	0.001403	0.000242	-	0.0061	-	[77]
	Estuary of Marche, Italy	-	0.000045	-	-	0.000315	-	0.002183	-	[78]
	Uglješnica River, Serbia	0.0002	0.01473	-	0.00408	0.04855	-	0.00998	0.00022	[62]
	Erenik River, Kosovo	-	0.007	0.029	0.044	0.014	-	0.042	-	[79]
	Europe rivers and lakes (1970 – 2017)	-	0.00062	0.00725	0.00848	0.00757	0.00395	0.09607	-	[31●●]
South America	Pardo River, Brazil	0.00214	0.00005	0.00188	0.00328	0.0018	0.00975	0.0133	-	[64]
	South American rivers and lakes (1970 – 2017)	-	0.00271	0.01248	0.01535	0.0325	0.01087	0.08347	-	[31●●]
North America	Mississippi River, US	-	0.00057	0.0002	0.0021	0.00031	-	0.0016	-	[80]
	Illinois River, US	-	0.0006	0.021	0.001	0.002	0.002	0.031	-	[81]
	North American rivers and lakes (1970 – 2017)	-	0.00361	0.02664	0.03813	0.02483	0.00584	0.19693	-	[31●●]
Latin America and Caribbean	San Pedro River, Mexico	0.16	0.014	0.212	0.2	-	0.3	-	-	[43]
	Chanchas River, Peru	0.0143	-	-	0.00257	0.00101	-	0.000375	-	[44]
	Puyango River Basin, Ecuador	0.015465	-	-	-	0.0274	-	-	0.0000046	[42]
	World average	0.00062	0.00008	-	0.00168	0.000079	-	0.0006	-	[82]
	Fresh toxicity reference values	0.15	0.002	0.011	0.009	0.003	0.052	0.018	-	[45]
Standards	Human health ambient water quality criteria of US EPA	0.000018	-	-	1.3	-	0.61	7.4	-	[40]
	WHO's drinking water quality guidelines	0.01	0.003	0.05	2	0.01	0.07	-	0.006	[41]

1011 **b)**

Region	Heavy metal concentration in the sediments (mg kg <sup>-1</sup> dry wt.)	As	Cd	Cr	Cu	Pb	Ni	Zn	Hg	References
Asia	Yangtze River, China	25.8567	0.4200	58.4667	46.4733	37.7533	-	148.8100	0.1933	[32]
	Jialu River, China	-	2.93	60.8	39.22	29.35	42.44	107.58	-	[83]
	Yongding River, China	-	-	47.61	24.71	35.47	40.45	94.75	-	[26]
	Yellow River, China	31	-	84.5	-	52	-	-	31	[84]
	Lianshan River, China	-	53.18	-	116.50	112.28	57.41	633.85	-	[85]

	Gomti River, India	-	5.0	16.2	23.2	46.2	23.9	76.3	-	[36]
	Subarnarekha River, India	-		111	69	75	42	100	-	[86]
	Swarnamukhi River Basin, India	-	0.2	85.25	100.9	21.39	2.43	63.4	-	[56●]
	Ganga, India	-	79.07	190.4	43.0	210.615	57.74	231.88	-	[68]
	Buriganga River, Bangladesh	15.54	7.74	530	62.1	65.16	47	52.975	-	[34]
	Korotoa River, Bangladesh	27.00	2.8	118	82	63	103	-	-	[23]
	Bangshi River, Bangladesh	1.93	0.61	98	-	60	-	-	-	[35]
	To Lich River, Vietnam	83.90	4.4	107.9	87.7	67.1	64.8	477.9	-	[25]
	Soan River, Pakistan	-	1.37	10.73	17.64	27.86	28.00	45.18	-	[69]
	Haraz River, Iran	33.55	3.50	28.05	32.10	26.35	43.55	73.80	-	[87]
	Al-Hawizeh Marsh, Iraq	3939.6	42.5	419.1	145.1	1602.4	-	-	-	[88]
	Ismailia Canal, Egypt	232.50	5.40	-	44.70	26.60	38.40	110.60	-	[89]
	Nile River, Egypt	-	-	274	81	23.2	112	221	-	[90]
	Asejire Reservoir, Nigeria	-	-	0.03	43.68	72.02	0.05	20.86	-	[91]
Africa	Qua Iboe River, Nigeria	-	5.67	28.52	43.72	231.52	2.6	-	-	[14●●]
	Okumeshi River, Nigeria	-	1.32	0.87	-	0.45	-	-	-	[92]
	Winam Gulf, Kenya	-	4.8	46.1	71.5	82.5	-	170.0	-	[93]
	Nil River, Algeria	-	2.34	-	38.38	61.50	96.20	-	-	[73]
	Odra River, Poland	95.33	8.47	64.67	99.33	113.33	51.00	1054.67	-	[46]
	Guadaira River, Spain	2	3	38	25	20	37	51	-	[94]
	Tinto River, Spain	-	12	151	2700	13,400	36	5280	-	[48]
	Tigris River, Turkey	12.44	7.90	158.35	2860.25	660.11	534.58	1061.54	-	[24]
	Yeşilirmak River, Turkey	-	0.55	-	38.7	17.3	79.2	45.5	-	[95]
	Gironde Estuary, France	-	1.11	-	36.62	58.99		235.08	-	[77]
Europe	Rivers of Latvia, Latvia	-	0.99	-	14.08	21.10	21.96		-	[96]
	River Po, Italy	-	3.7	-	90.1	98.5	161	645	-	[30]
	Lambro River, Italy	-	2.1	-	90.1	98.5	161.0	305.0	-	[30]
	Uglješnica River, Serbia	-	-	-	-	-	-	-	-	[62]
	Pasvik River, N. Fennoscandia	-	3.84	-	6495	62	6490	439	-	[12]
	Danube River, Central and western Europe	388	32.9	556.5	8088	541.8	173.3	2010	-	[47]
	Erenik River, Kosovo	-	-	625.0	62.3	14.8	-	157.0	-	[79]

	Axios River, Greece	40	11	180	93	140	188	271	-	[97]
	Tees River, UK	-	5.95	-	76.9	6880	-	1920	-	[98]
South America	Pardo River, Brazil	0.68	0.045	24.525	17.735	8.27	6.75	34.86	-	[64]
North America	Illinois River, US	-	2	-	19	28	-	81	-	[81]
	South Platte River, US	31	22	71	480	270	-	3700	-	[99]
Latin America and Caribbean	Rimac River, Peru	1543	31	71	796	2281	23	8076	-	[27]
	Almendares River, Cuba	-	4.3	23.4	420.8	189	-	708.8	-	[100]
	San Jorge River, Colombia	1.8	1159	-	6656	7.2	105	1064	0.31	[101]
	Culiacan River Estuary, Mexico	-	0.55	-	27.95	29.2	45.85	115.5	-	[102]
	Siete River, Ecuador	842.8	0.73	-	483.7	20.3	5960.9	132.5	1	[103]
	World sediment river average	-	1	100	100	150	90	350	-	[104]
	Surface rock average	-	0.13	97	32	20	49	129	-	
Standards	NOAA ERL <sup>a</sup>	8.2	1.2	81	34	46.7	20.9	150	0.15	[49]
	NOAA ERM <sup>a</sup>	70	9.6	370	270	218	51.6	410	0.71	[49]
	TRV <sup>b</sup>	6	0.6	26	16	31	16	110	-	[45]

"-" Not available; <sup>a</sup> National Oceanic and Atmospheric Administration (NOAA)'s effects range low (ERL) and effects range median (ERM); <sup>b</sup> Freshwater sediment toxicity reference value

1013 Table 2. Classifications of heavy metal contamination and adverse effect levels on  
 1014 ecosystem based on assessment indexes

Index	Classification	Contamination degree	References
<i>I<sub>geo</sub></i>	$I_{geo} < 0$	Uncontaminated	[9]
	$0 \leq I_{geo} < 1$	Uncontaminated to moderately contaminated	
	$1 \leq I_{geo} < 2$	Moderately contaminated	
	$2 \leq I_{geo} < 3$	Moderately to heavily contaminated	
	$3 \leq I_{geo} < 4$	Heavily contaminated	
	$4 \leq I_{geo} < 5$	Heavily to extremely contaminated	
	$I_{geo} \geq 5$	Extremely contaminated	
<i>CF</i>	$CF < 1$	Low degree	[7]
	$1 \leq CF < 3$	Moderate degree	
	$3 \leq CF < 6$	Considerable degree	
	$CF \geq 6$	Very high degree	
<i>EF</i>	$EF < 2$	No enrichment	[105]
	$2 \leq EF < 5$	Moderate enrichment	
	$5 \leq EF < 20$	Significant enrichment	
	$20 \leq EF < 40$	Very high enrichment	
	$EF \geq 40$	Extremely high enrichment	
<i>PCI</i>	$PCI < 1$	Low contamination	[12]
	$1 \leq PCI < 3$	Moderate contamination	
	$PCI \geq 3$	Severe or very severe contamination	
<i>HQ</i>	$HQ < 0.1$	No adverse effects	[13]
	$0.1 \leq HQ < 1$	Potential hazards	
	$1 \leq HQ < 10$	Moderate hazards	
	$HQ \geq 10$	High hazards	
<i>mHQ</i>	$mHQ < 0.5$	Nil to very low severity of contamination	[14●●]
	$0.5 < mHQ < 1$	Very low severity of contamination	
	$1 < mHQ < 1.5$	Low severity of contamination	
	$1.5 < mHQ < 2$	Moderate severity of contamination	
	$2 < mHQ < 2.5$	Considerable severity of contamination	
	$2.5 < mHQ < 3$	High severity of contamination	
	$3 < mHQ < 3.5$	Very high severity of contamination	
$mHQ > 3.5$	Extreme severity of contamination		
<i>DC</i>	$DC < 8$	Low degree	[7]
	$8 \leq DC < 16$	Moderate degree	
	$16 \leq DC < 24$	Considerable degree	
	$DC \geq 24$	Very high degree	
<i>mCd</i>	$mCd < 1.5$	Uncontaminated	[16]
	$1.5 \leq mCd < 2$	Slightly contaminated	
	$2 \leq mCd < 4$	Moderately contaminated	
	$4 \leq mCd < 8$	Moderately to heavily contaminated	
	$8 \leq mCd < 16$	Heavily contaminated	
	$16 \leq mCd < 32$	Severely contaminated	
	$mCd \geq 32$	Extremely contaminated	
<i>PLI</i>	$PLI = 0$	Perfection	[17]
	$PLI < 1$	Baseline level	
	$PLI > 1$	Contaminated	
<i>PI<sub>Nemerow</sub></i>	$PI < 0.7$	Unpolluted	[54]
	$0.7 \leq PI < 1$	Slightly polluted	
	$1 \leq PI < 2$	Moderately polluted	
	$2 \leq PI < 3$	Heavily polluted	
	$PI \geq 3$	Severely polluted	
<i>MPI</i>	$MPI < 1$	Unpolluted	[55]
	$1 \leq MPI < 2$	Slightly polluted	
	$2 \leq MPI < 3$	Moderately polluted	
	$3 \leq MPI < 5$	Moderately-heavily polluted	
	$5 \leq MPI < 10$	Heavily polluted	
	$MPI \geq 10$	Severely polluted	



<i>MI</i>	$MI < 0.3$	Very pure	[18]
	$0.3 \leq MI < 1$	Pure	
	$1 \leq MI < 2$	Slightly affected	
	$2 \leq MI < 4$	Moderately affected	
	$4 \leq MI < 6$	Strongly affected	
	$MI \geq 6$	Seriously affected	
<i>HPI</i>	$HPI < 100$	Safe for human consumption	[8]
	$HPI \geq 100$	Not safe for human consumption	
<i>CSI</i>	$CSI < 0.5$	Uncontaminated	[19]
	$0.5 \leq CSI < 1$	Very low severity of contamination	
	$1 \leq CSI < 1.5$	Low severity of contamination	
	$1.5 \leq CSI < 2$	Low to moderate severity of contamination	
	$2 \leq CSI < 2.5$	Moderate severity of contamination	
	$2.5 \leq CSI < 3$	Moderate to high severity of contamination	
	$3 \leq CSI < 4$	High severity of contamination	
	$4 \leq CSI < 5$	Very high severity of contamination	
	$CSI \geq 5$	Ultra-high severity of contamination	
<i>ECI</i>	$ECI < 2$	Uncontaminated	[14●●]
	$2 < ECI < 3$	Uncontaminated to slightly contaminated	
	$3 < ECI < 4$	Slightly to moderately contaminated	
	$4 < ECI < 5$	Moderately to considerably contaminated	
	$5 < ECI < 6$	Considerably to highly contaminated	
	$6 < ECI < 7$	Highly contaminated	
	$ECI > 7$	Extremely contaminated	
<i>PERI</i>	$PERI < 110$	Low risk	[7]
	$110 \leq PERI < 220$	Moderate risk	
	$220 \leq PERI < 440$	High risk	
	$PERI \geq 440$	Significantly high risk	
<i>mRAC</i>	$mRAC < 1\%$	No potential adverse effect	[57]
	$1\% \leq mRAC < 9\%$	Low potential adverse effect	
	$10\% \leq mRAC < 29\%$	Medium potential adverse effect	
	$30\% \leq mRAC < 49\%$	High potential adverse effect	
	$mRAC \geq 50\%$	Very high adverse effect	
<i>mPEL<sub>Q</sub></i>	$mPEL_Q \leq 0.1$	Low degree of contamination	[58]
	$0.1 < mPEL_Q \leq 1.5$	Medium-low degree of contamination	
	$1.5 < mPEL_Q \leq 2.3$	High-medium degree of contamination	
	$mPEL_Q > 2.3$	High degree of contamination	
<i>mERM<sub>Q</sub></i>	$mERM_Q \leq 0.1$	Low priority site	[59]
	$0.1 < mERM_Q \leq 0.5$	Medium-low priority site	
	$0.5 < mERM_Q \leq 1.5$	High-medium priority site	
	$mERM_Q > 1.5$	High priority site	

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