

28 **ABSTRACT**

29 Seagrasses are important marine foundation species, which are presently threatened
30 by coastal development and global change worldwide. The molecular mechanisms
31 that drive seagrass responses to anthropogenic stresses, including elevated levels of
32 nutrients such as ammonium, remains poorly understood. Despite the evidence that
33 seagrasses can assimilate ammonium by using glutamine synthetase (GS) / glutamate
34 synthase (glutamine-oxoglutarate amidotransferase or GOGAT) cycle, the regulation
35 of this fundamental metabolic pathway has never been studied at the gene expression
36 level in seagrasses so far. Here, we combine (i) reverse transcription quantitative real-
37 time PCR (RT-qPCR) to measure expression of key genes involved in the
38 GS/GOGAT cycle, and (ii) stable isotope labelling and mass spectrometry to
39 investigate ¹⁵N-ammonium assimilation in the widespread Australian species *Zostera*
40 *muelleri* *subsp. capricorni* (*Z. muelleri*). We demonstrate that exposure to a pulse of
41 ammonium in seawater can induce changes in GS gene expression of *Z. muelleri*, and
42 further correlate these changes in gene expression with ¹⁵N-ammonium uptake rate in
43 above- and below-ground tissue.

44 **1. INTRODUCTION**

45 Seagrasses are a group of vascular flowering plants whose evolutionary origins date
46 back to the Cretaceous, approximately 100 million years ago (Larkum et al., 2006).
47 They are composed of a true root-rhizome system (i.e. below-ground tissue)
48 anchoring them to the seabed and a canopy of leaves (i.e. above-ground tissue)
49 responsible for oxygen exchange, nutrient uptake and light capture (Larkum et al.,
50 2006, Zimmerman et al., 1987) (Figure 1A). Worldwide, there are approximately 72
51 seagrass species in 12 genera (Short et al., 2011). Seagrasses play an important role in
52 coastal ecosystems by (i) providing food and shelter for commercially important fish
53 (Beck et al., 2001), (ii) enhancing sediment accretion (Koch et al., 2013), (iii)
54 protecting the coastline from wave energy (Larkum et al., 2006) and (iv) sequestering
55 carbon, being responsible for 10 to 18% of the carbon accumulation in the ocean
56 (Macreadie et al., 2014a).

57

58 Despite their global importance, seagrasses are at risk worldwide. Significant losses
59 of seagrass meadows have been reported (Calleja et al., 2007, Cambridge and
60 McComb, 1984, Duarte, 2002) and Orth et al. (2006) declared a “global crisis for
61 seagrass meadows”. Previous studies have established a causal link between large die-
62 off events and environmental stresses, having been triggered by climate change and
63 anthropogenic impacts, e.g. ocean warming, increased frequency of heavy rainfall and
64 severe weather as well as agricultural/urban runoff, eutrophication, boat mooring and
65 dredging (Orth et al., 2006, Ralph et al., 2007, York et al., 2015). Eutrophication and
66 enrichment of nutrients, especially nitrogen, in coastal waters have been shown to be
67 one of the major factors of seagrass decline and have thus received increasing
68 attention within the past decades (e.g. Short and McRoy, 1984, Touchette and

69 Burkholder, 2007, Alexandre et al., 2010, Apostolaki et al., 2012). A range of
70 mechanisms have been proposed to explain reduced seagrass survival under
71 anthropogenic nutrient enrichment including algal turbidity (Morris and Tomasko,
72 1993), i.e. stimulation of algal overgrowth, which shades the underlying seagrass,
73 limiting light availability during daytime and reducing O₂ supply in darkness, both
74 leading to diminished plant fitness (Kemp et al., 1983, Brodersen et al., 2015,
75 Rasmussen et al., 2012). Direct nutrient toxicity effects on seagrass growth and
76 survival have also been reported (Burkholder et al., 2007), but the underlying
77 mechanisms have remained unclear.

78

79 Seagrasses are believed to have evolved in nitrogen-limited waters, and these plants
80 may have maximized nitrogen uptake/assimilation processes during infrequent
81 nutrients pulses (Burkholder et al., 1994). Previous research has shown that uptake of
82 inorganic nitrogen from the water column or the sediment pore water can supply up to
83 73% of seagrass annual need (Pedersen and Borum, 1993) (Figure 1). This nitrogen
84 supply is critical to balance the losses of nitrogen due to (i) denitrification in the
85 rhizosphere and (iii) increase in meadow size and density (Pedersen and Borum,
86 1993). While the nutrients source (i.e. sediment pore-water or water column) and
87 tissue (i.e. above or below-ground tissue) responsible for nutrients assimilation in
88 seagrass are commonly considered to vary depending on availability and
89 concentration, sediment pore water is often considered to be the primary source of
90 nitrogen (see: Touchette and Burkholder, 2000 for review).

91

92 The physiology for ammonium assimilation to amino acids by seagrass tissues is
93 complex involving the following two steps: (i) transport from seawater into the cell

94 and (ii) assimilation into amino acids. Ammonium transport into the cell is generally
95 view as a passive movement through membrane channels but few studies have
96 suggested the presence of a feedback inhibition mechanism indicative of a more
97 active transport for certain seagrass species such as *Zostera marina* (Touchette and
98 Burkholder, 2000, Short and McRoy, 1984). Among the molecular mechanisms
99 driving nutrient assimilation into amino acids, the glutamine synthetase (GS) /
100 glutamate synthase (glutamine-oxoglutarate amidotransferase or GOGAT) pathway of
101 ammonium incorporation (Goodwin and Mercer, 1983) is an important and
102 fundamental component of ammonium assimilation in seagrass (see: Touchette and
103 Burkholder, 2000 for review). This metabolic pathway involves the incorporation of
104 ammonium into glutamine by GS and the subsequent regeneration of glutamate from
105 glutamine and α -ketoglutarate by GOGAT (Goodwin and Mercer, 1983) (Figure 1B).
106 Indeed, several studies focusing on this metabolic pathway in seagrass have
107 demonstrated the influence of the surrounding environment on the enzymatic activity
108 of GS and its involvement in ammonium assimilation (Kraemer et al., 1997,
109 Alexandre et al., 2010, see: Touchette and Burkholder, 2000 for review). However,
110 the regulation of this fundamental metabolic pathway has never been studied at the
111 gene expression level so far in seagrass.

112

113 Our ability to predict the impacts of anthropogenic stresses such as nutrients
114 enrichment on seagrass meadows is dependent on how well we understand the
115 mechanisms that drive the cellular response to nutrients enrichment. Understanding
116 these underlying mechanisms also has importance to policy and management
117 responses, especially when it comes to understanding why some seagrass meadows
118 appear to vary in their tolerance to nutrient loading. Seagrasses respond to changing

119 environments by regulating molecular pathways to prevent or abate physiological
120 damage. These molecular events precede population-level changes and could be
121 useful biomarkers if they can be linked to specific physiological or ecological events
122 (Macreadie et al., 2014b). Thus, seagrass researchers have recently taken advantage of
123 the power of genomics and transcriptomics to address questions about the molecular
124 mechanisms of ecological responses to environmental perturbations (Franssen et al.,
125 2011, Franssen et al., 2014, Dattolo et al., 2014, Serra et al., 2012, Lauritano et al.,
126 2015, Salo et al., 2015, Brakel et al., 2014). Following the recent publication of
127 *Zostera marina* genome (Olsen et al., 2016), the seagrass community is now eagerly
128 awaiting the progress that will be promoted by integrating molecular biology and gene
129 expression into the spectrum of physiological and metabolic changes involved in the
130 response of seagrass to environmental stress.

131

132 In the present study, we conducted such a multi-parameter analysis by monitoring (i)
133 the initiation of the GS/GOGAT cycle by differential expression of GS and GOGAT
134 transcripts using RT-qPCR and (ii) the assimilation of ammonium by measuring the
135 incorporation of ^{15}N using mass spectrometry in the above- and below-ground tissue
136 of the seagrass *Zostera muelleri*. We hypothesised that seawater enriched in
137 ammonium would affect levels of expression for these key genes in nitrogen
138 assimilation in *Z. muelleri*. Moreover, for the first time, we correlated changes in gene
139 expression to the overall uptake rate of nutrients within the above- and below-ground
140 tissue.

141 **2. MATERIAL AND METHODS**

142 *2.1. Collection and maintenance of seagrass*

143 *Zostera muelleri* plants with intact sediment core of 5 cm in diameter and 10 cm in
144 depth were collected at Pelican Banks, Port Curtis (Queensland, Australia, ambient
145 salinity: 32 and water temperature: 24°C) and transported to the aquarium facility at
146 University of Technology Sydney (UTS), where they were acclimatised for two
147 months under an irradiance of 150 $\mu\text{mol photons m}^{-2} \text{ s}^{-1}$ over a 12:12 light:dark cycle.
148 Ambient water temperature was maintained at 24°C by heater-chillers; practical
149 salinity *sensu* (Lewis, 1980) was kept between 31.0 - 33.0 and adjusted when
150 necessary using deionized water / seawater. During the acclimation period, nutrients
151 concentration was monitored weekly in the aquaria and remained low (NO_2^- , <1 μM ;
152 NO_3^- , <5 μM ; NH_4^+ , <1 μM). Further, seagrass leaves were cleaned using thumb and
153 forefinger with gloves gently, and water was changed on a fortnightly basis to
154 minimise epiphyte growth. After the initial acclimatisation period, the labelling
155 experiment (as described below) was conducted to examine the effect of elevated
156 nutrients on both ammonium assimilation and gene expression in *Z. muelleri*.

157

158 *2.2. Labelling experiment*

159 Plants were exposed to two experimental treatments (i.e. ammonium enriched and
160 control treatments) across a total of six tanks (three 4 litres tank replicates per
161 treatment with a total of 8 shoots per tank; closed water system yet continuously
162 stirred using one powerhead pump for each tank). During the 36 h of the experiment,
163 the conditions were as follows: practical salinity at 32.0, pH 8.1, irradiance of 150
164 $\mu\text{mol photons m}^{-2} \text{ s}^{-1}$ over a 12:12 light:dark cycle and 2mM of dissolved inorganic
165 carbon. Three replicate samples of *Z. muelleri* (one plant for each of the three tank

166 replicates) were collected for each time point (0, 6, 12 and 36 h) and for each
167 treatment (ammonium enriched and control treatments) for further isotopic analysis
168 and gene expression analysis using RT-qPCR.

169 For the ammonium-enriched treatment, seawater used during acclimation was emptied
170 and replaced by dispensing artificial seawater enriched in ^{15}N -ammonium in the
171 aquarium tanks. The artificial seawater was prepared freshly using ultrapure water (no
172 detectable Dissolved Inorganic Nitrogen, recipe adapted from (Harrison et al., 1980)),
173 initially free of ammonium, which was then enriched in ^{15}N -labelled ammonium by
174 adding $^{15}\text{NH}_4\text{Cl}$ powder to a final concentration of 20 μM . This concentration is
175 above the natural concentration of ammonium in seawater and may represent elevated
176 environmental levels of ammonium caused by nutrients enrichments in coastal waters
177 (Koop et al. 2001). For example, a previous study reported that ammonium
178 concentration varied between <0.5 and 40 mM in northern Port Phillip Bay (VIC,
179 Australia) during periods of low runoff (January 1995) and high runoff (September
180 1993) respectively (Longmore et al., 2000). The $^{15}\text{NH}_4\text{Cl}$ powder had a ^{15}N isotopic
181 abundance of 98% (commercially available from Sigma, Castle Hill, NSW,
182 Australia).

183 For the control treatment, seawater used during acclimation was emptied and replaced
184 by dispensing artificial seawater, which was prepared freshly using ultrapure water as
185 described above (no detectable Dissolved Inorganic Nitrogen, recipe adapted from
186 Harrison et al., 1980), but this time free of ^{15}N - ammonium. It is important to note
187 that this experimental design could have induced nitrogen limitation in the control
188 treated plants. However, because our labelling experiment was focused on rapid
189 physiological response of seagrass (i.e. 36 hours experiment), this experimental
190 design was used to ensure that (i) the artificial seawater used for the control treatment

191 was initially nutrients-free and (ii) ¹⁵N-labelled ammonium was the only source of
192 nutrients in the ammonium-enriched treatment. For each plant sample collected,
193 above- and below-ground tissues were separated using scissors and resulting samples
194 were stored separately at -80°C for further isotopic and RT-qPCR analysis.

195

196 2.3. Nitrogen content and uptake rate

197 Prior to nitrogen content and ¹⁵N/¹⁴N analysis, samples of above- and below-ground
198 seagrass tissue were freeze-dried and weighed. Nitrogen content (%) and ¹⁵N/¹⁴N
199 ($\delta^{15}\text{N}$, in ‰) were then analyzed in triplicate using an elemental analyzer interfaced to
200 a continuous flow isotope ratio mass spectrometer at the Stable Isotope Facility,
201 Department of Plant Sciences (UC Davis, California, US). The Increase in ¹⁵N/¹⁴N
202 ratio for each sample was calculated in ‰ as follows:

$$203 \quad \delta^{15}\text{N} = (N_{mes}/N_{nat}) - 1 \times 10^3$$

204 where,

205 N_{mes} : ¹⁵N/¹⁴N measured in labelled samples;

206 N_{nat} : natural abundance of ¹⁵N/¹⁴N measured in unlabelled samples.

207 The nitrogen uptake rate per mg of tissue (ρ) was then expressed in ng N h⁻¹ mg⁻¹ and
208 was calculated by normalizing the ¹⁵N-incorporation measured (i) to the average
209 nitrogen content (% of dry mass) of the tissue and (ii) to the time of incubation
210 according to the equation previously described (Dugdale and Wilkerson, 1986):

$$211 \quad \rho = ((N_{mes} - N_{nat}) / ((N_{enr} - N_{nat}) \times T_{inc})) \times N_{content} \times 10^3$$

212 where,

213 N_{mes} : ¹⁵N/¹⁴N measured in labelled samples;

214 N_{nat} : natural abundance of ¹⁵N/¹⁴N measured in unlabelled samples.

215 Nenr: ¹⁵N-enrichment of the incubation medium
216 Tinc: incubation time
217 Ncontent: average nitrogen content (%) measured by using combustion (950°C)
218 method.

219

220 *2.4. Quantitative Real-Time PCR (qPCR)*

221 The present study conforms to the Minimum Information for Publication of
222 Quantitative Real-Time PCR guidelines (Bustin et al., 2009). In this section, we
223 indicate the essential information, *sensu* (Bustin et al., 2009), required to allow
224 reliable interpretation of the corresponding qPCR results.

225

226 *2.5. Primer design*

227 In-depth analysis of *Zostera muelleri* Transcriptomics Database (Hayward et al, in
228 prep), revealed transcripts encoding proteins with high similarities to the domains of
229 GS and GOGAT proteins already identified in the seagrass *Zostera marina* (Olsen et
230 al., 2016). The sequences were blasted for isoforms
231 (<http://blast.ncbi.nlm.nih.gov/Blast.cgi>) and functional domains of two of these
232 sequences, coding for Glutamine synthetase (GS; KMZ65307) and Glutamate
233 synthase (GOGAT; KMZ61057) respectively, were used as a template to design
234 sequence-specific primers for RT-qPCR using the software, Primer3 0.4.0 (Rozen and
235 Skaletsky, 2000) (Source code available at <http://fokker.wi.mit.edu/primer3/>; Table
236 1).

237

238 *2.6. RNA extraction and cDNA synthesis*

239 Above-ground and below-ground samples (i.e. pieces from the second youngest leaf
240 and youngest rhizome internode) from control and ammonium-enriched treatments
241 were collected at 0, 6, 12, and 36 hours after incubation, snap-frozen in liquid
242 nitrogen and stored in -80°C freezer until further analysis ($n = 3$). Samples were then
243 ground into powder in liquid nitrogen using a mortar and pestle. RNA was then
244 extracted using the RNA mini kit according to manufacturer's instructions (PureLink
245 TM RNA Mini Kit, Life Technologies). Column purification was carried out to remove
246 gDNA using PureLink TM DNase (Life Technologies). The RNA quantity was
247 assessed using Qubit RNA HS assay kit (Invitrogen) and quality checks were
248 performed using Nanochip technology (Agilent 2100 Bioanalyzer). High-quality
249 RNA samples (integrity number >6) were used for further cDNA synthesis and RT-
250 qPCR analysis. The cDNA was generated using a High Capacity Reverse
251 Transcription kit (Applied Biosystems) in a 20 μ l reaction using 200 ng of total RNA
252 as template.

253

254 2.7. RT-qPCR

255 RT-qPCR gene expression quantification was performed in three technical replicates
256 using SYBR Green PCR Master Mix (Applied Biosystems) on a Step One Plus Real-
257 Time PCR System (Applied Biosystems). Amplification of 10 μ L reactions with 1:50
258 dilutions of cDNA from control and treated seagrasses and 1 μ M of each specific
259 primers were placed in 96-well optical plates with the following PCR conditions:
260 incubation at 95°C for 10 min, then 50 cycles of 95°C, 60°C, 68°C for 30 s each
261 followed by 95°C for 15 s, 60°C for 1 min and 95°C for 15 s. The RT-qPCR
262 efficiency for each gene and each treatment was determined from a cDNA dilution
263 gradient of 243, 81, 27, 9 and 3 ng and a linear regression model (Pfaffl, 2001). The

264 corresponding RT-qPCR efficiencies were calculated according to the equation
265 described by Radonic et al (Radonic et al., 2004):

$$266 \quad PCR \text{ efficiency} = (10^{\lceil -\frac{1}{slope} \rceil} - 1) \times 100$$

267 All of the RT-qPCR efficiencies obtained for the different primers were between 89-
268 100% (Table 1; Supplementary Table S1), with a calibration coefficient >0.95
269 similarly to previous studies (Winters et al., 2011, Bergmann et al., 2010). A *no*
270 *template control*, as well as a *no reverse transcription control* was generated for each
271 gene and each treatment to ensure that the PCR reactions were free of DNA
272 contamination. Replicate variability of the CT values between the 3 technical
273 replicates, which were run on the same plate, was examined for each sample-gene
274 combination. Repeatability of the assay between the technical replicates was
275 consistent across the different genes with the replicate variability falling within the set
276 limit of <0.5 cycles for all the sample-gene combinations tested.

277

278 2.8. Data acquisition

279 Data from RT-qPCR was analysed using the Step One Plus Software (Ver. 2.3;
280 Applied Biosystems). Expression levels were determined as the number of cycles
281 needed for the amplification to reach a fixed threshold in the exponential phase of the
282 RT-qPCR reaction (Walker, 2002). The cycle threshold (CT) was set at 0.03 for all
283 genes. To validate changes in target genes expression, CT were imported into the
284 qbase+ software package (Biogazelle). Expression data for target genes was first
285 normalized against the best combination of 3 reference genes as previously described
286 (Adenosylhomocysteinase, AHCY; Glyceraldehyde 3-phosphate dehydrogenase,
287 GADPH; and Translation initiation factor 1 subunit beta, Elof1; Supplementary Table

288 S1, Supplementary Figure S1) (Schliep et al., 2015) using GeNorm algorithm within
289 qbase+ software package (Biogazelle) and then transformed into quantities using
290 maximum efficiency of 1.00 (or 100%) to obtain Calibrated Normalized Relative
291 Quantities (CNRQ) (Hellemans et al., 2007).

292

293 *2.9. Statistical analysis*

294 Statistical analyses were done using Repeated Measures Analyses of Variance with
295 PERMANOVA+ software in PRIMER v6 (Anderson et al., 2008). The analyses
296 tested the null hypothesis that there no difference in these data among ammonium
297 enriched and control treatments and/or among sampling time (0, 6, 12 and 36 hours).
298 Univariate analyses of variance of ¹⁵N-enrichment and gene expression data were
299 undertaken based on Euclidean dissimilarity matrices (Anderson et al., 2008) using
300 the factors described in Supplementary Table S2. Where significant factors' effects
301 were detected, post hoc permutational pair-wise comparison using PERMANOVA+
302 were undertaken to identify the levels of factors in which differences occurred. P-
303 values were based on 4999 permutations and results were considered significant at P
304 < 0.05. Differential gene expression data are presented on a log scale relative to
305 control. Throughout the paper, values given are mean±SE (*n*=3 plants). Because of
306 biomass limitation, we used a random design for the factor aquaria and only one plant
307 was collected for each of the aquaria replicate (1, 2 and 3) for each time point (0, 6,
308 12 and 36 h) and for each treatment (ammonium enriched and control treatments).
309 Therefore, we could not add aquaria as a factor in our statistical analysis (i.e. *n*=1
310 plant per aquaria per time point per treatment but *n*=3 plants per time point per
311 treatment in total). While adding aquaria as an additional factor and nesting it under
312 treatment would have potentially increased the statistical power as well as the testing

313 of specific temporal variation, our experimental design allowed to testing for
314 treatment-specific temporal variation in ^{15}N enrichment and gene expression.

315 3. RESULTS AND DISCUSSION

316 3.1. Nitrogen uptake in above- and below-ground tissue of *Zostera muelleri*

317 The dynamics of ^{15}N -ammonium incorporation was followed and quantified in the
318 above- and below-ground tissue fractions of *Z. muelleri* plants by measuring the
319 $^{15}\text{N}/^{14}\text{N}$ ratio after tissue separation. The $^{15}\text{N}/^{14}\text{N}$ mass spectrometry data indicated
320 rapid uptake of ^{15}N -ammonium by *Z. muelleri* plants (Figure 2). The above-ground
321 tissue displayed a substantial ^{15}N -enrichment after incubation in enriched seawater
322 with a $11093 \pm 1412 \text{ ‰}$ ^{15}N -enrichment within the first 6 h of labelling experiment
323 (pair-wise comparison between control and treated seagrass plants at $t=6$ h, $p=0.003$,
324 Supplementary Table S2B), followed by further increase in the next 6 h of the
325 experiment ($14457 \pm 438 \text{ ‰}$ ^{15}N -enrichment at $t=12$ h, pair-wise comparison between
326 control and treated seagrass plants, $p<0.001$, Supplementary Table S2B) and
327 accompanied by progressive increase from 12 h onward ($21960 \pm 5868 \text{ ‰}$ ^{15}N
328 enrichment at $t=36$ h, pair-wise comparison between control and treated seagrass
329 plants, $p= 0.016$, Figure 2A, Supplementary Table S2B). In contrast, the ^{15}N -labeling
330 of the below-ground tissue remained substantially weaker during the entire
331 experiment (Figure 2B), with an average ^{15}N -enrichment 3 to 4 times weaker than the
332 one observed in the above-ground tissue. The below-ground tissue showed a
333 significant increase in $^{15}\text{N}/^{14}\text{N}$ ratio in response to enriched seawater with a 2582
334 $\pm 636 \text{ ‰}$ ^{15}N -enrichment within the first 6 h of labelling experiment (pair-wise
335 comparison between control and treated seagrass plants at $t=6$ h, $p= 0.017$,
336 Supplementary Table S2D), followed by further increase ($4935 \pm 1700 \text{ ‰}$ ^{15}N -
337 enrichment at $t=12$ h ; pair-wise comparison between control and treated seagrass
338 plants, $p= 0.021$, Supplementary Table S2D, Figure 2B). Because ^{15}N -ammonium was
339 added in the water column, the slight delay observed in ^{15}N -enrichment in below-

340 ground tissue compared to the above-ground tissue is likely to reflect the time needed
341 for (i) basipetal translocation of ^{15}N -compounds (Zimmerman and Alberte, 1996)
342 and/or (ii) ^{15}N -ammonium diffusion from the water column into the sediment pore
343 water directly in contact with seagrass below-ground tissue.

344 By normalizing the ^{15}N -incorporation to the average nitrogen content of the above-
345 and below-ground fractions and to the time of incubation following the equation of
346 (Dugdale and Wilkerson, 1986), it was possible to estimate the ^{15}N taken up per hour
347 by the above- and below-ground tissue, respectively. In the first 12 hours of the
348 experiment, the above-ground tissue took up nitrogen with a specific rate of 267 ± 27
349 $\text{ngN h}^{-1} \text{mg}^{-1}$. This uptake rate is in the range of values found previously in the leaves
350 of *Z. marina* at the same initial ammonium concentration (Thursby and Harlin, 1982),
351 confirming the capacity of *Z. muelleri* above-ground tissue to uptake nitrogen rapidly
352 in response to a pulse of nutrients in the water column. Comparison between above-
353 and below-ground tissue uptake rates further allows establishing the relative
354 contribution of above- and below-ground tissue to the ammonium uptake capabilities.
355 In this respect, our data indicates that the above-ground tissue incorporated 8 times
356 more nitrogen than below-ground tissue during the first 12 h of the labelling
357 experiment (specific uptake rates of: $267 \pm 27 \text{ ngN h}^{-1} \text{mg}^{-1}$, $n=6$ for above-ground
358 tissue; $33 \pm 11 \text{ ngN h}^{-1} \text{mg}^{-1}$, $n=6$ for below-ground tissue; Figure 2C). This particular
359 result indicates a significant heterogeneity in the access and therefore ability of above-
360 and below-ground tissue to uptake nitrogen from a pulse of nutrients in the water
361 column; the above-ground fraction being the most active, but not the only site of
362 assimilation. It is important to note that these uptake rate calculations assume that the
363 ^{15}N isotope fraction of the ammonium was the same in the water above and below the
364 sediment. Therefore, the uptake reported for the below-ground tissue is likely to be

365 underestimated, but the uptake rate for above-ground tissue remains valid. It is also
366 possible that nitrogen-levels and fauna present in the sediment might have had an
367 impact on our data but we did not measure these parameters as our labelling
368 experiment was focused on the physiological responses of seagrass to a short pulse of
369 nutrients (i.e. 36 hours experiment after 2 months acclimation under laboratory
370 conditions). Further studies focussing on longer term response should therefore
371 investigate the effects of different nitrogen-levels and fauna present in the sediment
372 on ammonium assimilation in *Z. muelleri*. Because the pulse of ¹⁵N-ammonium was
373 experimentally introduced into the water column, the difference in the intensity and
374 the kinetics on which the ¹⁵N-label reached the above- and the below-ground tissue
375 suggest that the ¹⁵N incorporated by below-ground tissue derives either (i) from rapid
376 assimilation by the leaves and then a delayed basipetal translocation of nitrogenous
377 compounds (Kaldy et al., 2013, Thursby and Harlin, 1982) or (ii) from the delayed
378 diffusion of ¹⁵N-ammonium into sediment pore water surrounding the below-ground
379 tissue.

380 *3.2. Expression of key assimilation genes in above- and below-ground tissue of*
381 *Zostera muelleri*

382 In parallel to ¹⁵N-ammonium assimilation, we were able to detect changes in the
383 expression of transcripts coding for GOGAT and GS in the different tissue fraction of
384 *Z. muelleri* over the course of the experiment using RT-qPCR (Figure 3). Among the
385 transcripts targeted in this study, GS transcripts was significantly upregulated in
386 above- and below-ground tissue of *Z. muelleri* in response to enriched seawater with a
387 significant effect of the incubation time (above-ground tissue: PERMANOVA
388 analysis, $p < 0.001$, Figure 3A, Supplementary Table S2E; below-ground tissue:
389 PERMANOVA analysis, $p = 0.042$, Figure 3C, Supplementary Table S2H).

390 Ammonium enriched seawater induced different temporal dynamics of GS transcripts
391 regulation in the above- and below-ground tissue. Indeed, the earliest significant gene
392 regulation occurred within the above-ground tissue, with GS being significantly
393 upregulated from 6 h onward (treatment: 3.3 up-regulation; pair-wise comparison,
394 $p=0.001$, Figure 3A, Supplementary Table S2F) while in the below-ground tissue GS
395 was significantly upregulated only after 12 hours (treatment: 9.2 up-regulation; pair-
396 wise comparison, $p= 0.012$, Figure 3C, Supplementary Table S2I). Furthermore, in
397 the above-ground tissue, this increase in expression level of GS transcript remained
398 stable and significant from 6 hours onwards (Figure 3A) while in the below-ground
399 tissue, the upregulation of GS transcript was only significant at 12 hours and
400 decreased back to control level after that (Figure 3C). In the light of nitrogen uptake
401 data, these gene expression data suggest that the delay observed in ^{15}N incorporation
402 in below-ground tissue could be linked (i) to the time needed for ^{15}N -ammonium to
403 diffuse into the sediment pore water directly in contact with seagrass and/or (ii)
404 basipetal translocation. Because the level of expression GS transcript declined
405 concomitantly with the ^{15}N -enrichment in the below-ground tissue, our results further
406 suggest that during our experiment, very little of the ^{15}N -ammonium diffused into the
407 porewater, indicating that the second hypothesis is more likely. Although sediment
408 pore water is generally considered to be the primary source of ammonium for
409 seagrasses (Touchette and Burkholder, 2000), our results are supported by previous
410 studies indicating that ammonium uptake by below-ground tissues can be limited by
411 diffusion into sediment pore water (Short and McRoy, 1984, Stapel et al., 1996). It is
412 important to note that our experiment was focused on investigating rapid
413 physiological and molecular response of seagrass and represents a very unrealistic
414 case, since most sediments contain very high concentrations of ammonium (up to mM

415 levels). Further, the relative contribution of seagrass below-ground tissue to
416 ammonium assimilation is highly variable and might depend on (i) the presence of
417 microbial population associated with sediment or seagrass tissue and (ii) on the level
418 of ammonium in the water column and on the type of substrata the plants are growing
419 on. For instance, (Thursby and Harlin, 1982, Terrados and Williams, 1997) indicated
420 that maximum rates of root ammonium uptake were substantially diminished when
421 leaves were exposed to increased ammonium concentration, suggesting that under
422 these conditions, (i) most of nutrient absorption was occurring from the water column
423 into the leaves and (ii) basipetal translocation of nitrogenous products was higher
424 relative to acropetal translocation. Future experimental studies combining stable
425 isotope incubation, characterisation of associated microbial population and gene
426 expression over a range of environmental conditions should also include simultaneous
427 measurement of (i) N-levels in the water column and in the sediment pore-water as
428 well as (ii) bioturbation in order to determine more precisely the environmental
429 factors affecting nutrient assimilation and exchanges in *Z. muelleri* and the molecular
430 regulations involved in the different tissues.

431

432 As for land plants, GS and GOGAT are widely recognized as key enzymes in seagrass
433 nitrogen metabolism. GS drives the assimilation of ammonium into glutamine first,
434 which is then followed by addition of carbon skeletons by GOGAT to produce two
435 glutamate molecules. However, the diversity of isoforms present in seagrasses for
436 these key genes in nitrogen assimilation remains unknown and future studies
437 characterizing the diversity of GS isoforms and their localisation within the different
438 tissues of seagrass are needed to reveal specific functional roles and possible
439 adaptations.

440 In higher plants, GOGAT catalyses the production of glutamate from glutamine and
441 α -ketoglutarate (Lam et al., 1996). While an increasing trend was observed in the
442 expression level of this transcript in both tissues in response to seawater enriched in
443 ammonium, no statistical difference was observed (above-ground tissue:
444 PERMANOVA analysis, $p=0.204$, Figure 3B, Supplementary Table S2G; below-
445 ground tissue: PERMANOVA analysis, $p=0.249$, Figure 3D, Supplementary Table
446 S2G), suggesting either (i) that the variability in our RT-qPCR assay was too high to
447 allow detecting statistical difference in the expression level of this transcript or (ii)
448 that the role of GOGAT regulation in ammonium assimilation in *Z. muelleri* may
449 depends on other modifications (e.g. post-translational) that may not be followed by
450 differential gene expression.

451

452 *3.3. Expression of key assimilation genes as potential bioindicators of nutrients* 453 *levels*

454 Ammonium assimilation in seagrasses is complex and variable, the first step of this
455 process, transport of ammonium into the cell, being active for certain species and
456 population of seagrass, while passive for others (Touchette and Burkholder, 2000). As
457 a result, sensitivity to nutrients pulse can vary among different species and
458 populations of seagrass (Touchette and Burkholder, 2000) and ammonium
459 concentration in seawater might not be a good predictor for seagrass ammonium
460 assimilation concentration. Given that GS/GOGAT cycle is widely recognized as the
461 universal route for the second step of this process, i.e. incorporation of ammonium
462 into amino acids through glutamate, the regulation of key genes involved in this
463 pathway could conceivably be used as an indicator for exposure of different species
464 and populations of seagrass to a sudden pulse of ammonium. In this respect, by

465 combining nitrogen uptake with gene expression data from our labelling experiment
466 ($t=6, 12$ and 36 hours), we built a series of regressions in an attempt to predict the
467 relationship between the expression level of GS and GOGAT transcripts and the
468 uptake rate measured in the above- and below-ground tissue of *Z. muelleri*. Although
469 regression might not be the best way to describe these relationships because of our
470 experimental sampling design, this supplementary data can be useful to suggest
471 correlation between ^{15}N -enrichment and differential expression of certain genes
472 (Supplementary Table S3 and Supplementary Figure S2). Among the different
473 transcripts investigated in this study, the level of expression of GS transcript was the
474 most strongly correlated to ^{15}N -enrichment in the different tissue fractions of *Z.*
475 *muelleri* (Supplementary Figure S2A and C). The ^{15}N -ammonium uptake rate
476 explained more than 70% of the variation in GS transcripts in the above-ground tissue
477 (Supplementary Figure S2A) and more than 50% in the below-ground tissue
478 (Supplementary Figure S2C). Such molecular markers could be used (i) to
479 characterize the nutritional history of different populations and species of seagrass in
480 order to make informed management decisions about the likely impact of
481 anthropogenic nutrient enrichment and (ii) to monitor changes in nutritional
482 conditions of seagrass meadows rapidly (i.e. within 6 hours of the onset of nutrients
483 enrichment), before any epiphyte growth and/or morphological changes occur (2 to 6
484 weeks) (Murray et al., 1999). However, it is important to note that such markers will
485 be more complex to implement and interpret than a simple environmental assay for
486 ammonium concentration. Further, given the dynamic kinetic that occurred during our
487 study, further studies need to investigate these gene regulation on an hourly basis,
488 especially in the context of diurnal cycle.

489

490 With the future advances of digital qPCR, the use of these markers might not even
491 require the step of housekeeping genes selection that is normally needed to normalize
492 RT-qPCR data (Schliep et al., 2015, Zmienko et al., 2015). Consequently, as long as a
493 threshold database can be established and maintained for different species, our results
494 may have implications for the management of seagrass meadows. Indeed, the
495 regulations in the expression level of transcripts such as GS has potential to be
496 developed as integrative indicators of nutrient conditions and anthropogenic nutrient
497 enrichment. However, more work is clearly needed regarding the applicability of this
498 technology as a management tool to monitor seagrass condition adjacent to coastal
499 development and associated nutrient inputs, more particularly including (i)
500 experiments with various environmental conditions to determine if the relationship
501 between ammonium levels and the expression level of GS and GOGAT transcripts in
502 seagrass can be influenced by others factors, especially light, temperature and salinity
503 (Touchette and Burkholder, 2000), (ii) field collections and experiments with a
504 broader range of nutrients levels in order to establish more precisely this dose-
505 response relationship and (iii) experiments over longer period to assess if these target
506 genes can provide early warnings of epiphyte growth and if their level of expression
507 can integrate nutrient stress over a longer timeframe (i.e. remain altered after exposure
508 to the pulsed event). In this respect, our results suggest that this molecular toolkit may
509 be used as early warning of nutrient-stress only and may require a different set of
510 genes for longer term nutrient impact assessment as they are more likely to affect
511 epiphytes, thereby creating also an issue of competition for light. Further, because
512 seagrass species have shown considerable variation in nutritional response, their
513 acclimation or adaptation on differing nutrient levels could have potential effects on
514 gene expression of GS and GOGAT, the response of these molecular signatures

515 remains therefore to be examined in other seagrass species and compared across
516 geographic regions. Nevertheless, further research combining isotopic labelling
517 experiments with gene expression analysis should enable us to significantly advance
518 our knowledge of the metabolism and nutritional physiology that lie at the very heart
519 of seagrass physiology.

520

521 In conclusion, this is the first study to investigate in parallel the expression of key
522 genes involved in GS/GOGAT cycle and the assimilation of ammonium in the
523 seagrass *Z. muelleri* in response to a pulse of nutrients. The present data constitute the
524 first evidence that exposure of a seagrass to a pulse of ammonium in seawater can
525 induce fast change in GS gene expression and suggest potential physiological
526 importance of this gene regulation for nitrogen metabolism in the different tissues of
527 this seagrass. These results not only provide new insights into the nutritional
528 physiology of *Z. muelleri* but also have implications for the development of molecular
529 markers for reactive monitoring of *Z. muelleri* meadows in response to anthropogenic
530 nutrient enrichment.

531

532 **ACKNOWLEDGMENTS**

533 We would like to thank anonymous referees and editor for their very constructive
534 comments. We also wish to thank P. Brooks and the technical staff of UTS and C3 for
535 their assistance in tank maintenance.

536

537 **COMPETING INTERESTS**

538 The authors declare no competing interests.

539

540 **AUTHOR CONTRIBUTIONS**

541 Conceived and designed the experiments: MP and PR. Performed the experiments:
542 SS, AC and MP. Analysed the data: GS, SS, AC and MP. Contributed
543 reagents/materials/analysis tools: PR, PM and MR. Wrote the paper: MP with
544 comments and suggestions from all authors.

545

546 **FUNDING**

547 MP, MS, MR and PR acknowledge funding through the Gladstone Port Corporation.
548 MR and PR acknowledge the support of an Australian Research Council Linkage
549 Grant LP11020045. PM was supported by an Australian Research Council DECRA
550 Fellowship DE130101084. The authors also thank UTS and C3 for strategic research
551 support.

552

553

554 **Table 1**

555 Target genes investigated in *Zostera muelleri* by using RT-qPCR. Accession numbers, primers sequences, amplicon length, melting
 556 temperature, and RT-qPCR efficiency are indicated. GS: Glutamine synthetase; GOGAT: Glutamate synthase.

TABLES AND FIGURES LEGEND

Name	Accession number	Primer forward sequence	Primer reverse sequence	Length (bp)	Tm	Efficiency (%)
GS	KMZ65307	TGACCCTAAGCCA ATTCCAG	CTTCATACCCACCAGG TGCT	92	60.0	97
GOGAT	KMZ61057	TGTCAGCTGGACC AAAGATG	AGGGCCAAGAAATCC CATAC	112	59.8	89

557 **Figure 1**

558

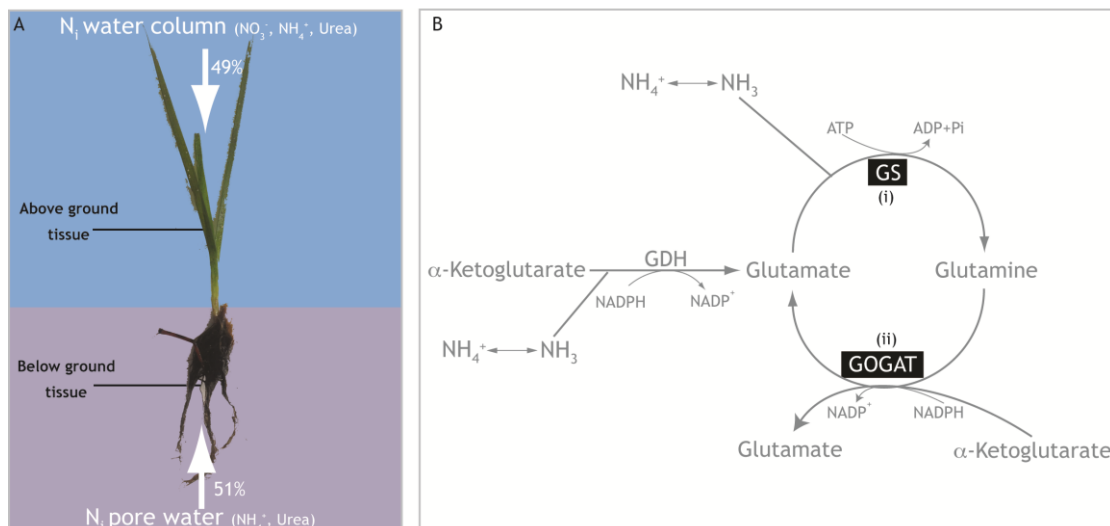
559 **Ammonium as a nitrogen source for seagrass.**

560 **A** Relative contributions of inorganic nitrogen (N_i : nitrate, NO_3^- ; ammonium, NH_4^+ ;
561 and urea) dissolved in water column (49%) and in porewater (51%), to the annual
562 external nitrogen requirements of seagrass (Pedersen and Borum, 1993).

563 **B** Reactions involved in ammonium uptake and assimilation in seagrass cells.
564 Ammonium enters the GS / GOGAT cycle, where (i) glutamate is aminated by the
565 enzyme glutamine synthetase (GS) to synthesize glutamine, which is then (ii)
566 transaminated by glutamate synthase (GOGAT) with the addition of carbon skeletons
567 (α -ketoglutarate) to finally produce two glutamate molecules. One of these glutamate
568 molecules can be recycled through the GS / GOGAT pathway, whereas the other can
569 be used to form more complex amino acids (Touchette and Burkholder, 2000). Black
570 boxes indicate key enzymes targeted in this study with coding genes identified in *Z.*
571 *muelleri* transcriptome (Hayward et al, in prep) and in *Z. marina* genome (Olsen et al.,
572 2016) and corresponding expression levels measured by RT-qPCR assay.

573

574



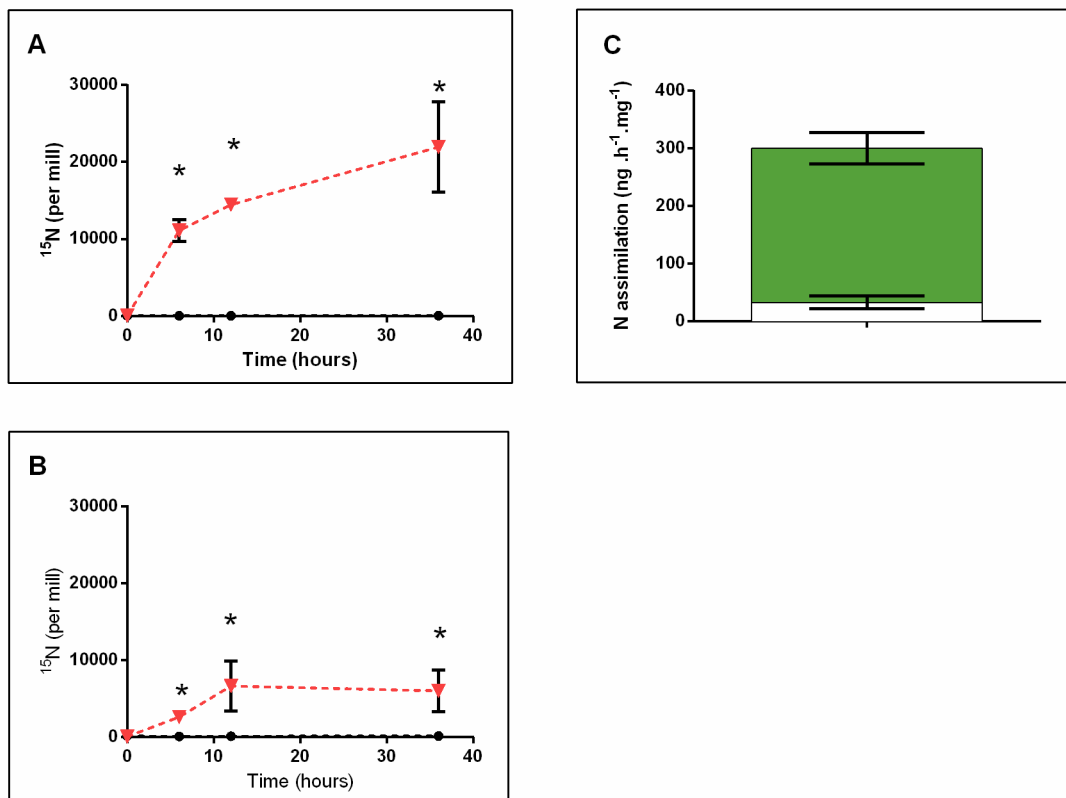
575

576
577
578
579
580
581
582
583
584
585
586
587
588
589
590
591
592
593

Figure 2

Quantification of nitrogen uptake in above- and below-ground tissue of *Zostera muelleri* in response to seawater enriched in ^{15}N -ammonium.

Time series of ^{15}N -enrichment in above- (A) and below-ground (B) tissue of *Zostera muelleri* for the duration of the experiment (36 hours). *Significant difference (pair-wise comparison, $P < 0.05$) between control treatment (black circle; no ^{15}N -enrichment in seagrass tissue above the natural fluctuations of the $^{15}\text{N}/^{14}\text{N}$ ratio, 0.00385 ± 0.00015) and seawater enriched in ^{15}N -ammonium (red triangle; average ^{15}N -enrichment in $^{15}\text{N}/^{14}\text{N}$ ratio in seagrass tissue, 0.04009 ± 0.00649). Error bars represent \pm SEM (3 plants). (C) Tissue specific nitrogen assimilatory capacity in *Zostera muelleri*. The total nitrogen uptake rate was calculated in above- (green) and below (white) ground tissue by normalizing the ^{15}N -incorporation to the average nitrogen content (% of dry mass) of *Z. muelleri* tissue and to the time of incubation according to the equation of Dugdale and Wilkerson (Dugdale and Wilkerson, 1986). The boxes indicate the mean value \pm SEM (6 plants).



594
595
596
597

598 **Figure 3**

599

600 **Ammonium-mediated expression of key assimilation genes in above- and below-**
601 **ground tissue of *Zostera muelleri***

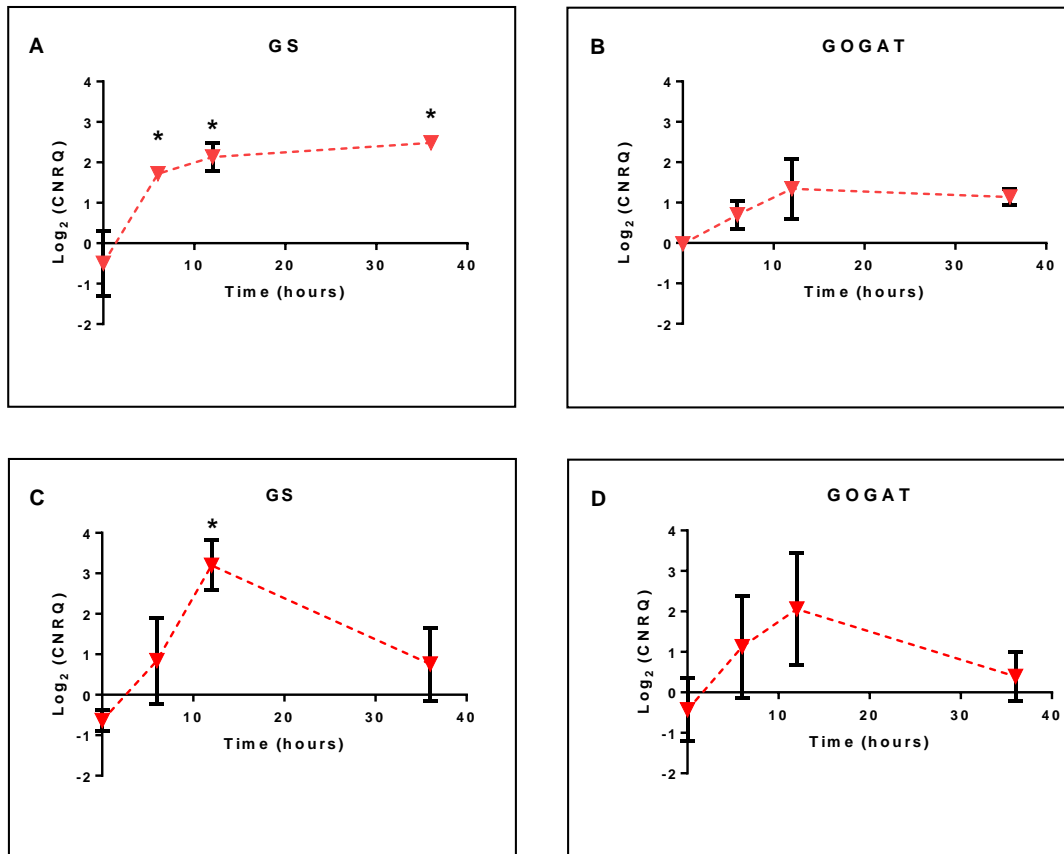
602 Differential gene expression as Log_2 (Calibrated Normalized Relative Quantities) in
603 ammonium-enriched treatment relative to control treatment in above-ground tissue
604 (A: GS; B: GOGAT) and below-ground tissue (C: GS; D: GOGAT) of *Zostera*
605 *muelleri* for the duration of the experiment (36 hours). *Significant difference (pair-
606 wise comparison, $P < 0.05$). Error bars represent \pm SEM (3 plants).

607

608

609

610



611 **REFERENCES**

- 612 ALEXANDRE, A., SILVA, J. & SANTOS, R. 2010. Inorganic nitrogen uptake and
613 related enzymatic activity in the seagrass *Zostera noltii*. *Marine Ecology*, 31,
614 539-545.
- 615 ANDERSON, M. J., GORLEY, R. N. & CLARKE, K. R. 2008. *PERMANOVA+ for*
616 *PRIMER: Guide to Software and Statistical Methods*, Plymouth, UK.
- 617 APOSTOLAKI, E. T., VIZZINI, S. & KARAKASSIS, I. 2012. Leaf vs. epiphyte
618 nitrogen uptake in a nutrient enriched Mediterranean seagrass (*Posidonia*
619 *oceanica*) meadow. *Aquatic Botany*, 96, 58-62.
- 620 BECK, M. W., HECK, K. L., ABLE, K. W., CHILDERS, D. L., EGGLESTON, D.
621 B., GILLANDERS, B. M., HALPERN, B., HAYS, C. G., HOSHINO, K.,
622 MINELLO, T. J., ORTH, R. J., SHERIDAN, P. F. & WEINSTEIN, M. P.
623 2001. The Identification, Conservation, and Management of Estuarine and
624 Marine Nurseries for Fish and Invertebrates: A better understanding of the
625 habitats that serve as nurseries for marine species and the factors that create
626 site-specific variability in nursery quality will improve conservation and
627 management of these areas. *BioScience*, 51, 633-641.
- 628 BERGMANN, N., WINTERS, G., RAUCH, G., EIZAGUIRRE, C., GU, J., NELLE,
629 P., FRICKE, B. & REUSCH, T. B. H. 2010. Population-specificity of heat
630 stress gene induction in northern and southern eelgrass *Zostera marina*
631 populations under simulated global warming. *Molecular Ecology*, 19, 2870-
632 2883.
- 633 BRAKEL, J., WERNER, F. J., TAMS, V., REUSCH, T. B. H. & BOCKELMANN,
634 A.-C. 2014. Current European *Labyrinthula zosterae* Are Not Virulent and
635 Modulate Seagrass (*Zostera marina*) Defense Gene Expression. *PLoS ONE*, 9,
636 e92448.
- 637 BRODERSEN, K. E., LICHTENBERG, M., PAZ, L.-C. & KÜHL, M. 2015.
638 Epiphyte-cover on seagrass (*Zostera marina* L.) leaves impedes plant
639 performance and radial O₂ loss from the below-ground tissue. *Frontiers in*
640 *Marine Science*, 2.
- 641 BURKHOLDER, J. M., GLASGOW, H. B. J. & COOKE, J. E. 1994. Comparative
642 effects of water-column nitrate enrichment on eelgrass *Zostera marina*,
643 shoalgrass *Halodule wrightii*, and widgeongrass *Ruppia maritima*. *Marine*
644 *Ecology Progress Series*, 105, 121-138
- 645 BURKHOLDER, J. M., TOMASKO, D. A. & TOUCHETTE, B. W. 2007.
646 Seagrasses and eutrophication. *Journal of Experimental Marine Biology and*
647 *Ecology*, 350, 46-72.
- 648 BUSTIN, S. A., BENES, V., GARSON, J. A., HELLEMANS, J., HUGGETT, J.,
649 KUBISTA, M., MUELLER, R., NOLAN, T., PFAFFL, M. W., SHIPLEY, G.
650 L., VANDESOMPELE, J. & WITWER, C. T. 2009. The MIQE Guidelines:
651 Minimum Information for Publication of Quantitative Real-Time PCR
652 Experiments. *Clinical Chemistry*, 55, 611-622.
- 653 CALLEJA, M. L., MARBÀ, N. & DUARTE, C. M. 2007. The relationship between
654 seagrass (*Posidonia oceanica*) decline and sulfide porewater concentration in
655 carbonate sediments. *Estuarine, Coastal and Shelf Science*, 73, 583-588.
- 656 CAMBRIDGE, M. L. & MCCOMB, A. J. 1984. The loss of seagrasses in Cockburn
657 Sound, Western Australia. I. The time course and magnitude of seagrass
658 decline in relation to industrial development. *Aquatic Botany*, 20, 229-243.

- 659 DATTOLO, E., RUOCCO, M., BRUNET, C., LORENTI, M., LAURITANO, C.,
660 D'ESPOSITO, D., DE LUCA, P., SANGES, R., MAZZUCA, S. &
661 PROCACCINI, G. 2014. Response of the seagrass *Posidonia oceanica* to
662 different light environments: Insights from a combined molecular and photo-
663 physiological study. *Mar Environ Res*, 101, 225–236.
- 664 DUARTE, C. M. 2002. The future of seagrass meadows. *Environmental*
665 *Conservation*, 29, 192-206.
- 666 DUGDALE, R. C. & WILKERSON, F. P. 1986. The use of ^{15}N to measure nitrogen
667 uptake in eutrophic oceans; experimental considerations. *Limnol Oceanogr*, 6,
668 673-689.
- 669 FRANSSSEN, S. U., GU, J., BERGMANN, N., WINTERS, G., KLOSTERMEIER, U.
670 C., ROSENSTIEL, P., BORNBERG-BAUER, E. & REUSCH, T. B. 2011.
671 Transcriptomic resilience to global warming in the seagrass *Zostera marina*, a
672 marine foundation species. *Proc Natl Acad Sci U S A*, 108, 19276-81.
- 673 FRANSSSEN, S. U., GU, J., WINTERS, G., HUYLMANS, A. K., WIENPAHL, I.,
674 SPARWEL, M., COYER, J. A., OLSEN, J. L., REUSCH, T. B. &
675 BORNBERG-BAUER, E. 2014. Genome-wide transcriptomic responses of
676 the seagrasses *Zostera marina* and *Nanozostera noltii* under a simulated
677 heatwave confirm functional types. *Mar Genomics*, 15, 65-73.
- 678 GOODWIN, T. W. & MERCER, E. I. 1983. *Introduction To Plant Biochemistry*,
679 New York, Pergamon.
- 680 HARRISON, P. J., WATERS, R. E. & TAYLOR, F. J. R. 1980. A broad spectrum
681 artificial seawater medium for coastal and open ocean phytoplankton. *Journal*
682 *of phycology*, 16, 28-35.
- 683 HELLEMANS, J., MORTIER, G., DE PAEPE, A., SPELEMAN, F. &
684 VANDESOMPELE, J. 2007. qBase relative quantification framework and
685 software for management and automated analysis of real-time quantitative
686 PCR data. *Genome Biology*, 8, 1-14.
- 687 KALDY, J., BROWN, C. & ANDERSEN, C. 2013. In situ ^{13}C tracer experiments
688 elucidate carbon translocation rates and allocation patterns in eelgrass *Zostera*
689 *marina*. *Marine Ecology Progress Series*, 487, 27-39.
- 690 KEMP, W. M., TWILLEY, R. R., STEVENSON, J. C., BOYNTON, W. R. &
691 MEANS, J. C. 1983. The Decline of Submerged Vascular Plants in Upper
692 Chesapeake Bay - Summary of Results Concerning Possible Causes. *Marine*
693 *Technology Society Journal*, 17(2), 78-89.
- 694 KOCH, M., BOWES, G., ROSS, C. & ZHANG, X.-H. 2013. Climate change and
695 ocean acidification effects on seagrasses and marine macroalgae. *Global*
696 *Change Biology*, 19, 103-132.
- 697 KRAEMER, G. P., MAZZELLA, L. & ALBERTE, R. S. 1997. Nitrogen Assimilation
698 and Partitioning in the Mediterranean Seagrass *Posidonia oceanica*. *Marine*
699 *Ecology*, 18, 175-188.
- 700 LAM, H. M., COSCHIGANO, K. T., OLIVEIRA, I. C., MELO-OLIVEIRA, R. &
701 CORUZZI, G. M. 1996. THE MOLECULAR-GENETICS OF NITROGEN
702 ASSIMILATION INTO AMINO ACIDS IN HIGHER PLANTS. *Annual*
703 *Review of Plant Physiology & Plant Molecular Biology*, 47, 569.
- 704 LARKUM, A., ORTH, R. & DUARTE, C. 2006. *Seagrasses: Biology Ecology and*
705 *Conservation*. Springer, Netherlands.
- 706 LAURITANO, C., RUOCCO, M., DATTOLO, E., BUIA, M. C., SILVA, J.,
707 SANTOS, R., OLIVÉ, I., COSTA, M. M. & PROCACCINI, G. 2015.

708 Response of key stress-related genes of the seagrass *Posidonia oceanica* in the
709 vicinity of submarine volcanic vents. *Biogeosciences*, 12, 4185-4194.

710 LEWIS, E. 1980. The practical salinity scale 1978 and its antecedents. *IEEE Journal*
711 *of Oceanic Engineering*, 5, 3-8.

712 LONGMORE, A. R., HEGGIE, D. T., FLINT, R., COWDELL, R. & SKYRING, G.
713 W. 2000. Impact of runoff on nutrient patterns in northern Port Phillip Bay,
714 Victoria. *Journal of Australian Geology and Geophysics*, 17, 203-210.

715 MACREADIE, P. I., BAIRD, M. E., TREVATHAN-TACKETT, S. M., LARKUM,
716 A. W. D. & RALPH, P. J. 2014a. Quantifying and modelling the carbon
717 sequestration capacity of seagrass meadows – A critical assessment. *Marine*
718 *Pollution Bulletin*, 83, 430-439.

719 MACREADIE, P. I., SCHLIEP, M., ; , RASHEED, M. A., ; , CHARTRAND, K. M.
720 & RALPH, P. J. 2014b. Molecular indicators of chronic seagrass stress: A new
721 era in the management of seagrass ecosystems? . *Ecological Indicators* 38,
722 279-281

723 MORRIS, L. J. & TOMASKO, D. A. 1993. Proceedings and conclusions of
724 workshops on submerged aquatic vegetation initiative and photosynthetically
725 active radiation. St. Johns River Water Management District, Palatka, FL.

726 MURRAY, L., STURGIS, B. R., BARTLESON, R. D., SEVERN, W. & KEMP, W.
727 M. 1999. Scaling submersed plant community responses to experimental
728 nutrient enrichment. In: BORTONE, S. A. (ed.) *Seagrasses: monitoring,*
729 *ecology, physiology and management*. CRC press.

730 OLSEN, J. L., ROUZE, P., VERHELST, B., LIN, Y. C., BAYER, T., COLLEN, J.,
731 DATTOLO, E., DE PAOLI, E., DITTAMI, S., MAUMUS, F., MICHEL, G.,
732 KERSTING, A., LAURITANO, C., LOHAUS, R., TOPEL, M., TONON, T.,
733 VANNESTE, K., AMIREBRAHIMI, M., BRAKEL, J., BOSTROM, C.,
734 CHOVIATIA, M., GRIMWOOD, J., JENKINS, J. W., JUETERBOCK, A.,
735 MRAZ, A., STAM, W. T., TICE, H., BORNBERG-BAUER, E., GREEN, P.
736 J., PEARSON, G. A., PROCACCINI, G., DUARTE, C. M., SCHMUTZ, J.,
737 REUSCH, T. B. & VAN DE PEER, Y. 2016. The genome of the seagrass
738 *Zostera marina* reveals angiosperm adaptation to the sea. *Nature*, 530, 331-5.

739 ORTH, R. J., CARRUTHERS, T. J. B., DENNISON, W. C., DUARTE, C. M.,
740 FOURQUREAN, J. W., HECK, K. L., HUGHES, A. R., KENDRICK, G. A.,
741 KENWORTHY, W. J., OLYARNIK, S., SHORT, F. T., WAYCOTT, M. &
742 WILLIAMS, S. L. 2006. A Global Crisis for Seagrass Ecosystems.
743 *BioScience*, 56, 987-996.

744 PEDERSEN, M. F. & BORUM, J. 1993. An annual nitrogen budget for a seagrass
745 *Zostera marina* population
746 *MEPS*, 101, 169-177.

747 PFAFFL, M. W. 2001. A new mathematical model for relative quantification in real-
748 time RT-PCR. *Nucleic Acids Research*, 29, 2002-2007.

749 RADONIC, A., THULKE, S., MACKAY, I. M., LANDT, O., SIEGERT, W. &
750 NITSCHKE, A. 2004. Guideline to reference gene selection for quantitative
751 real-time PCR. *Biochemical and Biophysical Research Communications*, 313,
752 856-862.

753 RALPH, P. J., DURAKO, M. J., ENRIQUEZ, S., COLLIER, C. J. & DOBLIN, M. A.
754 2007. Impact of light limitation on seagrasses. *Journal of Experimental*
755 *Marine Biology and Ecology*, 350, 176-193.

- 756 RASMUSSEN, J. R., OLESEN, B. & KRAUSE-JENSEN, D. 2012. Effects of
757 filamentous macroalgae mats on growth and survival of eelgrass, *Zostera*
758 *marina*, seedlings. *Aquatic Botany*, 99, 41-48.
- 759 ROZEN, S. & SKALETSKY, H. 2000. Primer3 on the WWW for general users and
760 for biologist programmers. *Methods Mol Biol*, 132, 365-86.
- 761 SALO, T., REUSCH, T. B. H. & BOSTRÖM, C. 2015. Genotype-specific responses
762 to light stress in eelgrass *Zostera marina*, a marine foundation plant. *Marine*
763 *Ecology Progress Series*, 519, 129-140.
- 764 SCHLIEP, M., PERNICE, M., SINUTOK, S., BRYANT, C. V., YORK, P. H.,
765 RASHEED, M. A. & RALPH, P. J. 2015. Evaluation of Reference Genes for
766 RT-qPCR Studies in the Seagrass *Zostera muelleri* Exposed to Light
767 Limitation. *Scientific Reports*, 5, 17051.
- 768 SERRA, I. A., LAURITANO, C., DATTOLO, E., PUOTI, A., NICASTRO, S.,
769 INNOCENTI, A. & PROCACCINI, G. 2012. Reference genes assessment for
770 the seagrass *Posidonia oceanica* in different salinity, pH and light conditions.
771 *Marine Biology*, 159, 1269-1282.
- 772 SHORT, F. T. & MCROY, C. P. 1984. Nitrogen Uptake by Leaves and Roots of the
773 Seagrass *Zostera marina* L. *Botanica Marina*.
- 774 SHORT, F. T., POLIDORO, B., LIVINGSTONE, S. R., CARPENTER, K. E.,
775 BANDEIRA, S., BUJANG, J. S., CALUMPONG, H. P., CARRUTHERS, T.
776 J. B., COLES, R. G., DENNISON, W. C., ERFTEMEIJER, P. L. A.,
777 FORTES, M. D., FREEMAN, A. S., JAGTAP, T. G., KAMAL, A. H. M.,
778 KENDRICK, G. A., JUDSON KENWORTHY, W., LA NAFIE, Y. A.,
779 NASUTION, I. M., ORTH, R. J., PRATHEP, A., SANCIANGCO, J. C.,
780 TUSSENBROEK, B. V., VERGARA, S. G., WAYCOTT, M. & ZIEMAN, J.
781 C. 2011. Extinction risk assessment of the world's seagrass species. *Biological*
782 *Conservation*, 144, 1961-1971.
- 783 STAPEL, J., AARTS, T., VAN DUYNHOVEN, B., DE GROOT, J., VAN DEN
784 HOOGEN, P. & HEMMINGA, M. 1996. Nutrient uptake by leaves and roots
785 of the seagrass *Thalassia hemprichii* in the Spermonde Archipelago,
786 Indonesia. *Marine Ecology Progress Series*, 134, 195-206.
- 787 TERRADOS, J. & WILLIAMS, S. 1997. Leaf versus root nitrogen uptake by the
788 surfgrass *Phyllospadixtorreyi*. *Marine Ecology Progress Series*, 149, 267-277.
- 789 THURSBY, G. B. & HARLIN, M. M. 1982. Leaf-root interaction in the uptake of
790 ammonia by *Zostera marina*. *Marine Biology*, 72, 109-112.
- 791 TOUCHETTE, B. W. & BURKHOLDER, J. M. 2000. Review of nitrogen and
792 phosphorus metabolism in seagrasses. *Journal of Experimental Marine*
793 *Biology and Ecology*, 250, 133-167.
- 794 TOUCHETTE, B. W. & BURKHOLDER, J. M. 2007. Carbon and nitrogen
795 metabolism in the seagrass, *Zostera marina* L.: Environmental control of
796 enzymes involved in carbon allocation and nitrogen assimilation. *Journal of*
797 *Experimental Marine Biology and Ecology*, 350, 216-233.
- 798 WALKER, N. J. 2002. A technique whose time has come. *Science*, 296, 557-559.
- 799 WINTERS, G., NELLE, P., FRICKE, B., RAUCH, G. & REUSCH, T. B. H. 2011.
800 Effects of a simulated heat wave on photophysiology and gene expression of
801 high- and low-latitude populations of *Zostera marina*. *Marine Ecology*
802 *Progress Series*, 435, 83-95
- 803 YORK, P. H., CARTER, A. B., CHARTRAND, K., SANKEY, T., WELLS, L. &
804 RASHEED, M. A. 2015. Dynamics of a deep-water seagrass population on the

805 Great Barrier Reef: annual occurrence and response to a major dredging
806 program. *Scientific Reports*, 5, 13167.
807 ZIMMERMAN, R. C. & ALBERTE, R. S. 1996. Effect of light/dark transition on
808 carbon translocation in eelgrass *Zostera marina* seedlings. *Marine Ecology*
809 *Progress Series*, 136, 305-309.
810 ZIMMERMAN, R. C., SMITH, R. D. & ALBERTE, R. S. 1987. Is growth of eelgrass
811 nitrogen limited? A numerical simulation of the effects of light and nitrogen
812 on the growth dynamics of *Zostera marina*. *Marine ecology progress series*,
813 41, 167-176.
814 ZMIENKO, A., SAMELAK-CZAJKA, A., GORALSKI, M., SOBIESZCZUK-
815 NOWICKA, E., KOZLOWSKI, P. & FIGLEROWICZ, M. 2015. Selection of
816 Reference Genes for qPCR- and ddPCR-Based Analyses of Gene Expression
817 in Senescing Barley Leaves. *PLoS ONE*, 10, e0118226.
818