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Advances in nanoparticles from gasoline direct injection engines: a focus on physical and chemical characterisation

Xiaochen Wang ^a, Wei-Hsin Chen ^{b,c,d}, Jianbing Gao ^{e,f,*}, Yuhan Huang ^g

^a School of Automobile, Chang'an University, Xi'an, 710064, PR China

^b Department of Aeronautics and Astronautics, National Cheng Kung University, Tainan 701, Taiwan

^c Research Center for Smart Sustainable Circular Economy, Tunghai University, Taichung 407, Taiwan

^d Department of Mechanical Engineering, National Chin-Yi University of Technology, Taichung 411,
Taiwan

^e School of Mechanical Engineering, Beijing Institute of Technology, Beijing, 10081, PR China

^f Institute for Transport Studies, University of Leeds, Leeds LS2 9JT, UK

^g Centre for Green Technology, School of Civil and Environmental Engineering, University of
Technology Sydney, NSW 2007, Australia

Xiaochen Wang, xc.wang@chd.edu.cn / wangxiaochen510@163.com

School of Automobile, Chang'an University

***Corresponding author:**

Jianbing Gao, gaojianbing@bit.edu.cn / redonggaojianbing@163.com

School of Mechanical Engineering, Beijing Institute of Technology

Abstract: With an increasing market share of gasoline direct injection (GDI) vehicles, high nanoparticulate emissions of GDI engines are of increasing concern due to their adverse impacts on both human health and the ecological environment. A thorough understanding of GDI nanoparticulate properties is required to develop advanced particulate filters and assess the exhaust toxicity and environmental impacts. Therefore, this paper aims to fill this gap by providing a comprehensive review of the physical and chemical characteristics of GDI nanoparticles from a distinctive perspective, including soot oxidation reactivity, morphology, nanostructure, surface chemistry, chemical components, and their correlations. This review begins with a brief description of in-cylinder soot and exhaust particulate features and a comparison between GDI and diesel particulates. Then, the impacts of engine operating parameters and aftertreatments on GDI soot features are discussed in detail. The characteristics of soot particles from oxygenated fuels are introduced. Finally, the conclusions and future research recommendations are presented.

Keywords: Nanoparticles; gasoline direct injection engines; soot oxidation reactivity; nanostructure; surface chemistry; chemical component

Nomenclature

AFR	Air/fuel ratio	XRF	X-ray fluorescence spectroscopy
bTDC	before top dead centre	XPS	X-ray photoelectron spectroscopy
DMC	Dimethyl carbonate	TWC	Three-way catalyst
DMF	2,5-dimethylfuran	ICP-OES	Inductively coupled plasma optical emission spectrometry
EC	Elemental carbon	BMEP	Brake mean effective pressure
EGR	Exhaust gas recirculation	EEPS	Engine exhaust particle sizer
XRD	X-ray diffraction	IMEP	Indicated mean effective pressure
GDI	Gasoline direct injection	PAHs	Polycyclic aromatic hydrocarbons
VOF	Volatile organics fraction	EDX	Energy dispersive X-ray spectrometer
GPF	Gasoline particulate filter	FTIR	Fourier-transform infrared spectroscopy
LNT	Lean NO _x trap	SAXS	Small-angle X-ray scattering
OC	Organic carbon	EELS	Electron energy loss spectroscopy
PFI	Port fuel injection	DSC	Differential scanning calorimetry
PM	Particulate matter	TEM	Transmission electron microscopy
PN	Particulate number	HRTEM	<i>High-resolution</i> transmission electron microscopy
RS	Raman spectroscopy	GC-MS	Gas chromatography and mass spectrometry
SCR	Selective catalytic reduction	SEM	Scanning electron microscopy
TGA	Thermogravimetric analysis	SMPS	Scanning mobility particle sizer
TOC	Thermal/optical carbon		

1 Introduction

2 *1.1 Background*

3 Nanoparticle emissions from vehicle exhaust cause adverse effects on both human health
4 and the environment. A recent study indicated that each $10 \mu\text{g}/\text{m}^3$ increase in ambient $\text{PM}_{2.5}$
5 concentration was associated with a 0.55-0.74% increase in mortality from all-cause,
6 cardiovascular, and respiratory [1]. In addition, extensive studies also linked the high ambient
7 particulate levels with human health such as lung and cardiovascular diseases, incident brain
8 tumors or even daily mortality [2]. Compared to diesel engines fitted with diesel particulate
9 filters, gasoline direct injection (GDI) engines emit more ultrafine particles (less than 100 nm)
10 [3], of which a high proportion could be deposited in the lung. With the popularity of GDI
11 vehicles in recent years, GDI particulates have become an emerging environmental concern
12 which has prompted the introduction of increasingly stringent particulate regulations. Taking
13 China 6b as an example, the particulate mass (PM) and particulate number (PN) emissions for
14 light-duty gasoline vehicles (Type I test) are limited to 3.0 mg/km and 6.0×10^{11} #/km,
15 respectively. These regulations have brought the control of ultrafine particulates from GDI
16 engines into sharp focus [4]. Meeting such regulations not only requires engine and fuel co-
17 optimisation but also necessitates the application of gasoline particulate filters (GPFs). To
18 improve the performance of GPFs, a profound understanding of the soot oxidation process are
19 necessary [5, 6].

20 Numerous studies performed on diesel engines indicate that soot oxidation reactivity
21 depends on its morphology, internal structure and chemical composition [7]. Therefore, a
22 fundamental knowledge of PM characteristics is crucial to developing more effective GPF
23 management systems and extending the catalyst life of GPF. Recently, Al Housseiny, et al. [8]
24 concluded that soot surface chemistry (i.e. oxygen content and the presence of carboxylic acid
25 groups in soot) affected inflammatory and oxidative responses in human lung epithelial cells.
26 In addition, Kwon, et al. [9] reviewed the characteristics of ultrafine particles from engines that
27 are relevant to human health. Polycyclic aromatic hydrocarbons (PAHs), which are the main

28 organic compounds of exhaust PM, are the genotoxic mutagens that form DNA adducts [10].
29 The physical and chemical characteristics of soot particles can provide useful information on
30 the PM formation and development of advanced particulate filters and help researchers to assess
31 exhaust toxicity and environmental impacts [6, 11-15].

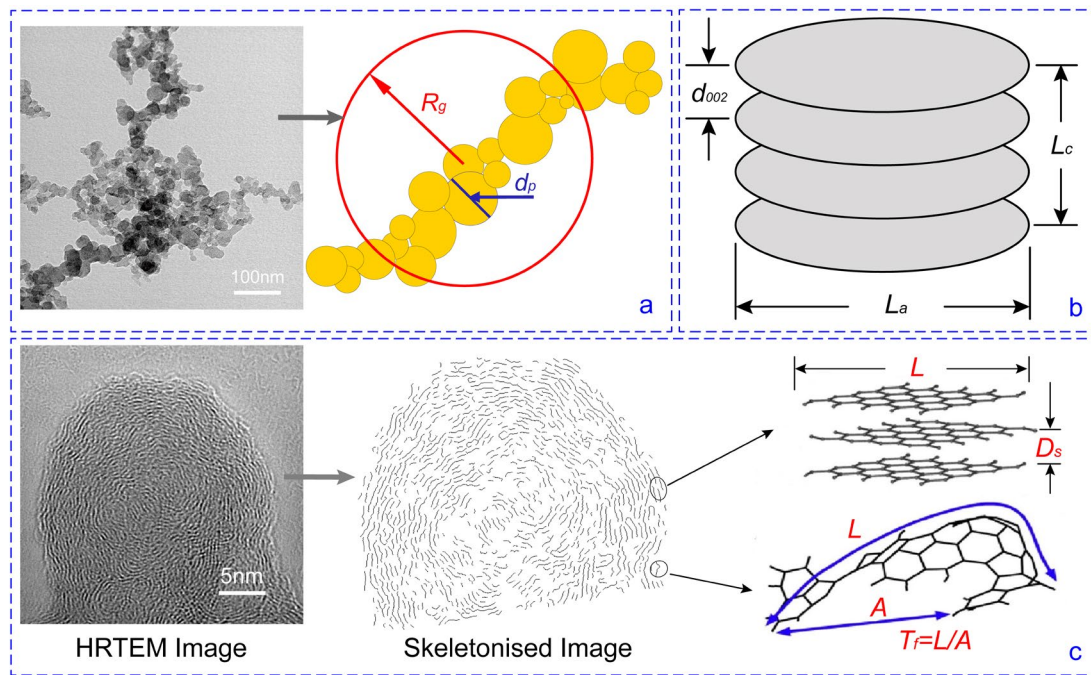
32 1.2 Soot characterisation methods

33 Collection methods may alter the properties of the particulate samples. Therefore, the
34 sampling methods should be carefully selected according to the research purposes and
35 performance of analytical techniques, thus to minimise the test uncertainties. It is recommended
36 that a thermophoretic probe is the most appropriate method for morphological and
37 nanostructural characterisation, whereas a particle trap is the most suitable method for thermal
38 and chemical surface analyses [16].

39 Soot oxidation reactivity is usually measured using thermal analysis techniques, such as
40 thermogravimetric analysis (TGA) [17, 18] and differential scanning calorimetry (DSC) [5, 19].
41 With controlled temperature and gas composition during the reaction, TGA measures the
42 variations of the loss of soot mass and DSC measures the energy released by soot oxidation.
43 Based on the measurements, the ignition temperature and activation energy are used to assess
44 the oxidation reactivity of soot samples [20, 21].

45 Detailed morphological features of soot particles at micro- and nano-scales are observed
46 by transmission electron microscopy (TEM) and *high-resolution* TEM (HRTEM), respectively.
47 Extracted from TEM images, the morphological parameters, including the mean primary
48 particle diameter (d_p), the radius of gyration (R_g) and fractal dimension (D_f), are often used to
49 study the primary particle and agglomerate sizes (see Figure 1a) [22, 23]. As shown in Figure
50 1c, HRTEM images are widely used to investigate the soot nanostructure by introducing the
51 fringe parameters including fringe length (L), tortuosity (T_f) and interlayer spacing (D_s) [24-26].
52 Likewise, X-ray diffraction (XRD) provides information on the soot crystallinity which is
53 defined by the interplanar distance between the graphene layers (d_{002}), stacking thickness (L_c)
54 and crystallite size (L_a) (as presented in Figure 1b) [27]. Note that L obtained from HRTEM

55 may qualitatively agree with L_a obtained from XRD [6]. However, as reported by Guerrero
 56 Peña, et al. [28], the crystallite size obtained from XRD is overestimated. In addition, Raman
 57 spectroscopy (RS) is frequently used to supplement the information regarding soot
 58 nanostructure [29]. Extracted from Raman spectra, the Raman parameters (the intensity/area
 59 ratios of bands) are used to study the degree of the order of soot samples [30].



60

61 Figure 1. Schematic of morphological features of soot particles at micro- and nano-scales: (a) an agglomerate
 62 particle, (b) lattice parameters estimated from XRD spectra (reproduced from Ref. [27]) and (c) fringe
 63 parameters from HRTEM images.

64 Chemical characterisation of PM usually contains the element composition, carbonaceous
 65 components (elemental carbon (EC) and organic carbon (OC)) and surface functional groups
 66 (SFGs). The most common method to reveal the elemental composition of soot particles may
 67 involve scanning electron microscopy (SEM) (or TEM) combined with an energy dispersive X-
 68 ray spectrometer (EDX) [31, 32]. The thermal/optical carbon (TOC) analyser is applied to
 69 determine the fractions of EC and OC [33, 34]. Regarding soot SFGs, both Fourier-transform
 70 infrared spectroscopy (FTIR) and X-ray photoelectron spectroscopy (XPS) are employed [35,
 71 36]. Through both spectra with curve fitting, the concentrations of SFGs, such as aliphatic C-
 72 H, hydroxyl group (C-OH) and carbonyl group (C=O), are obtained. PAHs are also evaluated,

73 which are adsorbed onto the PM or portioned in the semi-volatile PM phase. PAHs are first
74 extracted from PM samples and then analysed using gas chromatography and mass
75 spectrometry (GC-MS) [37-39]. Although the above analytical methods have been commonly
76 used in most publications, some other techniques are also infrequently utilised. Among them,
77 the techniques that are used to characterise soot nanostructure and graphitisation include
78 electron energy loss spectroscopy (EELS), X-ray near edge absorption spectroscopy and small-
79 angle X-ray scattering (SAXS). Moreover, nuclear magnetic resonance, X-ray fluorescence
80 spectroscopy (XRF) and inductively coupled plasma optical emission spectrometry (ICP-OES)
81 are employed to determine the PM's elemental and chemical compositions. Detailed
82 information regarding the fundamentals of these analytical techniques and their applications in
83 soot characterisation is beyond the scope of this work. For further details, the reader is referred
84 to several previous reviews [6, 11, 40-43] and the references therein.

85 ***1.3 Motivation and outline of this review***

86 Recently, researchers have conducted numerous experimental and numerical studies to
87 investigate the soot formation mechanisms and mitigation measures. There are a few reviews
88 on the PM and PN emissions from GDI engines. Qian, et al. [44] systematically summarised
89 the PM emissions from modern gasoline-fuelled engines, including the effect of fuel properties,
90 engine operating conditions and driving cycles on PM mass and PN emissions. Additionally,
91 PM and PN emissions in various GDI engines and their control methods to meet the current
92 emission regulations have been presented by Überall, et al. [4], Raza, et al. [45], and Awad, et
93 al. [46]. Although several reviews focused on the features of soot particles from diesel engines
94 [6, 11, 12], a comprehensive summary of soot physicochemical features from GDI engines is
95 still absent given GDI soot is attracting more and more attention. Moreover, GDI soot has
96 significant differences with diesel soot in physicochemical features and the correlations among
97 these features.

98 Therefore, this paper aims to fill this gap by providing a comprehensive review of the
99 physical and chemical characteristics of GDI soot particles. This study will focus on

100 physicochemical features, including soot oxidation reactivity, soot morphology and
101 nanostructure, graphitisation degree, chemical composition, and surface chemistry. Although
102 the above features are imperative to characterise GDI PM, they cannot be fully examined due
103 to the limited technique accessibility. Consequently, the present review tries to build a whole
104 picture of GDI soot features based on vast experimental investigations. The remaining of this
105 paper is organised as follows. First, Section 2 introduces the PM characteristics of GDI engines,
106 including both in-cylinder and exhaust soot, and the comparisons between the properties of GDI
107 and diesel exhaust particulates. Next, Section 3 systematically describes the effects of engine
108 operating parameters on the characteristics of GDI PM, including engine speed and load,
109 equivalence ratio, fuel injection strategies and exhaust gas recirculation. Then, Sections 4
110 discusses in detail the effects of the application of exhaust after-treatment technologies on GDI
111 PM characteristics, namely three-way catalyst (TWC) and GPF. Finally, the conclusions and
112 future perspectives are presented. This review is expected to provide a thorough understanding
113 of GDI soot properties which are critical for the assessment of health and environmental impacts
114 of GDI nanoparticles and the development of effective mitigation measures.

115 2 PM characteristics of GDI engines

116 2.1 *In-cylinder soot features*

117 The formation of soot particles in combustion engines involves complex physical and
118 chemical processes, including fuel pyrolysis, nucleation, coalescence, surface growth,
119 agglomeration and oxidation [44, 47]. The evolution of in-cylinder soot structure can provide
120 useful information on the actual soot formation process inside the engine combustion chamber
121 [48]. For this purpose, Kook et al. [49-52] developed a novel in-cylinder soot sampling system
122 in an optical wall-guided GDI engine. Unlike other studies which used optical diagnosis [53]
123 and numerical simulation [54] methods, they used TEM images of the sampled particles to
124 characterise their morphology and internal nanostructure, which provided detailed information
125 regarding the in-cylinder soot status.

126 TEM images showed that in-flame soot exhibited chain-like and/or ring shapes, and the

127 samples that were located on the exhaust valve side had a higher particle concentration and
 128 larger primary particles and agglomerates than those located on the intake valve side [49]. As
 129 shown in the natural flame luminosity images, those on the exhaust valve side were surrounded
 130 by wall wetting-induced poor fire, which persisted for a longer period. Therefore, longer soot
 131 residence time enhanced soot surface growth and aggregation, thus leading to larger primary
 132 particles and agglomerates [49]. Subsequently, the changes in the morphology of the in-flame
 133 particles were investigated under piston- and liner-wetting conditions, which were obtained
 134 from the selected start of the injection (SOI) of 320 °CA and 180 °CA before top dead centre
 135 (bTDC), respectively. The results showed that early injection with piston-wetting produced soot
 136 particles with a larger gyration radius and primary particle diameter, lower fractal dimension
 137 and a more graphited structure, which was indicated by a longer fringe length, less tortuosity
 138 and smaller layer spacing compared to the later injection with liner wetting, as shown in Table
 139 1 [50]. These phenomena can be explained by the fact that the soot particles that were collected
 140 from the piston-wetting condition suffered a longer residence time, which led to a more
 141 carbonised structure.

142 Table 1. Quantitative analysis of in-cylinder soot morphology and nanostructure under various wall-wetting
 143 conditions (data from Ref. [50]).

Average values of structural parameters	Wall-wetting conditions	
	Piston-wetting (SOI = 320°CA bTDC)	Liner-wetting (SOI = 180°CA bTDC)
d_p	23.91 ± 0.24 nm	22.01 ± 0.19 nm
R_g	68.55 ± 3.33 nm	60.82 ± 1.93 nm
D_f	1.69	1.71
L	1.01 ± 0.02 nm	1.00 ± 0.01 nm
T_f	1.18 ± 0.01	1.19 ± 0.01
D_s	0.405 ± 0.002 nm	0.432 ± 0.001 nm

144 Note that in Refs. [49-52], three to seven injection cycles (i.e. firing cycles) were adopted
 145 for each sampling run, and the soot was directly deposited onto the TEM grid surface due to
 146 the thermophoretic force. Therefore, the time-sequential in-cylinder soot features are missing.
 147 This problem can be addressed by the total cylinder sampling system [55], which has long been
 148 applied in diesel engines [56]. At a predetermined sampling crank angle, the electronic control
 149 unit cut diaphragms in the engine cylinder head. This allows the in-cylinder gas to escape

150 rapidly out of the combustion chamber into a sampling bag, which is pre-filled with nitrogen
151 [57]. With this sampling apparatus, the evolution of in-cylinder diesel soot features in various
152 combustion phases has been widely analysed [36, 58, 59]. Therefore, further research is
153 necessary to reveal the in-cylinder GDI soot features as a function of the crank angle by
154 combining the total cylinder sampling system in GDI engines and multiple analytical
155 techniques, especially for soot chemical features.

156 *2.2 Exhaust particulate features*

157 Compared to in-flame soot, exhaust soot showed a lower fringe separation and more
158 defined core-shell boundaries at both injection pressures [51], which means that exhaust soot is
159 more difficult to oxidise than in-cylinder soot. The result is in agreement with the subsequent
160 work of the authors [50, 52], who reported that exhaust soot exhibited a longer fringe length,
161 smaller fringe tortuosity and fringe separation. This indicates a more ordered nanostructure,
162 regardless of injection timing and test fuel.

163 GDI soot reactivity is a key factor that affects GPF regeneration behaviour. Soot with
164 higher reactivity is easier to oxidise, which benefits GPF regeneration. Generally, soot reactivity
165 is inherently governed by its physicochemical features. As soot reactivity is often evaluated by
166 the TGA methodology, it is influenced by the heating program, oxidising conditions and sample
167 load [60, 61]. There are many examples of soot oxidation kinetics in the literature. According
168 to Hu, et al. [62], the activation energy of soot particles derived from the worldwide harmonised
169 light-duty vehicle test cycle (WLTC) was in the range of 91 kJ/mol–133 kJ/mol, depending on
170 the oxidation temperature and calculated methodology. In Ref. [63], the activation energy of
171 soot samples that were collected at five different engine conditions ranged from 125 kJ/mol to
172 142 kJ/mol.

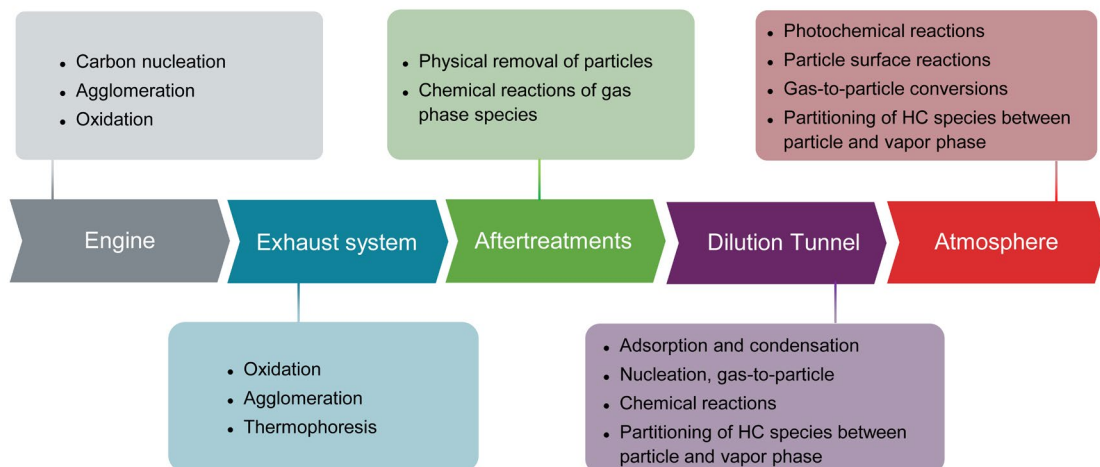
173 TEM images show that GDI exhaust particles can be classified into the following two
174 groups: carbonaceous particles and non-carbonaceous particles [64]. The former usually
175 exhibits cluster-like and chain-like soot [64], while the latter contains Ca-rich, S-rich, Fe-rich
176 and Zn-rich particles, with morphologies of nearly circular, irregular and prismatic shapes [31].

177 It is worth noting that ultrafine ash particles are in almost equal amounts to or higher than soot
178 particles, which maybe because a large amount of soot that is generated in the combustion
179 chamber tends to form soot agglomerates that are larger than 100 nm [31]. This result suggests
180 that the larger amount of ash in ultrafine particles should be taken into consideration to evaluate
181 their environmental and human health effects. As ash particles mainly originate from the in-
182 cylinder consumed lubricating oil, Seong, et al. [65] identified engine oil-derived ash
183 nanoparticles by dosing various formulated engine oil into the fuel system of a GDI engine.
184 The STEM results showed that ash particles exist as single or multi-agglomerated forms and
185 are separated from or deposited on aggregated soot.

186 Generally, HRTEM images reveal that GDI exhaust soot exhibits the core-shell
187 nanostructure and presents a more graphited structure than in-cylinder soot. Moreover, the
188 primary particles with an outer shell that is interrupted by multiple cores were observed, which
189 infers that in the combustion chamber, an abundant growth of nuclei and the coalescence of
190 soot may simultaneously occur before each core grows into a primary particle [66]. In addition
191 to the core-shell structure (also regarded as a ‘mature’ structure), a small number of amorphous
192 particles (i.e. an isotropic graphite crystalline but with no obvious graphene layers) and
193 disordered particles (i.e. no crystalline structure) were observed [37].

194 Regarding the element components, an EDX analysis of soot particles emitted from a GDI
195 vehicle (Euro V) over the WLTC shows that the main components are C and O, followed by S,
196 whereas the minor components include Ca, P, Na, Si, Zn (mainly from lubricating oil) and Fe
197 (mainly from engine wear) [67]. This is generally consistent with another study [68], which
198 identified using inductively-coupled plasma mass spectrometry that the emission factors of the
199 main elements in PM can be ordered from high to low as follows: S > Fe > Ca > Al > P > Na.
200 As reported in Ref. [69], under the WLTC, the particulate-PAH emissions from GDI vehicles
201 were $4.8 \pm 1.4 \mu\text{g}/\text{km}$, which is three orders of magnitude higher than those from PFI vehicles.
202 Moreover, the benzo [a] pyrene (BaP) emission factors for GDI vehicles were 4.3 times higher
203 than those for PFI vehicles.

204 Figure 2 shows particle and gas phase processes of diesel engine particulate matter
 205 emissions into the atmosphere. Particulates from the engine exhaust valve into the atmosphere
 206 involve the exhaust system, the aftertreatment devices, the dilution tunnel and the atmosphere.
 207 Along the exhaust pipe, the changes and/or reactions occur in sequence, including oxidation,
 208 agglomeration and thermophoresis in the exhaust system and the physical removal and
 209 chemical reactions in the aftertreatment devices. After exiting the exhaust pipe, the particle
 210 processes cover HC adsorption and condensation; nucleation in the dilution tunnel; and
 211 diffusion and continuous dilution in the atmosphere [70, 71]. Although the effect of
 212 implementing TWC and GPF on soot features has been partly studied, studies on variations in
 213 soot features along the entire exhaust path of GDI engines are limited (see Section 3.2).
 214 Previously, Gao, et al. [72] found that diesel PM collected at different tailpipe positions has
 215 different physicochemical properties. However, the change in the GDI particulate features as a
 216 function of the exhaust distance and local temperature is unknown. Regarding the particle
 217 processes that occur in the dilution tunnel and the subsequent atmosphere, the soot morphology
 218 and chemical composition may change significantly due to both transport and physical removal
 219 processes and chemical transformations [73]. Therefore, the differences in GDI soot features
 220 from the exhaust valve into the atmosphere must be clarified in the future.



221
 222 Figure 2. Schematic of the particle and gas-phase processes of diesel engine PM emissions into the atmosphere
 223

(reproduced from Ref. [71]).

224 **2.3 Comparisons between GDI and diesel particulates**

225 Under the high combustion temperature and pressure conditions in both GDI and diesel
226 engines, some similarities in the soot features are expected between diesel and GDI exhaust
227 particles. However, the physicochemical features of soot particles from GDI and diesel engines
228 have some inherent differences in terms of morphology, nanostructure, chemical components
229 and oxidation reactivity due to the differences in the fuel properties, mixture formation
230 strategies, ignition and combustion processes. Therefore, a comparison between diesel and GDI
231 exhaust particles is made by checking their similarities and differences regarding soot reactivity
232 and its related features.

233 Through TGA, Oo, et al. [74] identified that GDI exhaust soot had a faster oxidation
234 process and a lower oxidation temperature, which demonstrates that it is easier to oxidise than
235 diesel exhaust soot. This is in good agreement with the observation of Wang-Hansen, et al. [75].
236 The more reactive nature of GDI soot indicates that it may be advantageous to the regeneration
237 of the particulate filter. Note that both studies on the comparison of soot oxidation were
238 performed under a given atmosphere. However, the exhaust environments for GDI and diesel
239 engines are different. For example, GDI engines with stoichiometric combustion have lower O₂
240 concentrations in the absence of NO₂ downstream of a three-way catalyst (TWC) compared to
241 diesel engines. Therefore, a further comparative study on soot oxidation for both engines could
242 be conducted under each exhaust-like gas atmosphere.

243 Concerning soot morphology, TEM images showed that diesel particulate was composed
244 of nearly all aggregates while GDI particulate was composed of a complex mixture of
245 aggregates covering single spheroids, nascent soot, and irregular soot [76]. As observed by
246 Gaddam [77], the GDI soot was ‘open and chain-like’, whereas the diesel soot was more like a
247 ‘compact assembly’. The differences in morphology can be verified by the morphological
248 parameters. The author pointed out that the D_f of the gasoline soot was 2.03, which was lower
249 than that of diesel soot with 2.39. This means that the GDI soot agglomerates exhibit a less
250 compact and more branched morphology [77]. However, Lu, et al. [78] identified using SEM

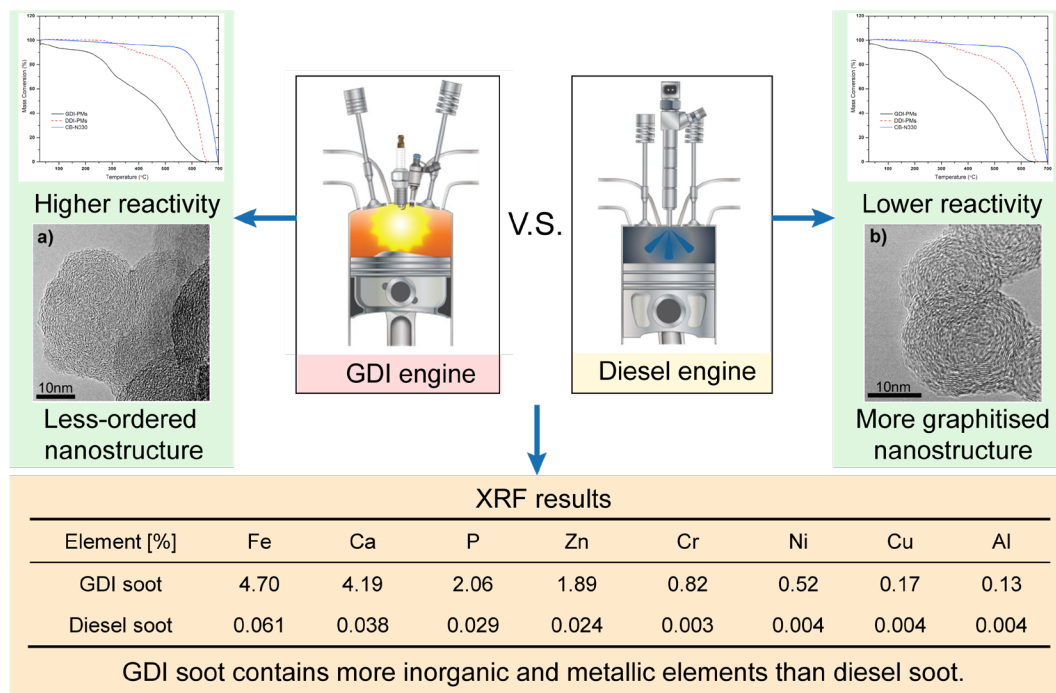
251 that the D_f of the soot sample from the gasoline vehicle was 1.769 ± 0.006 , which lies in between
252 that of particulates from the heavy-duty diesel vehicle (1.752 ± 0.014) and light-duty diesel
253 vehicle (1.789 ± 0.076). In addition, in Ref. [79], the D_f of the diesel particulates was within
254 the range of 1.70–1.90, whereas that of the GDI particulates was located in wider distribution,
255 ranging from 1.45 to 2.21. Unlike D_f , GDI soot often exhibits a smaller mean primary particle
256 diameter than diesel soot [74, 79, 80].

257 Regarding the nano-scale level, HRTEM images reveal that GDI soot presents a more
258 amorphous and disordered structure, whereas diesel soot shows a more graphitic and organised
259 morphology [66, 77]. Likewise, the skeletonised images show that the GDI soot contains more
260 irregularly patterned, shorter and more curved fringes than diesel soot [81]. Correspondingly, a
261 shorter fringe length [74, 77] and larger tortuosity and interlayer spacing [79] were observed
262 for the GDI particulate samples. In contrast to diffusion combustion in diesel engines, it is
263 speculated that premixed combustion in GDI engines provides fewer hydrocarbons and carbon
264 in the soot formation process, thus decreasing the graphitisation of particles. Additionally, the
265 difference in fuel properties between gasoline and diesel, particularly the content and species
266 of aromatic hydrocarbons, may partly account for the observation. Compared to diesel fuel,
267 gasoline contains a larger proportion of smaller PAHs, with less electronic stabilisation and
268 thermodynamic stability. During fuel pyrolysis, smaller PAHs tend to produce C_2 fragments
269 and further PAHs containing five-membered rings, and the latter has a more curved
270 nanostructure [82]. The less-ordered structure of GDI soot agrees with the higher soot oxidation
271 reactivity obtained from TGA.

272 Regarding the elemental composition, Oo, et al. [74] determined through TEM-EDX that
273 GDI soot had a lower fraction of carbon but a higher percentage of oxygen than diesel soot. As
274 analysed by XRF, the particulate sample that was collected from a GDI vehicle contained more
275 oil additive elements, such as P, Ca, Zn, and wear metals, such as Fe and Al, compared to a
276 diesel vehicle [80]. The XPS results show that only C, O and N are detected in diesel soot,
277 whereas additional inorganic elements, such as S, P and Zn, are observed in GDI soot [80].
278 Regarding PAH emissions, the patterns of PAHs of GDI vehicles are similar to those of diesel

279 vehicle, considering the larger numbers of nanoparticles and the adsorbed genotoxic PAHs.
 280 Furthermore, note that the sum of 16 US Environmental Protection Agency (EPA) priority
 281 control PAHs in GDI exhausts (without GPF) is higher than that in diesel exhausts (equipped
 282 with DPF) [38]. The result implies that GDI vehicles would need particulate filters to reduce
 283 both nanoparticles and genotoxic PAHs.

284 Figure 3 illustrates the differences between GDI and diesel particulates. The three
 285 differences are summarised as follows: i) GDI soot is generally more reactive and easier to
 286 oxidise than diesel soot; ii) GDI soot exhibits smaller primary particles with a less-ordered
 287 nanostructure than diesel soot, whereas the trend in fractal dimension is not conclusive; and iii)
 288 GDI soot contains more inorganic and metallic elements (such as P, Ca and Zn) than diesel soot,
 289 which originate from lubricating oil additives and engine wear.



290
 291 Figure 3. Differences between GDI and diesel particulates (TG curves and XPF results were from Oo, et al. [74],
 292 TEM images were from Uy, et al. [80]).

293 3 Influences of engine operating parameters on GDI soot features

294 Various engine operating parameters influence the in-cylinder combustion conditions and

295 exhaust atmosphere, which can affect the physicochemical characteristics of GDI exhaust
296 particles. Many studies have investigated the impact of operating conditions on soot features.
297 This section focuses on this topic in detail.

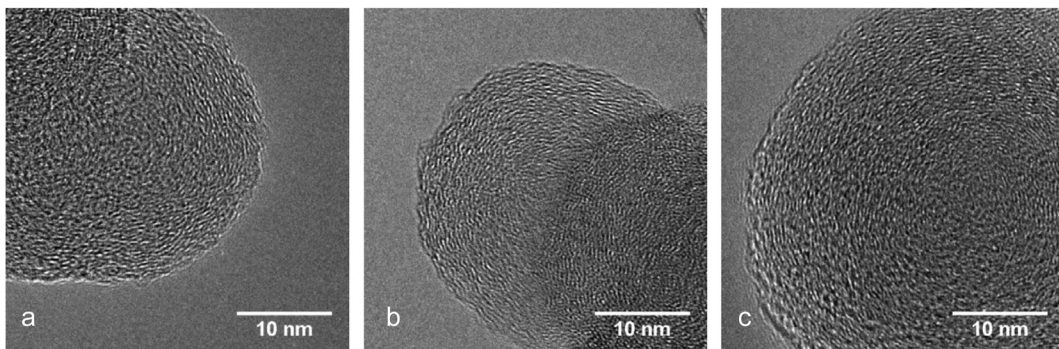
298 *3.1 Engine speed and load*

299 Choi, et al. [63] evaluated the soot oxidation reactivity of GDI particulates generated at
300 the cold-idle and the following two different steady-state conditions: 1250 r/min-25% load and
301 1500 r/min-50% load. In contrast to diesel soot, variations in GDI soot reactivity are less
302 sensitive to the engine operating point. The authors explained that in GDI engines, flame
303 propagation combustion provides a more consistent in-cylinder fuel-air mixture and flame
304 temperature conditions, irrespective of the engine load and speed, which is different from the
305 nature of diffusion combustion in diesel engines. In addition, they ascribed the differences in
306 GDI soot reactivity to the changes in organic fractions and ash content with the various
307 operating conditions rather than the soot structure. However, other studies [39, 83] found that
308 soot that is produced at a higher engine load is more difficult to oxidise.

309 Collected at various operating points, agglomerated particles exhibit similar morphology
310 with a chain-like structure that comprises a few to hundreds of primary particles [81, 84].
311 However, the trends regarding the primary particle size with the change in engine load are
312 conflicting. Seong, et al. [81] observed that the mean primary particle diameter increased from
313 20 nm to 29 nm, as the engine load increased from 25% to 75% at the speed of 1500 r/min.
314 Likewise, Gaddam, et al. [84] found that elevated engine load (increasing fuel quantity from 11
315 mg/cycle to 21 mg/cycle) increased the primary particle size from 21 nm to 25 nm and resulted
316 in a larger gyration diameter and fractal dimension. However, Sharma, et al. [85] reported that
317 the mean primary particle diameter of gasoline soot decreased from 31.22 nm to 20.57 nm when
318 the load was increased from 50% to 100%. The authors explained that a higher maximum in-
319 cylinder temperature at higher load conditions leads to the predominance of particle oxidation
320 over particle nucleation.

321 Figure 4 presents the HRTEM images of GDI soot collected at various operating points

322 [63]. Visually, the three samples have the typical core-shell nanostructure, as the inner core
323 contains fringes with random orientations, and the outer shell is composed of parallel and
324 concentrically oriented lamellae. Additionally, the visual examination reveals that similar
325 orders of crystalline structures are observed amongst the three samples. In particular, in the
326 cold-idle condition, the soot particle presents a clear concentric fringe pattern. Although the
327 Raman Spectra further validates that GDI particulate samples have a similar soot nanostructure,
328 irrespective of engine operating points, no fringe parameters that were extracted from the
329 HRTEM images were quantified [63]. Quantitative fringe parameters from another work [77]
330 show that soot that is generated at a high load (fuel quantity = 21 mg) has a shorter fringe length
331 but a larger fringe tortuosity than soot collected at a low load (fuel quantity = 11 mg). However,
332 the fringe discrepancies in both conditions are slight. Zhang, et al. [86] pointed out that the
333 interlayer spacing of idle soot was the largest, followed by the soot sample at the cold start
334 condition. In addition, the interlay spacing of soot particles increased when the engine load
335 increased from 3 bar to 7 bar brake mean effective pressure (BMEP) at the fixed engine speed
336 of 2000 r/min, whereas it slightly decreased when the engine speed increased from 1200 r/min
337 to 2000 r/min at the fixed engine load of 5 bar BMEP. The fringe analysis from Ref. [77, 86]
338 indicates that an elevated engine load emits soot particles with a less-ordered nanostructure.
339 Interestingly, this trend contradicts the commonly accepted result of diesel soot that higher
340 engine loads lead to a more ordered nanostructure due to the higher temperature, more fuel-rich
341 regions and longer diffusion combustion phase [87, 88].



342
343 Figure 4. HETEM images of GDI soot collected at: (a) cold-idle; (b) 1250 r/min-25%; and (c) 1500 r/min-50%
344 (adapted from Ref. [63]).

345 Regarding the chemical composition, the PM sample that was collected at the cold-idle
346 condition has a much higher volatile organics fraction (VOF) compared to those generated in
347 hot steady-state conditions [63]. When the engine load increases, the VOF generally decreases,
348 while the soot fraction increases [39, 64]. This is because a higher engine load operation induces
349 a higher exhaust temperature, and more volatile organic compounds are emitted in gas-phase
350 forms rather than particle-phase forms. As indicated in Ref. [37], when the engine load is
351 increased from 60 N·m to 120 N·m, the vapor-phase PAHs with two, three and four rings and
352 the particulate-phase PAHs with four, five and six rings gradually increased, especially for the
353 two-ring PAH. Under the high engine load of 120 N·m, the peak concentrations of PAHs were
354 found at the engine speed of 2000 r/min. When the engine was operating at 60 N·m, the
355 minimum values were found at 2500 r/min.

356 **3.2 Equivalence ratio**

357 The equivalence ratio is an important parameter governing engine particle emissions due
358 to its influence on both the oxygen concentration and the combustion temperature. With the
359 increasing equivalence ratio, more fuel availability may result in higher soot surface growth
360 rates, along with the restraint of soot oxidation caused by the lower in-cylinder oxygen
361 concentration [89]. Moreover, the higher in-cylinder temperature under fuel-rich conditions
362 may improve both the particle surface growth and oxidation. These in-cylinder changes may
363 alter the soot features.

364 According to the previous studies [79, 84], there is no consensus regarding the impact of
365 the equivalence ratio on agglomerates and primary particle sizes. Wu, et al. [79] reported that
366 the soot particles that were collected at the air/fuel ratio (AFR) of 14.7 had the highest D_f and
367 D_p , which reveals that particulate samples under stoichiometric conditions exhibit the most
368 compactly clustered agglomerates with the largest primary particle size. They further stated that
369 under the fuel-rich condition (AFR = 13), the reduced d_p was caused by the predominant effect
370 of particle oxidation because of the high temperature. However, under the fuel-lean conditions
371 (AFRs = 16 and 18), thanks to the excess air, enhanced particle oxidation reduced the primary

372 particle size. In contrast, in Ref. [84], it was pointed out that with the increasing AFR, the mean
373 primary particle diameter increased from 19 nm to 23 nm. Endo, et al. [90] found that soot from
374 a pool fire on the wall surface exhibits a larger primary particle diameter, whereas that from
375 bulk-gas in the cylinder tends to be smaller in size, which is consistent with the results in Ref.
376 [49]. Furthermore, the authors found that a larger proportion of large particles can be observed
377 for fuel-rich combustion (ARF = 12.5), which indicates that the pool fire soot is dominant.
378 However, Seong, et al. [91] pointed out that d_p exhibited no statistically significant changes
379 with an equivalence ratio from 0.98 to 1.13. The authors explained that for various equivalence
380 ratios, the spark timings were adjusted in the experiments to keep the same CA50 (the crank
381 angle at which 50% of the cumulative heat is released).

382 Like primary particle size, the effect of the equivalence ratio on soot nanostructure is
383 inconclusive. At a nanoscale level, particulate samples from various equivalence ratios present
384 the typical shell-core structure with different degrees of graphitisation. Table 2 shows the
385 average values of soot nanostructural parameters versus AFR obtained from HRTEM. As
386 reported by Gaddam, et al. [84], soot from fuel-rich combustion (AFR = 13) displayed a more
387 amorphous structure, with the smallest fringe length but largest tortuosity amongst three
388 samples from AFRs of 17, 15 and 13. In addition, no obvious differences in the soot
389 nanostructure were observed between stoichiometric (AFR = 15) and fuel-lean (AFR = 17)
390 conditions. However, according to Wu, et al. [79], the particulate sample that was generated
391 under the stoichiometric condition showed the shortest fringe length, largest fringe tortuosity
392 and smallest fringe separation distance. Their results of the sp^2/sp^3 ratio that was obtained from
393 the XPS supported the HRTEM observation that soot particulates with a stoichiometric
394 operation exhibited a more disordered nanostructure. The authors further attributed the different
395 trends in soot structure to the different engine configurations and test fuel properties. From a
396 fundamental perspective, it can be hypothesised that increasing the equivalence ratio results in
397 more fuel but lower oxygen availability, which leads to higher carbonisation and a more ordered
398 structure. However, at the same time, a higher in-cylinder temperature caused by a higher
399 equivalence ratio affects both soot carbonisation and oxidation. Therefore, to clarify the effect

400 of the equivalence ratio on soot nanostructure, it should be combined with computational fluid
 401 dynamics (CFD) [92] or in-cylinder optical diagnostics (such as spatiotemporal flame
 402 luminosity [93]) to determine the combustion characteristics that are related to soot formation
 403 and oxidation.

404 Table 2. Soot nanostructural parameters versus AFR obtained via HRTEM.

Reference	Engine type	Operating point	Sampling method	AFR	Nanostructural parameters		
					L (nm)	T_f	D_s (nm)
<u>Gaddam, et al. [84]</u>	single cylinder, 549 cc, 11.97:1	n = 2100 r/min,	thermophoretic sampling	13	0.72	1.210	-
		fuel quantity = 11 mg/cycle		15	0.82	1.175	-
				17	0.80	1.176	-
<u>Wu, et al. [79]</u>	4-cylinder, 1.498 L, 12:1	n = 2000 r/min,	Dilution sampling	13	0.91	1.46	0.39
		fuel quantity = 46.3 mg/cycle		14.7	0.83	1.49	0.38
		49.3 mg/cycle		16	0.87	1.47	0.40
				18	0.85	1.44	0.41

405 The results that are derived from the FTIR showed that soot produced at fuel-rich
 406 combustion had significant organic content, which indicates that soot oxidation requires lower
 407 activation energy [77, 84]. However, Zelenyuk, et al. [94] determined that fuel-rich combustion
 408 increased the particulate mass with a higher EC than fuel-lean combustion. The authors stated
 409 that the number of fractal EC particles increased when the equivalence ratio was increased from
 410 0.90 to 2.10, which indicates a more incomplete combustion. In addition, more than 90% of
 411 EC-dominated particles are observed for the lean stratified (2000 r/min, 2 bar BMEP, and $\lambda =$
 412 2.10) condition, whereas PM from the lean homogeneous (2000 r/min, 4 bar BMEP, and $\lambda =$
 413 1.54) operation contains a relatively large fraction of organic particles, which are mainly
 414 composed of heavy alkanes and alkenes. This observation agrees with the results in Ref. [95],
 415 where the experiments were operated in lean-stratified, lean-homogeneous and stoichiometric
 416 modes. The results showed that pre-GPF PM in the lean-stratified mode had the largest EC,
 417 whereas the lean-homogeneous mode had the largest OC.

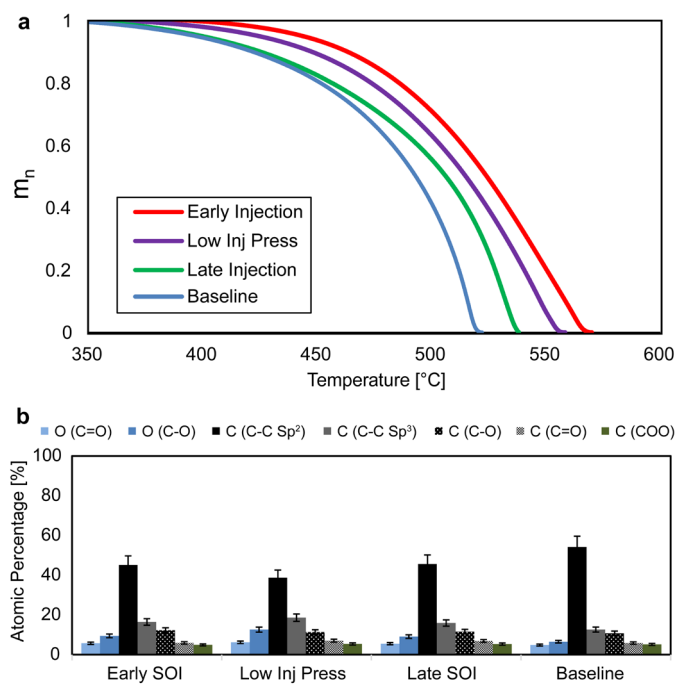
418 Lean GDI combustion can effectively increase fuel economy and reduce CO₂ emissions.
 419 However, the focus should be placed on both PM and PN emissions control under lean
 420 operations, especially under stratified-lean GDI combustion [95, 96]. In addition, the above

421 discussions reveal that there are inconsistent opinions regarding the effect of the equivalence
422 ratio on soot morphology and nanostructure. Therefore, more work is required to
423 comprehensively characterise the soot particulates from lean GDI combustion.

424 **3.3 Fuel injection pressure**

425 For pre-mixed flames in GDI engines, soot formation is dependent on the quality of fuel/air
426 mixing. Therefore, fuel injection parameters, including injection timing and pressure,
427 significantly affect the time available for premixing, fuel evaporation and fuel impingement,
428 thereby influencing soot formation and oxidation [97]. Increasing the fuel pressure reduces the
429 diameter of the fuel droplets, which improves atomisation and results in a more homogenous
430 fuel-air mixture. However, it is worth noting that an over-high injection pressure increases spray
431 penetration length, which may increase the opportunity of fuel impingement [45].

432 Koch, et al. [98] studied the variations in soot oxidation regarding the pressure changing
433 from 50 bar to 150 bar. The derivative thermogravimetry (DTG) results showed that a higher
434 oxidation temperature (defined as the temperature at the maximum oxidation rate [T_{max}]) was
435 observed for the injection pressures ($P_{inj.}$) of 50 bar and 75 bar, which indicates a lower
436 oxidation reactivity. However, the soot samples that were generated at $P_{inj.} > 87.5$ bar presented
437 a lower T_{max} (approximately 525 °C), which reveals that they are more reactive regarding
438 oxidation. Similarly, Easter [99] identified using both isothermal and non-isothermal tests that
439 soot for $P_{inj.} = 60$ bar was more reactive regarding oxidation than that for $P_{inj.} = 20$ bar (non-
440 isothermal test results were shown in Figure 6a). The trend is consistent with the fact that a
441 higher oxidation reactivity of diesel particulates was obtained at the higher injection pressure
442 conditions [11, 100, 101].



443
 444 Figure 5. TG (a) and XPS (b) results for various particulate samples with respect to injection parameters (Early
 445 injection: SOI = 325 °CA bTDC, P_{inj} = 6 MPa; Low injection pressure: SOI = 287 °CA bTDC, P_{inj} = 2 MPa;,
 446 Late injection: SOI = 150 °CA bTDC, P_{inj} = 6 MPa; Baseline: SOI = 287 °CA bTDC, P_{inj} = 6 MPa) (adapted
 447 from Ref. [99]).

448 As observed using SEM and TEM, Potenza, et al. [102] pointed out that when the rail
 449 pressure increased from 50 bar to 120 bar, the particulate cluster morphology changed from
 450 simple nanoparticle chains to nano-belts with core-shell structures, nanotubes and textured
 451 nanofoils. As reported by Kim, et al. [51], the radius of gyration and primary particle diameter
 452 for soot collected at P_{inj} = 150 bar were 50.10 ± 2.10 and 22.22 ± 0.22 nm, respectively, which
 453 are lower than 55.80 ± 3.07 and 22.53 ± 0.24 nm for soot generated at P_{inj} = 50 bar. This
 454 suggests that the higher injection pressure produced reduced soot agglomerates with smaller
 455 primary particles because of the improved soot oxidation. Extracted from HRTEM images, $L =$
 456 0.517 ± 0.046 nm and $T_f = 1.321 \pm 0.012$ of P_{inj} = 87.5 bar soot are similar to the values for P_{inj} .
 457 = 100 bar soot [98]. However, there is a significant difference regarding the fringe separation
 458 distance for the above soot samples. For example, D_s of P_{inj} = 87.5 bar soot was 0.405 ± 0.042
 459 nm, which is significantly lower than that of P_{inj} = 100 bar soot (i.e. 1.187 ± 0.042 nm) [98].
 460 Similarly, in the study by Kim, et al. [51], exhaust soot that was emitted by P_{inj} = 50 bar had a

461 lower fringe separation distance than that produced by $P_{inj.} = 150$ bar. However, the
462 discrepancies regarding the fringe separation distance between both samples were relatively
463 smaller. These results indicate that increasing injection pressure results in a more disordered
464 soot nanostructure.

465 Additionally, changes in the injection pressure influence PM composition. SEM-EDX
466 shows that low injection pressure soot only presents carbon clusters, whereas high injection
467 pressure soot exhibits traces of metallic polycrystalline agglomerates regarding the piston wear
468 phenomenon [102]. When the injection pressure increased from 150 bar to 200 bar, the EC ratio
469 tended to decrease, while the OC ratio increased [103]. Both vapor-phase and particulate-phase
470 PAH concentrations gradually reduced when the injection pressure increased. However, when
471 the injection pressure achieved 80 bar, the decline in the PAH concentration became less
472 obvious [37]. The differences in oxygen-containing SFGs between different injection pressures
473 are small, as presented in Figure 6 (b) [99]. In addition to the above discussions, there are
474 limited studies on the effect of injection pressure on aliphatic C-H SFGs. As the injection
475 pressure increased, a lower total PN concentration with a smaller geometric mean diameter was
476 observed [104, 105]. As particulates with smaller mean diameters have a more available surface
477 area for the adsorption of toxic compounds, more studies on the chemical properties of
478 particulates produced by higher injection pressures are required.

479 **3.4 Fuel injection timing**

480 Generally, early injection during the intake stroke is beneficial to homogenous fuel/air
481 mixture formation. However, fuel that is injected too early (close to the top dead centre of the
482 intake stroke) may cause fuel impingement on combustion chamber surfaces, which is the main
483 source of pool fires. This causes high PM emissions [105, 106]. As fuel injection timing shifts
484 from during the intake stroke to during the compression stroke, the time for air-fuel mixing
485 becomes insufficient, thereby increasing the PM emissions. In addition, for wall-guided
486 injection systems, the fuel spray may impinge on the cylinder walls and form a fuel film and
487 more wall wetting. The injection timing determines not only the time available for pool

488 evaporation but also the amount of fuel impinging on the piston or the cylinder walls [107]. As
489 discussed earlier, injection timing significantly influences soot formation and oxidation, and
490 further affects soot oxidation reactivity and related features.

491 Koch, et al. [98] investigated the effect of injection timing on soot oxidation reactivity by
492 varying the start of injection (SOI) from 340 °CA to 220 °CA bTDC at the step of 30 °CA. The
493 results showed that the lowest T_{max} of 520 °C (i.e. the highest oxidation reactivity) was achieved
494 for the SOI of 310 °CA bTDC, which was related to the smallest soot particles that were
495 obtained by an engine exhaust particle sizer (EEPS) under this condition. Easter, et al. [108]
496 compared the oxidation reactivity of soot particles that were produced at three injection timings,
497 namely 325 °CA, 287 °CA and 150 °CA bTDC, respectively. Figure 6 (a) shows the TG results,
498 which indicate that soot reactivity from high to low can be ordered as follows: baseline
499 (287 °CA) > retarded SOI (150 °CA) > advanced SOI (325 °CA). Furthermore, the authors
500 found that the trend in soot reactivity agrees with the ash content. The above results of sweeping
501 SOI consistently showed that a SOI that leads to the highest soot reactivity and advanced or
502 retarded injection may reduce the reactivity.

503 Morphological observation showed that GDI soot that was generated at various injecting
504 timings presented a typical chain-like agglomerate structure, which was similar to diesel
505 agglomerates [81, 109]. The TEM analysis showed that both the agglomerated gyration radius
506 and the primary particle diameter decreased as the injection timing retarded. This result is in
507 agreement with the observation of Seong, et al. [81], who set a wider injection parameter from
508 330 °CA to 190 °CA bTDC. In addition, the authors claimed that the influence of injection
509 timing on the fractal dimension was insignificant. Barone, et al. [110] found that compared to
510 the optimised SOI of 280 °CA bTDC (regarding low PN concentrations), early fuel injections
511 (320 °CA bTDC) produced more particles with a larger primary particle diameter, which was
512 attributed to the increased fuel-rich zones in the cylinder. The consistent result that was obtained
513 from the TEM images that the primary particle size decreases along with the retarded injection
514 timing was confirmed in Ref. [111] through ultra-SAXS.

515 The effect of injection timing on soot nanostructure is more complicated, and the results
516 vary. Easter [99] identified using both the Raman parameter of the D1/G peaks and visual
517 observation of TEM images that varying the SOI had a negligible effect on soot nanostructure.
518 This observation was further confirmed by quantitative analysis in Ref. [112], which reported
519 that advanced injection timing from 303 °CA to 335 °CA bTDC had a negligible effect on both
520 fringe length and tortuosity. However, although other scholars manifest that the injection timing
521 may influence the soot nanostructure, their results are inconsistent. According to Koch, et al.
522 [98], although soot particles with a SOI of 310 °CA bTDC exhibited a lower fringe length and
523 separation distance, they showed a comparable fringe tortuosity than that of a SOI of 220 °CA
524 bTDC. They concluded that a later injection resulted in poor homogenisation of the air/fuel
525 mixture, which increased the graphene layer. This was further related to higher soot reactivity.
526 Pan [113] found that regarding the retarded injection timing from 320 °CA to 240 °CA bTDC,
527 the fringe tortuosity decreased, and the separation distance increased. Note that the variations
528 in fringe tortuosity and separation distance may conflict, as reduced fringe tortuosity is usually
529 accomplished by decreasing the separation distance and creating a more graphited arrangement
530 of carbon layers [114]. Considering this, the author further confirmed using EELS that retarded
531 injection timing led to a more ordered nanostructure. On the contrary, Kim, et al. [50] pointed
532 out that an early injection with piston top wetting (320 °CA bTDC) had a smaller fringe
533 separation distance than a late injection with liner wetting (180 °CA bTDC), which reveals that
534 retarded injection timing causes a less-ordered structure. When the injection timing was
535 retarded from 280 °CA to 220 °CA bTDC, Gaddam, et al. [84] reported that the fringe length
536 decreased from 0.82 nm to 0.74 nm, whereas the fringe tortuosity slightly increased from 1.175
537 to 1.190. The contradictory findings from various researchers could be explained by the
538 different engine configurations (i.e. injector position and wall-guided/spray-guided) and fuel
539 properties (such as aromatic content and research octane number).

540 Unlike the above morphological analysis, investigations into the impact of SOI on soot
541 chemical features were limited. Easter [99] found that the VOF increased with the retarded
542 injection timing, and the differences between SOIs of 287 °CA and 150 °CA aTDC were small.

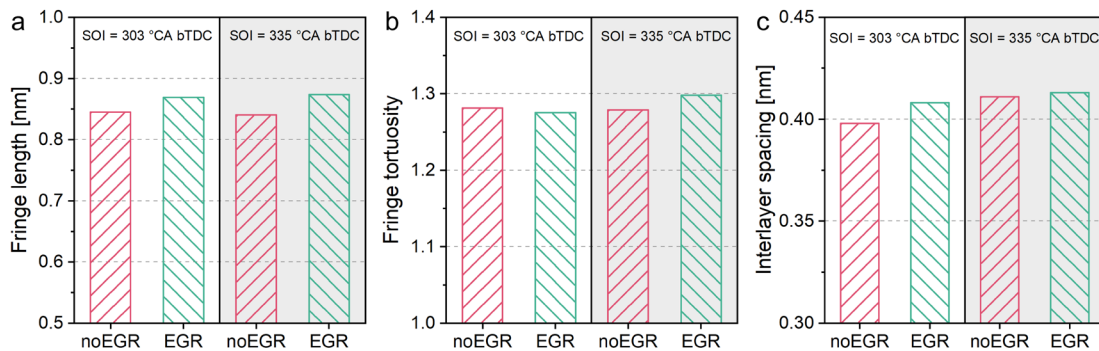
543 When obtained using the XPS, both early and late injection timings had a higher sp^3/sp^2 ratio.
544 However, the oxygen-containing SFGs, including C-O, C=O and O-C=O, were similar among
545 samples with three SOIs, as seen in Figure 6 (b) [99, 108]. In addition, Pan [113] reported that
546 PAHs with four, five and six rings generally increased with the retarded injection timing.

547 **3.5 EGR**

548 EGR is a common technique that is used to control NO_x emissions in GDI engines. At part
549 load, EGR acts as a diluent in unburnt gas, thus decreasing the peak in-cylinder temperature
550 and NO_x formation rates. The introduction of EGR could also change the PM features. Bogarra
551 et al. [115, 116] examined the oxidation reactivity of soot samples collected from non-EGR and
552 EGR with a rate of 19%, respectively. DTG curves showed that EGR soot exhibited a lower
553 oxidation temperature than non-EGR soot. However, the authors found that the difference in
554 the oxidation temperature between the two samples may be due to the different mass availability
555 of soot rather than the variations in particle structure. According to the result of calculated
556 activation energy, both samples showed similar values of approximately 100 kJ/mol, which
557 confirmed the authors' speculation [115, 116].

558 From the morphology perspective, applying EGR reduced the primary particles' size and
559 gyration radius. The reduced primary particle size may be attributed to the EGR capability of
560 inhibiting soot nucleation, which overcomes the decreasing effect of post-flame oxidation
561 caused by the lower temperature [117, 118]. The discrepancies in both the fractal dimension
562 and number of primary particles per agglomerate were small, with or without EGR [117]. With
563 a speed of 2000 r/min and the indicated mean effective pressure (IMEP) of 3 bar, applying EGR
564 from 3.3% to 18% increased both the fringe tortuosity and separation distance. However, it had
565 a negligible effect on fringe length, which indicates a more amorphous nanostructure [113].
566 Previously, Al-Qurashi, et al. [119] deconvoluted the chemical, dilution or thermal effects of
567 EGR on the diesel soot features. Considering the different soot formation mechanisms
568 regarding applying EGR for diesel and GDI engines, more fundamental research is required to
569 quantify the weight of these three effects on the variations in the morphology and nanostructure

570 of GDI soot. Figure 8 presents the effect of the EGR with a ratio of 19% on fringe parameters
 571 under different SOIs, according to the data in Ref. [112]. Interestingly, EGR slightly increased
 572 the fringe length and interlayer spacing, but no significant trends in tortuosity were observed.
 573 The authors further concluded that adopting 19% EGR had a negligible influence on soot
 574 nanostructure [112]. In diesel engines, the impact of EGR on soot nanostructure and oxidation
 575 is related to the operating point [120, 121]. In this regard, more work is required to clarify the
 576 reasons behind these different trends.



577
 578 Figure 6. Effect of EGR on fringe parameters under different SOIs: (a) fringe length, (b) fringe tortuosity, and (c)
 579 interlayer spacing (data from Ref. [112]).

580 The influence of EGR on PAH concentrations is dependent on engine operating conditions,
 581 according to various researchers [37, 113]. At 2000 r/min and IMEP = 7.5 bar, An, et al. [37]
 582 found that applying EGR with a lower ratio (i.e. 2.5%) significantly reduced the particulate-
 583 phase PAH concentration, compared with the non-EGR condition. With the further increase of
 584 the EGR ratio, a negligible change in the particulate-phase PAH concentration was observed.
 585 However, in Ref. [113], at 2000 r/min and IMEP = 3 bar, increasing the EGR ratio from 0% to
 586 18% gradually increased the particulate-phase PAH concentrations. As particulate-phase PAHs
 587 with five or six rings were predominantly absorbed on soot cores [37], it may verify the
 588 assumption that the impact of EGR on soot nanostructure and oxidation in GDI engines is
 589 related to the operating point.

590 In addition to the operating parameters that are discussed above, PM emissions from GDI
 591 engines are influenced by other factors, such as spark timing, intake boost, multiple injections

592 and injector deposits. However, there are limited studies on the impact of these factors on PM
593 features. Therefore, further research on the effects of these factors is necessary to better reveal
594 the control strategy of particulate aftertreatments of GDI engines.

595 4 Influences of exhaust aftertreatment technologies on soot features

596 Exhaust aftertreatments, such as TWC and GPF, are essential to control both gaseous and
597 particulate pollutants to meet the increasingly stringent emission regulations. Amongst them,
598 TWCs have been widely used for decades due to their effective removal of HC, CO and NO_x
599 emissions. It should be noted that the impact of TWCs on soot particles is not negligible
600 although they are not specifically designed to reduce particle emissions in the exhaust [122].
601 As HC emissions are important components of semi-volatile particles, TWCs can remove the
602 precursors of semi-volatile components and significantly reduce the exhaust particulate mass
603 and number [123]. In addition, the high temperature and the presence of catalysts in the TWC
604 may influence the soot features. As mentioned earlier, TWCs greatly affect PM emissions as
605 GPFs are designed to trap the exhaust particles. A previous study on diesel soot suggested that
606 understanding the changes in soot features along the aftertreatments contributes to the
607 optimisation of soot elimination [124]. Therefore, the effects of TWC and GPF on soot features
608 are summarised below.

609 4.1 TWC

610 Bogarra, et al. [112] observed that the TWC had no significant effect on soot reactivity, as
611 the residence time of PM in the TWC is too limited. However, Choi, et al. [63] stated that PM
612 that was collected from the TWC downstream showed higher soot oxidation reactivity than that
613 generated from the TWC upstream. This is attributed to the higher ash content in PM after TWC,
614 which provides the catalytic effect on soot oxidation.

615 Endo, et al. [90] found that a reduction in the mean primary particle diameter from 17.6
616 nm before TWC to 15.0 nm after TWC was observed under the operating condition. The PM is
617 primarily composed of dry soot, which is covered by a soluble organic fraction (SOF). When
618 the exhaust flow passes through, larger particles are collected on the coat in the TWC, and some

619 SOFs become oxidised due to the presence of the catalyst. This can be testified in Ref. [63],
620 where it was reported that under the given operating conditions, TWC reduced VOF from 5.14%
621 to 2.02% due to the oxidation of organic components. Therefore, the use of TWC decreases the
622 size of the primaries. Regarding the soot nanostructure, Bogarra, et al. [112] pointed out that
623 the fringe parameters exhibited no significant changes before and after TWC. The authors
624 speculated that the soot may be deposited in the TWC by diffusion rather than be oxidised.
625 However, numerous studies show that due to the presence of the catalyst, soot is oxidised easier
626 [125-127]. Such a change in the soot oxidation process may change the soot structure. It can be
627 hypothesised that a relatively higher exhaust temperature with the presence of a catalyst in the
628 TWC could influence the soot features. Therefore, more studies are required to clarify the effect
629 of TWC on soot reactivity and its related features.

630 **4.2 GPF**

631 Liu, et al. [128] pointed out that by combining thermophoretic sampling with TEM
632 observations, the use of GPF has a negligible effect on the visual morphology of soot
633 agglomerates. However, as expected, the number of primary particles in a certain TEM image
634 significantly reduces after GPF. Moreover, the D_f of agglomerates after GPF was lower than
635 that of before GPF, which indicates that the agglomerates after GPF present a more branched
636 and less compact structure [128]. According to Endo, et al. [90], the mean radii of the gyration
637 of agglomerates before and after GPF were 56.5 nm and 82.0 nm, respectively. The significant
638 increase in agglomerate size after GPF could be because particles are easily collected in the
639 GPF [90]. However, the trends in the morphological parameters of particulates from a GDI
640 vehicle during transient drive cycles are different from the above results from the steady-state
641 engine tests. Saffaripour, et al. [129] performed a comparative study on the effect of GPF on
642 the size and morphology of particulates from a GDI vehicle over the Federal Test Procedure
643 (FTP-75) and the US06 Supplemental Federal Test Procedure (US06) transient drive cycles.
644 Their results showed that the fractal dimension of soot aggregates was not affected by GPF.
645 Moreover, over the whole range of driving cycles, particulates from the vehicle without GPF
646 presented slightly higher agglomerates but smaller primary particle sizes than those collected

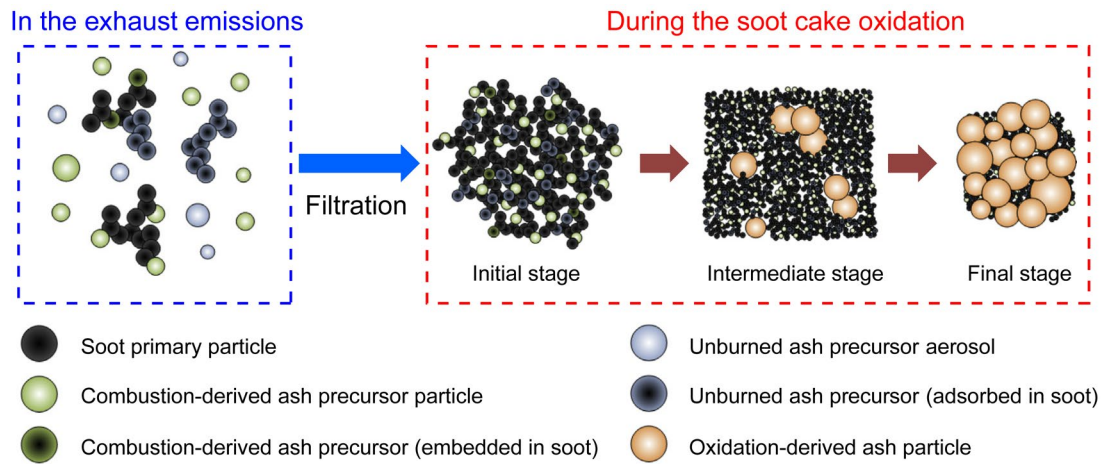
647 from the GPF-equipped vehicle.

648 Extracted from HRTEM images, soot collected from the GPF downstream showed a
649 smaller mean fringe length, fringe tortuosity and interlayer spacing than the sample from the
650 GPF upstream [128]. Soot with a more ordered structure generally presents a longer fringe
651 length and smaller tortuosity and interlayer spacing. Based on the results reported in Ref. [128],
652 it is difficult to quantify and classify whether the reduced fringe length or smaller tortuosity and
653 interlayer spacing is governing the trends in soot nanostructure. Therefore, a supplementary
654 analysis using Raman spectrometer or/and XRD could be conducted to assess the variations in
655 soot nanostructure between the GPF upstream and downstream.

656 As reported in Ref. [130], the application of GPF significantly reduces the emissions of
657 particulate- and vapor-phase PAHs. Moreover, for various GDI vehicles over the LA92 cycle,
658 employing GPF reduced the particulate- and vapor-phase PAHs by 98% and 57.5% on average,
659 respectively. A similar phenomenon was observed by Munoz, et al. [131], who pointed out that
660 the effects of GPF on removing semi-volatile compounds were not as strong as for solid
661 particles. This is attributed to the volatility of the semi-volatile PAHs, which enables them to
662 escape from the soot-loaded filter at higher temperatures. Having compared four types of GPFs,
663 it is uncertain whether GPFs with coated catalysts could remove PAHs. However, GPFs
664 exhibited higher filtration efficiencies in PAHs over the cold WLTC than over the hot WLTC
665 [131].

666 In modern GDI engines, both TWC and GPF are the indispensable exhaust aftertreatments
667 that meet the Euro VI or future emission regulations. The above discussions show that although
668 some research has been conducted on the effect of TWC or GPF on the soot features, the results
669 are often contradictory. In addition, studies that focus on variations of soot oxidation reactivity
670 and related features, such as morphology, nanostructure and chemical components, in the entire
671 exhaust path of GDI engines are limited. In this regard, investigations into the variations in GDI
672 soot features along the exhaust aftertreatments, including before TWC, after TWC (i.e. before
673 GPF), after GPF and inside the GPF, are encouraged. It can be speculated that soot inside the

674 GPF has a higher oxidation reactivity than soot collected the GPF upstream. Figure 7 displays
 675 the interactions of soot and ash particles during the soot oxidation processes [63]. As stated by
 676 Choi, et al. [63], ash particles, particularly the combustion-derived ash precursor significantly
 677 promotes soot oxidation due to the catalytic effect. With the accumulation of soot particles in
 678 the filter, ash fraction increases, and thus the promoting effect on soot oxidation may be more
 679 apparent.



680
 681 Figure 7. Conceptual drawings of the interactions of soot and ash particles during the soot oxidation processes
 682 (adapted from Ref. [63]).

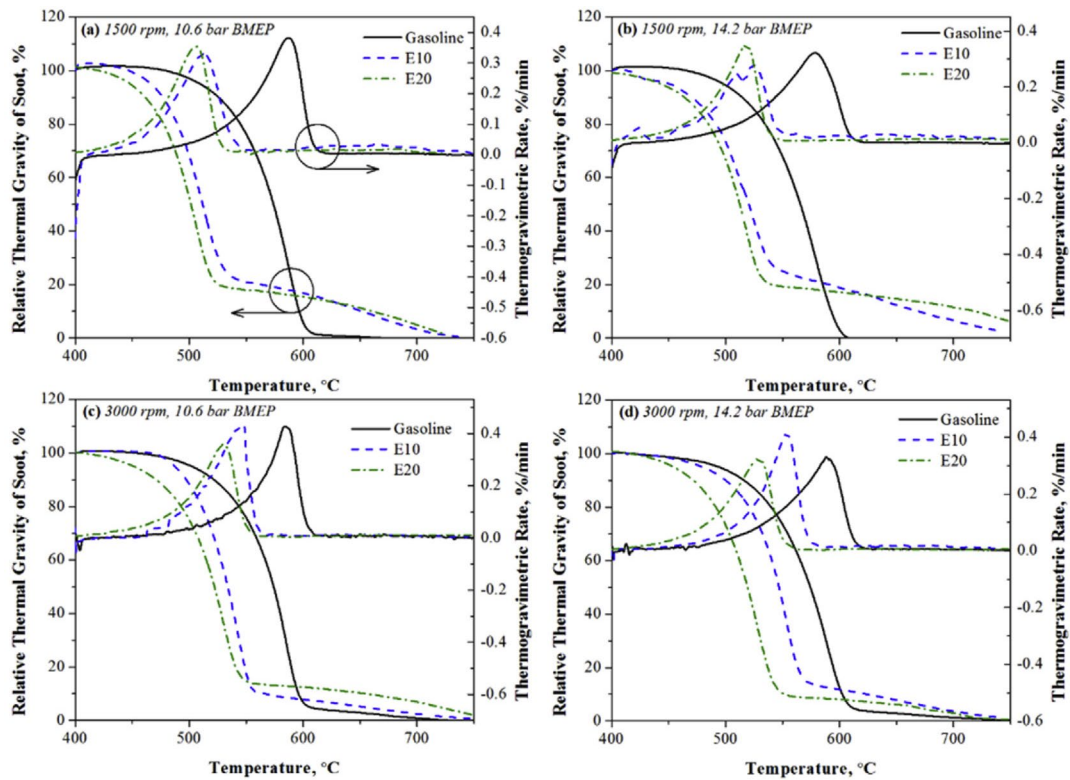
683 3 The features of soot particles from oxygenated fuels

684 Fuel reformulation changes the physical and chemical features of the fuel and influences
 685 the air-fuel mixing process and soot-forming combustion reaction pathways [132, 133]. From
 686 the soot formation perspective, the coalescence during the combustion of unsaturated
 687 hydrocarbon (especially the PAHs) initiates soot formation. Mixing with oxygenates dilutes the
 688 content of unsaturated hydrocarbon in the fuel, which benefits the inhibition of nascent soot
 689 formation. Another effect of oxygenated fuel combustion is the promotion of soot oxidation.
 690 However, in practical GDI engines, blending with oxygenated fuels may affect fuel injection
 691 and spray due to the modified vapor pressure, heating value and enthalpy of vaporisation. As
 692 these factors may compete, it remains inconclusive regarding whether mixing with oxygenated
 693 fuels causes an increase or decrease in PM mass and PN emissions from current GDI engines
 694 [44, 45]. Here, we discuss the effect of mixing with oxygenated fuels on soot features. Ethanol

695 is the most popular additive amongst the oxygenates used in GDI engines due to its high octane
696 number, oxygen content and latent heat of vaporisation [134]. In addition to the ethanol, other
697 oxygenated additives, such as methanol, n-butanol, 2,5-dimethylfuran (DMF) and dimethyl
698 carbonate (DMC), are used in GDI engines. However, their impacts on soot features have rarely
699 been evaluated, as the available literature has mainly focused on the variations in soot
700 morphological structures with the addition of these additives. Table 3 lists the summary of
701 physicochemical features of soot particles from oxygenated fuels.

702 The soot samples from gasoline-ethanol blends exhibited lower activation energies and
703 higher oxidation reactivity, which was independent of the combustion systems (air-guided or
704 wall-guided), ethanol content and operating points [39, 83]. Figure 10 presents the
705 thermogravimetric curves of particulate samples emitted from gasoline fuel and
706 gasoline/ethanol blends. With the increase of ethanol concentration from 10% to 20%,
707 oxidation temperatures (defined as the temperatures at which the maximum thermogravimetric
708 rate in Figure 10) of soot samples gradually decreased, indicating a growing trend towards
709 oxidation [39]. The activation energy of pure ethanol-derived soot was the lowest at 83 kJ/mol,
710 followed by DMF-derived soot (109 kJ/mol), E25 (25 vol% ethanol + 75 vol% gasoline)-
711 derived soot (124 kJ/mol) and gasoline-derived soot (153 kJ/mol) [83]. This is consistent with
712 the result of Guerrero Peña, et al. [135], who claimed that the diffusion flame soot of E20 (20
713 vol% ethanol + 80 vol% gasoline) was the most reactive, whereas the reactivity of soot
714 generated from gasoline was the least. Regarding DMF20 (20 vol% DMF + 80 vol% gasoline)
715 soot, the oxidation reactivity is between that of gasoline and E20. In their study, the highest
716 reactivity of E20 soot was attributed to it having the highest oxygen and hydrogen content via
717 elemental analysis, an amorphous nanostructure characterised by HRTEM, EELS and XRD and
718 the highest concentrations of oxygenated and aliphatic functional groups on its surface, as
719 examined by FTIR [135]. Moreover, soot collected from DMC8 (8 vol% DMC + 92 vol%
720 commercial gasoline) displayed a lower starting and maximum oxidation temperature than soot
721 generated from commercial gasoline, which indicates higher oxidation reactivity [136].
722 Similarly, for most diesel engines, the combustion of oxygenated fuels tends to emit soot with

723 higher oxidation reactivity [88, 137, 138].



724

725 Figure 8. Thermogravimetric curves of particulate samples emitted from gasoline fuel and gasoline/ethanol
726 blends (adapted from Ref.[39]).

727 Particulates from gasoline mixing with oxygenated fuels display spherical chain-like or
728 cluster-like agglomerates and show no significant visual differences in micro-scale morphology
729 amongst most PM samples. However, regarding E25 soot in Ref. [139], the TEM images of the
730 particles showed the coexistence of fractal-like agglomerates and some small HC droplets.
731 According to the TEM images, mixing with ethanol and DMC generally decreases the
732 agglomerate and primary particle sizes [52, 136, 139]. The presence of oxygen and the absence
733 of aromatics reduce the average number of primary particles per agglomerate, which results in
734 a smaller agglomerate size. As is widely accepted, the primary particle size results from the
735 competition of surface growth and soot oxidation. From a fuel component perspective, both the
736 dilution and oxygen effects of the addition of oxygenated fuels not only suppress the soot
737 surface growth but also promote soot oxidation. This is in agreement with the soot luminosity
738 images that were observed by Gao, et al. [52], which show that mixtures with higher ethanol

739 contents significantly reduced the pool fires and soot formation, despite possibly higher wall
740 wetting caused by a larger injection duration and higher density and viscosity. It is interesting
741 to note that the soot morphology between gasoline and B33 (33 vol% butanol +67 vol%
742 gasoline) is relatively slight, as reported by Hergueta, et al. [139]. This may be attributed to the
743 consequence of the fuel spray patterns and combustion conditions. Another explanation is the
744 coagulation mechanism related to the higher combustion temperatures of B33 combustion.
745 However, Sharma and Agarwal [85, 140] found that the impact of gasohols on the primary
746 particle size is connected to the engine load, as the d_p reduced as the fuel oxygen content at the
747 part-load increased, whereas at full load, the d_p increased as the fuel oxygen content increased.
748 Moreover, the authors stated that the soot agglomerates mainly presented chain-like structures,
749 and no significant difference was observed between gasoline and gasohols. However, Lee, et al.
750 [141] found that E20 yielded no statistically pronounced impact on the mean primary particle
751 diameter.

752 A visual inspection of the HRTEM images showed that primary particles emitted from
753 gasoline and its blends present the typical shell-core structure, i.e. graphene-like segments with
754 longer fringes at the outer shell region and PAH layers with chaotic orientations at the inner
755 core region [52, 136, 140]. However, the addition of oxygenated fuels changes the soot
756 nanostructure, as the combustion-related parameters, such as residence time, reaction
757 temperature and oxygen concentration, differ. Generally, adding oxygenates into gasoline
758 produces particles with a more amorphous structure, as shown by both HRTEM and RS [52,
759 136, 142]. When the ethanol concentrations increased from 0% to 60%, the fringe length almost
760 linearly reduced, along with the growth in both fringe tortuosity and separation distance, which
761 indicates a less graphitised structure [52]. The higher A_{D1}/A_G (area ratio of D1 band to G band)
762 and A_{D3}/A_G (area ratio of D3 band to G band) of DMC8 soot implied that it was smaller in
763 graphitic and higher in amorphous concentration [136]. This phenomenon may be due to the
764 lower in-cylinder temperature [136] and shorter soot residence time [52], which reduces the
765 graphitisation process of the graphene layers. However, as reported by Lee, et al. [141], the
766 impact of E20 on soot nanostructure is sensitive to injection timing. For example, at the SOI of

767 310 °CA bTDC, E20 soot showed a more amorphous structure than gasoline soot, whereas at
768 the SOI of 250 °CA bTDC, the primary particles from both fuels exhibited graphitic structures.

769 Regarding the chemical components, mixing with ethanol and DMC increased the VOF
770 content in PM [39, 136]. This may be because adding oxygenated fuels inhibits the formation
771 of soot precursors and reduces the fraction of elemental soot in PM. As expected, Luo, et al.
772 [39] found that E10 (10% ethanol + 90% gasoline, by vol.) decreased the concentration of most
773 PAHs, except those with small aromatic rings, such as naphthalene. Similarly, with the increase
774 of ethanol in gasoline from 10% to 83%, the elemental carbon fraction was markedly reduced
775 [34]. Moreover, Sharma, et al. [85] reported that particulates from gasohols had relatively lower
776 trace metal concentrations than that from gasoline.

777 From the publications available, mixing with oxygenated fuels increases the soot oxidation
778 reactivity, which can benefit the regeneration of GPFs, as a lower energy input is required to
779 oxidise the accumulated particles. Generally, oxygenated fuel-derived soot particles emitted
780 smaller agglomerates with smaller primaries, which could cause the absence of soot cake and
781 reduce filtration efficiency. Therefore, for the blended fuels, the design may need to be
782 optimised by changing the filter material porosity or/and the filter wall size to achieve a balance
783 between pressure drop and filtration efficiency. Moreover, soot produced from oxygenated fuels
784 presents a less-ordered nanostructure. This seems to be highly linked to higher oxidation
785 reactivity. It is worth noting that studies on the chemical properties of soot particulates from
786 oxygenated fuels are limited. Further research is necessary to clarify the impact of adding
787 oxygenated fuels on soot chemical features, such as surface functional groups.

Table 3. Summary of physicochemical features of soot particles from oxygenated fuels.

References	Engine type	Test fuel	Analytic methods	Main results
<u>Luo, et al. [39]</u>	wall-guided, 2.0 L, 9.2:1	Gasoline, E10, E20	TGA, GC-MS	Ethanol enhances soot oxidation reactivity and increases VOF content. Blended fuels reduce concentrations of most PAHs, except those with small aromatic rings.
<u>Wang, et al. [83]</u>	spray-guided, 0.565L, 11.5:1	Gasoline, E25, DMF, E100	TGA	Ranking of soot reactivity: E100 > DMF > E20 > Gasoline.
<u>Lee, et al. [141]</u>	-, 0.549 L, 11.97:1	Gasoline, E20	TEM	E20 has a negligible effect on the mean primary particle diameter. At an SOI of 250 °CA bTDC, primary particles from both fuels exhibit graphitic structures, whereas at an SOI of 310 °CA bTDC, the E20 particle shows a more amorphous structure.
<u>Gao, et al. [52]</u>	wall-guided, 0.5 L, 10.5:1	Gasoline, E10, E35, E60	TEM	Increasing ethanol concentration decreases the sizes of agglomerates and primary particles, and stretches the agglomerate structures. As the ethanol ratio increases, more amorphous nanostructures are observed.
<u>Sharma, et al. [85]</u>	wall-guided, 0.5 L, 10.5:1	Gasoline, E15, M15	TEM, STEM-EDS	At part load, d_p reduces as the fuel oxygen content increases. At full load, d_p increases as the fuel oxygen content increases. Insignificant difference in agglomerates was observed between the particles from various test fuels. The concentration of trace metals is lower in E15 than gasoline and M15.
<u>Hergueta, et al. [139]</u>	air-guided, 2.0 L, 10:1	Gasoline, E25, B33	TEM	E25 soot exhibits the smallest and the most chain-like agglomerates with the smallest primary particle size. Gasoline soot presents the largest and the most compact agglomerates with the largest primary particle size. The difference in soot morphology between gasoline and B33 is relatively small.
<u>Sharma, et al. [142]</u>	wall-guided, 0.5 L, 10.5:1	Gasoline, E15, B15, M15	ICP-OES, RS, FTIR, HRTEM	Particulates from gasohols exhibit a more amorphous structure than those from gasoline. Particulates from gasohols have lower trace metal concentrations than those from gasoline.

<u>Sharma, et al.</u> <u>[140]</u>	wall-guided, 0.5 L, 10.5:1	Gasoline, B15	TEM	Compared to gasoline, B15 soot exhibits a smaller d_p at partial load and a larger d_p at full load. Soot agglomerates present chain-like structures, and there is insignificant difference between both fuels.
<u>Karavalakis, et al.</u> <u>[34]</u>	wall-guided, 5.3 L, -	E10, E51, E83, iBu55	TOC	Increasing ethanol content reduces the EC fraction significantly. Regarding iBu55, the EC concentration is almost the same as the OC fraction.
<u>Chan, et al.</u> <u>[136]</u>	air-guided, 2.0 L, 10:1	Gasoline, DMC8	TGA, TEM, RS	DMC8 soot has a larger proportion of VOF and a higher oxidation reactivity. DMC8 soot shows smaller primary particle and agglomerates sizes than gasoline soot and a more disordered structure.

790 5 Conclusions and outlook

791 GDI particulates have become an emerging environmental concern with the penetration of
792 GDI engine-powered vehicles in recent years. GPFs are essential for GDI engines to meet the
793 increasingly stringent particulate regulations. A better understanding of PM characteristics is
794 crucial to develop smart GPF management systems and extend the catalyst's life of GPF. In
795 addition, the physical and chemical characteristics of soot particles can provide useful
796 information on PM formation and the assessment of exhaust toxicity and environmental impacts.
797 Therefore, this review attempted to comprehensively describe the physicochemical features of
798 soot particulates from GDI engines.

799 In-flame soot samples collected on the exhaust valve side or generated from the piston-
800 wetting condition have a longer residence time and exhibit a more carbonised structure.
801 However, compared to in-flame soot, GDI exhaust soot presents a more stable and graphitic
802 nanostructure, which makes it harder to oxidise. In addition to carbonaceous soot, ash particles,
803 which are mainly originated from the in-cylinder consumed lubricating oil, are observed in GDI
804 exhaust particulate samples. Accordingly, in addition to the main components (e.g. C and O),
805 element components including Ca, P, Na, Si and Zn are detected in soot particles from GDI
806 vehicles. Compared to diesel exhaust particles, GDI soot contains more inorganic and metallic
807 elements. Additionally, GDI soot is usually smaller, less ordered and more reactive than diesel
808 soot. However, further researches are needed to present a complete picture of soot features from
809 in-cylinder to the atmosphere, including: i) revealing instantaneous features of in-cylinder soot
810 as a function of crank angle degrees; and ii) clarifying the differences in soot features from the
811 exhaust valve into the atmosphere.

812 The effects of engine operating parameters (i.e. engine speed and load, equivalence ratio,
813 fuel injection pressure and timing, and EGR) and aftertreatments (i.e. TWC and GPF) are
814 systematically discussed. However, no consensus has been reached which could be attributed
815 to the complexity of analytic techniques, the variations in engine specifications and fuel
816 properties, and the heterogeneity of soot samples. The effects of engine speed and load on GDI

817 soot features are smaller than on diesel soot. Higher injection pressures promote soot oxidation,
818 and there is an optimal SOI to enhance soot oxidation reactivity. Both EGR and TWC have a
819 minor effect on soot reactivity. Although GPF significantly reduces PAH emissions, it has a
820 minor effect on soot morphology. Future studies need to evaluate more influencing factors, such
821 as spark timing, intake boost, multiple injections and injector deposits.

822 **Mixing** with oxygenated fuels enhances the soot oxidation reactivity and benefits the
823 regeneration of GPFs. In addition, oxygenated fuel-derived soot particles have smaller
824 agglomerates with smaller primaries, which implies that the design of the filter material porosity
825 and wall size must be optimised to maintain the filtration efficiency. Regarding the fuel
826 components and properties, more work is required to investigate the effect of the physical and
827 chemical properties of gasoline fuel on soot features.

828

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